

Protein interactions related to biological ice nucleation, allergies, and inflammation

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geb. am 05.03.1990 in Neuwied

Dissertation

zur Erlangung des Grades

'Doctor rerum naturalium (Dr. rer. nat.)'

im Promotionsfach Chemie

am Fachbereich Chemie, Pharmazie und Geowissenschaften der

Johannes Gutenberg-Universität in Mainz

Paul Crutzen Graduate School

Mainz, März 2020

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Mainz, 3 March 2020, Anna Theresa Kunert

Der größte Feind der Qualität ist die Eile.

(Henry Ford)

Abstract

Proteins are involved in many environmental and physiological processes that are relevant for climate and public health, including ice nucleation, allergies, and inflammation. Upon exposure to reactive oxygen and nitrogen species, proteins can undergo chemical modifications influencing their physical, chemical, and biological properties. This dissertation addresses two aspects of current atmospheric and biomedical research related to protein interactions: (i) the ice nucleation activity of proteins and characterization of biological ice nuclei (IN), and (ii) the chemical reactivity and inflammatory potential of proteins acting as allergens or damage-associated molecular patterns (DAMPs).

In the first part of the dissertation, a fully-automated high-throughput Twin-plate Ice Nucleation Assay (TINA) was developed for efficient analysis and characterization of biological IN in laboratory and field samples. The instrument was used to investigate bacterial, fungal, and chemically modified IN as well as air particulate matter. Experiments with fungal ice nuclei from *Fusarium* revealed that the cell-free IN are smaller than 100 kDa and that molecular aggregates can be formed in aqueous solution. The IN activity was not affected by long-term storage, freeze-thaw cycles, and exposure to atmospherically relevant concentrations of ozone and nitrogen dioxide, but it was strongly reduced by heat treatment confirming that the *Fusarium* IN are proteinaceous. Moreover, TINA was used in related further studies investigating the effects of antifreeze proteins, salts, pH changes, and electrostatic interactions on the activity of bacterial IN from *Pseudomonas syringae*. The results provide new insights and help to unravel the molecular mechanisms of biological ice nucleation.

In the second part of the dissertation, the influence of chemical modifications on the allergenic and inflammatory potential of proteins was investigated, building on a review of air pollution effects on allergies. For the grass pollen allergen Phl p 5, the products and kinetics of reactions with O_3/NO_2 and $ONOO^-$ were characterized over a wide range of experimental conditions. The degrees of nitration and oligomerization were higher for $ONOO^-$ than O_3/NO_2 , extending the mechanistic insights gained in related studies with other allergenic proteins and reference substances. Furthermore, chemical modification of the proteinous DAMPs α -Synuclein, heat shock protein 60, and high-mobility-group box 1 protein by $ONOO^-$ was found to enhance innate immune responses mediated by pattern recognition receptors, pro-inflammatory transcription factors, and cytokines (TLR4, NF- κ B, TNF- α , IL-1 β , IL-8). The mechanistic insights gained in these studies contribute to a molecular understanding of adverse health effects caused by oxidative stress and environmental change in the Anthropocene.

Zusammenfassung

Proteine sind in vielen Umwelt- und physiologischen Prozessen involviert einschließlich Eiskeimbildung, Allergien und Entzündungen, die für Klima und öffentliche Gesundheit relevant sind. Durch Wechselwirkungen mit reaktiven Sauerstoff- und Stickstoffspezies können Proteine chemisch modifiziert werden, was ihre physikalischen, chemischen und biologischen Eigenschaften verändert. Diese Dissertation befasst sich mit zwei Aspekten der aktuellen Atmosphären- und biomedizinischen Forschung in Bezug auf Proteinwechselwirkungen: (i) der Eisaktivität von Proteinen und der Charakterisierung von biologischen Eiskeimen, sowie (ii) der chemischen Reaktivität und dem Entzündungspotential von Proteinen, die als Allergene oder Schadens-assoziierte molekulare Muster (*engl.: damage-associated molecular patterns*, DAMPs) wirken.

Im ersten Teil der Dissertation wurde ein vollautomatischer Hochdurchsatz-Gefrierassay (*Twin-plate Ice Nucleation Assay*, TINA) für die effiziente Analyse und Charakterisierung von biologischen Eiskeimen in Labor- und Feldproben entwickelt. Das Instrument wurde für die Untersuchung von bakteriellen, pilzlichen und chemisch modifizierten Eiskeimen sowie von Luftstaub eingesetzt. Experimente mit pilzlichen Eiskeimen von *Fusarium* zeigten, dass die zellfreien Eiskeime kleiner als 100 kDa sind und dass in wässriger Lösung molekulare Aggregate geformt werden können. Die Eisaktivität wurde durch Langzeitlagerung, Gefrier-Tau-Zyklen und atmosphärisch relevanten Konzentrationen von Ozon und Stickstoffdioxid nicht beeinträchtigt, jedoch wurde sie durch Hitzebehandlungen stark reduziert, was bestätigt, dass Eiskeime von *Fusarium* Proteine enthalten. Darüber hinaus wurde TINA in ähnlichen weiteren Studien genutzt um die Effekte von Antieisproteinen, Salzen, pH-Veränderungen und elektrostatischen Wechselwirkungen auf die Aktivität der bakteriellen Eiskeime von *Pseudomonas syringae* zu untersuchen. Die Ergebnisse liefern neue Erkenntnisse und helfen die molekularen Mechanismen der biologischen Eiskeimbildung aufzuklären.

Im zweiten Teil der Dissertation wurde, aufbauend auf einem Review über die Auswirkungen von Luftverschmutzung auf Allergien, der Einfluss chemischer Modifikationen auf das allergene und entzündliche Potenzial von Proteinen untersucht. Für das Graspollenallergen Phl p 5 wurden die Produkte und Kinetiken der Reaktion mit O_3/NO_2 and $ONOO^-$ unter verschiedenen experimentellen Bedingungen charakterisiert. Die Nitrierungs- und Oligomerisierungsgrade waren für $ONOO^-$ höher als für O_3/NO_2 , was die mechanistischen Erkenntnisse erweitert, die in verwandten Studien mit anderen allergenen Proteinen und Referenzsubstanzen gewonnen wurden. Des weiteren wurde herausgefunden, dass die chemische Modifizierung von proteinösen DAMPs (α -Synuklein, Hitzeschockprotein 60, High-Mobility-Group-Protein B1) durch $ONOO^-$ die angeborene Immunantwort verstärkt, die durch Mustererkennungsrezeptoren, entzündungsfördernden Transkriptionsfaktoren und Zytokine (TLR4, NF- κ B, TNF- α , IL-1 β , IL-8) vermittelt wird. Die in diesen Studien gewonnenen mechanistischen Erkenntnisse tragen zu einem besseren molekularen Verständnis der gesundheitsschädlichen Auswirkungen bei, die durch oxidativen Stress und Umweltveränderungen im Anthropozän verursacht werden.

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1 Introduction

1.1 Atmospheric Aerosols and Biological Ice Nucleation

Atmospheric aerosols play important roles in climate, atmospheric chemistry, and public health (Andreae and Rosenfeld, 2008; Pöschl, 2005; Pöschl and Shiraiwa, 2015; Shiraiwa et al., 2017). Primary biological aerosol particles, in short bioaerosols, comprise living or dead organisms, reproductive units, and fragments or excretions from plants and animals (e.g., algae, bacteria, fungal spores, plant pollen, proteins and other biopolymers) (Després et al., 2012; Fröhlich-Nowoisky et al., 2016), spanning a size range from nanometers up to about a tenth of a millimeter with average atmospheric residence times between minutes and weeks (Burrows et al., 2009; Pöschl and Shiraiwa, 2015). They are directly released from the biosphere into the atmosphere, and in the course of atmospheric transport, airborne particles can undergo physical and chemical transformation, stress, and biological aging before they are deposited back to the ground (Fig. 1) (Fröhlich-Nowoisky et al., 2016; Pöschl et al., 2005; Pöschl and Shiraiwa, 2015).

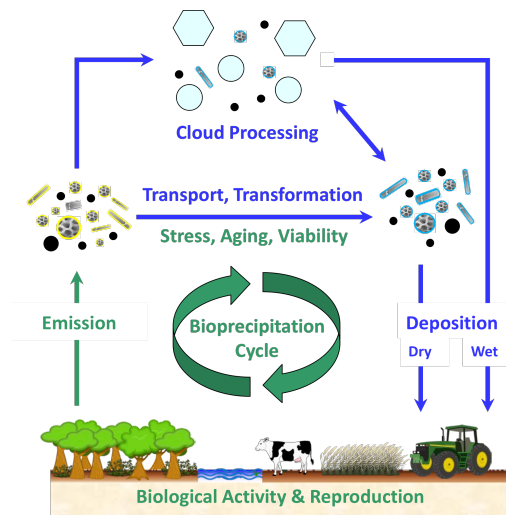


Figure 1: Bioaerosol cycling in the Earth system. Bioaerosols are directly released from the biosphere into the atmosphere, where they can be involved in cloud formation and precipitation and undergo physical and chemical transformation, stress, and biological aging before they are deposited back to the ground. Viable bioparticles can contribute to biological reproduction and further emission, which is particularly efficient when coupled to the water cycle (bioprecipitation). Figure adapted from Fröhlich-Nowoisky et al. (2016).

Some bioaerosols such as certain bacteria or plant pollen can act as ice nuclei (IN), but the impact of biological IN on cloud glaciation and the formation of precipitation is still poorly understood, and aerosol-cloud interactions are among the largest uncertainties in the assessment and understanding of climate change, the Earth system, and the Anthropocene (Boucher et al., 2013; Pöschl and Shiraiwa, 2015). Several studies suggest a triggering effect of biological IN for cloud formation and precipitation (Creamean et al., 2013; DeMott and Prenni, 2010; Failor et al., 2017; Hanlon et al., 2017; Joly et al., 2014; Petters and Wright, 2015; Pratt et al., 2009; Stopelli et al., 2015, 2017), and former studies have shown that biological particles are more efficient than mineral IN (Després et al., 2012; Hill et al., 2014; Hoose and Möhler, 2012; Huffman et al., 2013; Möhler et al., 2007; Pratt et al., 2009). Thus, biological IN are assumed to play an important role for clouds in the temperature range from -15 to 0 °C (DeMott and Prenni, 2010; Morris et al., 2014; Murray et al., 2012).

The best characterized biological IN are common plant-associated bacteria of the genera *Pseudomonas*, *Pantoea*, and *Xanthomonas* (Garnham et al., 2011; Govindarajan and Lindow, 1988; Graether and Jia, 2001; Green and Warren, 1985; Hill et al., 2014; Kim et al., 1987; Ling et al., 2018; Šantl-Temkiv et al., 2015; Schmid et al., 1997; Wolber et al., 1986), and, recently, an ice-nucleation-active (IN-active) *Lysinibacillus* was found (Failor et al., 2017). The first identified IN-active fungi were strains of the genus *Fusarium* (Hasegawa et al., 1994; Pouleur et al., 1992; Richard et al., 1996; Tsumuki et al., 1992). To date, a few more fungal genera with varying initial freezing temperatures such as *Isaria farinosa* (~ -4 °C), *Mortierella alpina* (~ -5 °C), *Puccinia* species (-4 to -8 °C), and *Sarocladium* (formerly named *Acremonium*) *implicatum* (~ -9 °C) have been identified as IN-active (Fröhlich-Nowoisky et al., 2015; Huffman et al., 2013; Morris et al., 2013; Richard et al., 1996). Although IN activity in fungi is known for decades, the IN could not be identified yet. Several studies hypothesize that a proteinaceous compound has to be involved, but the molecular structure remains unknown (Fröhlich-Nowoisky et al., 2015; Hasegawa et al., 1994; Kunert et al., 2019; Pouleur et al., 1992; Pummer et al., 2015; Tsumuki and Konno, 1994).

A wide range of different instruments has been developed and applied for the analysis of IN in immersion freezing experiments, including cloud chambers (e.g., Bundke et al., 2008; Möhler et al., 2003; Stratmann et al., 2004), continuous flow diffusion chambers (e.g., Kanji and Abbatt, 2009; Rogers et al., 2001; Salam et al., 2006), and droplet freezing assays (e.g., Budke and Koop, 2015; Stopelli et al., 2015; Whale et al., 2015). The latter type of instrument is the only one, which is suitable to measure very small IN concentrations in environmental samples that are active at temperatures above -10 °C (Stopelli et al., 2015). The basic operating principle of a droplet freezing assay is the simultaneous cooling of a defined number of aqueous droplets with equal volume in picoliter to milliliter range. Experiments can be performed using a stepped temperature profile or a constant cooling rate, which is similar to those in slowly ascending clouds ($< 1 \text{ m s}^{-1}$) where precipitation is initiated by the formation of ice crystals ($< 1 \text{ °C min}^{-1}$) (Stopelli et al., 2015). The droplet freezing

is detected by either digital cameras based on the reduction of light transmission upon freezing of a liquid sample (e.g., Budke and Koop, 2015; Stopelli et al., 2015; Whale et al., 2015) or infrared cameras based on the latent heat release upon phase change of liquid water to ice (Harrison et al., 2018; Kunert et al., 2018; Zaragotas et al., 2016). The determination of frozen droplets at a given temperature or after a certain time interval enables the quantitative assessment of IN, which was established by Gabor Vali in 1971.

1.2 Protein Modification, Allergies, and Inflammation

Air pollutants can chemically modify proteins either directly in the environment or indirectly in the human body by inducing oxidative stress and inflammation (Reinmuth-Selzle et al., 2017). Responsible for these modifications are reactive oxygen or nitrogen species (ROS/RNS), such as ozone (O_3), hydroxyl radicals, hydrogen peroxide, superoxide (O_2^-), nitrogen oxide (NO), nitrogen dioxide (NO_2) or peroxynitrite ($ONOO^-$), which react with oxidation-sensitive amino acids like cysteine, histidine, methionine, phenylalanine, tryptophan, and tyrosine (Tyr), as well as aliphatic side chains and the peptide backbone (Bachi et al., 2013; Mudd et al., 1969; Sharma and Graham, 2010). Among them, the aromatic amino acid tyrosine is favored in the reaction of proteins with ROS/RNS resulting in the formation of 3-nitrotyrosine (NTyr) and dityrosine cross-links (Kampf et al., 2015; Reinmuth-Selzle et al., 2014; Walcher et al., 2003). Nitrotyrosine can be linked to many severe acute and chronic diseases, such as sepsis, cancer, cardiovascular and neurodegenerative diseases, and it is an important biomarker for inflammation and oxidative stress (Ischiropoulos et al., 1992; Ischiropoulos, 2009; Greenacre and Ischiropoulos, 2001; Hodara et al., 2004; Turko and Murad, 2002). Also protein dimerization and oligomerization are supposed to have a strong influence on the immunogenicity of proteins (Reinmuth-Selzle et al., 2017; Ziegler et al., 2020).

In the atmosphere, allergenic proteins can react with O_3 and NO_2 resulting in oxidized and nitrated proteins as well as protein oligomerization and degradation (Franze et al., 2005; Shiraiwa et al., 2012). The reaction requires a two-step mechanism, in which the protein reacts first with ozone, forming long-lived reactive oxygen intermediates (ROIs), such as tyrosyl radicals (Fig. 2). In a second step, the ROI reacts either with nitrogen dioxide resulting in the formation of 3-nitrotyrosine, combines with another ROI forming dityrosine cross-links, or undergo further oxidation reactions (Kampf et al., 2015; Liu et al., 2017; Shiraiwa et al., 2011).

A similar two-step mechanism is known for protein modification inside the human body. During oxidative stress and inflammation, NO and O_2^- are formed that can react in a diffusion-controlled reaction to the strong oxidizing and nitrating intermediate $ONOO^-$ (Beckman et al., 1990). Under physiological conditions, an equilibrium exists between $ONOO^-$ and its protonated form, peroxy-

trous acid (ONOOH), that depends on the local pH (80 % ONOO⁻ at pH 7.4) (Radi et al., 2001). The anion can react with carbon dioxide (CO₂) forming nitrosoperoxycarbonate (ONOOCO₂⁻), which in turn decomposes to NO₂ and a carbonate radical (CO₃⁻) (Gunaydin and Houk, 2009). These one-electron oxidants can attack the aromatic ring of a tyrosine residue, leading to the formation of a tyrosyl radical (Fig. 2). In a second step, the ROI either combines with NO₂ to yield 3-nitrotyrosine or reacts with another ROI forming dityrosine cross-links, similar to the reaction with O₃ and NO₂ (Gunaydin and Houk, 2009; Pfeiffer et al., 2000). Notably, the reactions involving free radicals are complex, depend on the pH, and yield a mixture of hydroxylated, nitrated, cross-linked, and degraded proteins (Davies and Delsignore, 1987; Gunaydin and Houk, 2009).

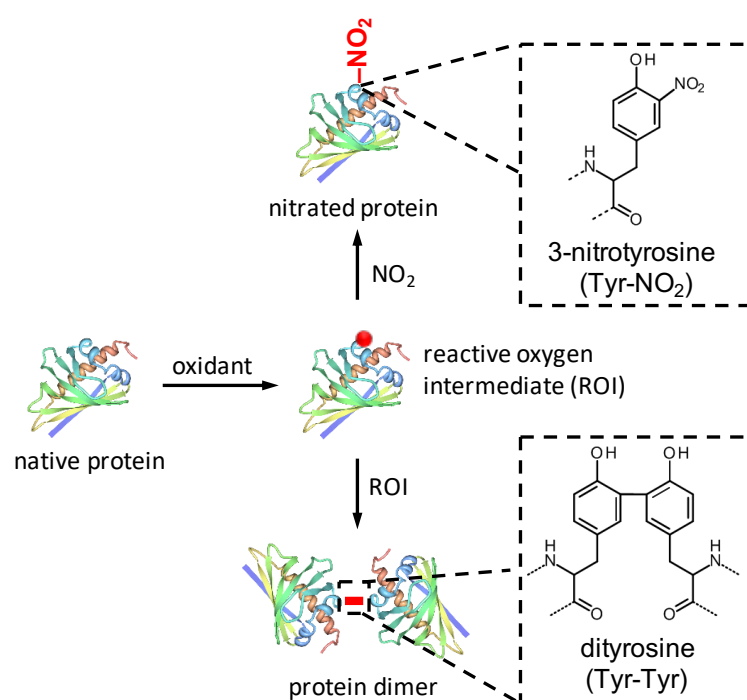


Figure 2: Two-step mechanism for the posttranslational modification of proteins. In the first step, a one-electron oxidant (O₃, NO₂, CO₃⁻) reacts with the aromatic ring of a tyrosine residue forming a reactive oxygen intermediate (ROI), such as a tyrosyl radical (red dot). In a second step, the ROI either combines with NO₂ to yield 3-nitrotyrosine or reacts with another ROI, forming dityrosine cross-links (red bar) and thereby protein dimers and higher oligomers. Figure adapted from Reinmuth-Selzle et al. (2017).

Posttranslational modifications can change the protein structure and stability, affect hydrophobicity and acidity of binding sites, and thereby alter the protein function (Abello et al., 2009; Ackaert et al., 2014; Greenacre and Ischiropoulos, 2001; Ischiropoulos, 2009; Gruijthuijsen et al., 2006; Karle et al., 2012). For example, the addition of a sterically demanding NO₂ group at the ortho position of the aromatic ring causes a significant shift in the pK_a value of the hydroxyl group from ~10 to ~7, thereby affecting the protein conformation (Bachi et al., 2013; Reinmuth-Selzle et al., 2017; Turko

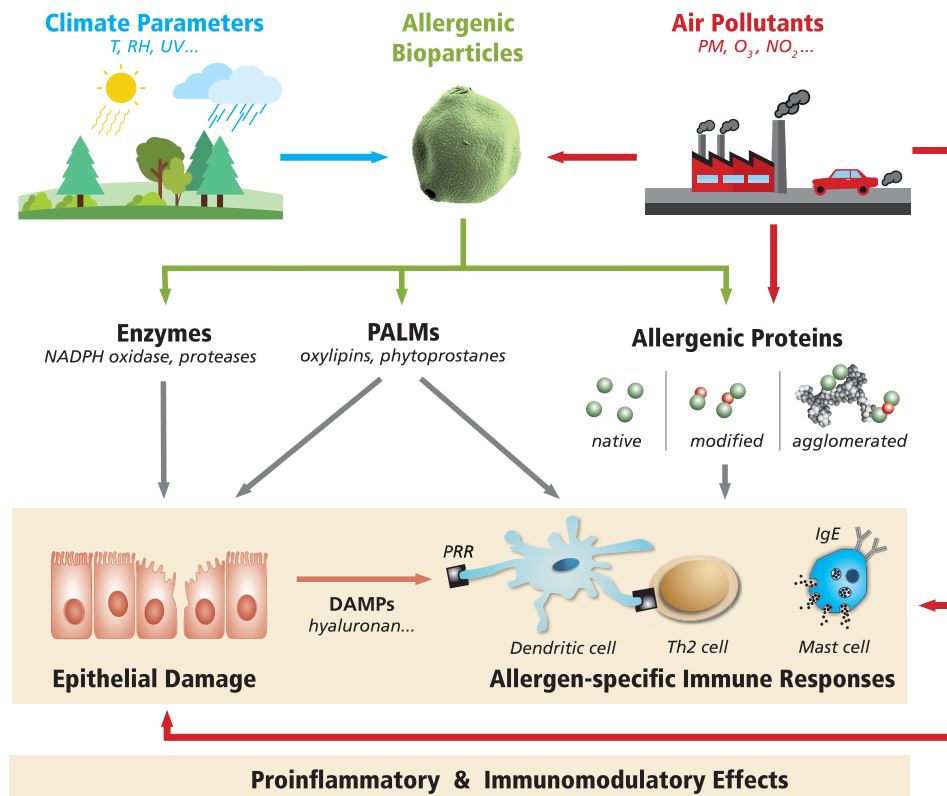


Figure 3: Pathways through which climate parameters and air pollutants can influence the release, potency, and effects of allergens and adjuvants: temperature (T), relative humidity (RH), ultraviolet (UV) radiation, particulate matter (PM), ozone and nitrogen oxides (O₃, NO_x), reduced nicotinamide adenine dinucleotide phosphate (NADPH) oxidase, pollen-associated lipid mediators (PALMs), damage-associated molecular patterns (DAMPs), pattern recognition receptors (PRR), type 2 T helper (Th2) cells, immunoglobulin E (IgE), allergenic proteins (green dots), and chemical modifications (red dots). Figure adapted from (Reinmuth-Selzle et al., 2017).

and Murad, 2002). A chemically modified protein with altered chemical and structural properties can act as adjuvant inducing or potentiating immunogenic reactions and inflammation (Fig. 3). Climate parameters and air pollution can further promote pro-inflammatory and immunomodulatory effects, as both factors can influence the release and thus the environmental abundance of allergenic proteins and biogenic adjuvants (Reinmuth-Selzle et al., 2017). For example, pollen grains can rupture under humid conditions or upon exposure to O₃ or NO₂, leading to subsequent release of cytoplasmic material into the atmosphere (Behrendt and Becker, 2001; Motta et al., 2006; Ouyang et al., 2016; Steiner et al., 2015; Taylor et al., 2004). Besides allergenic proteins, pollen and fungal spores also release other compounds that can act as adjuvants. In particular, the release of nonallergenic, bioactive, pollen-associated lipid mediators (PALMs) with pro-inflammatory and immunomodulatory effects can trigger and enhance allergies (Fig. 3) (Reinmuth-Selzle et al., 2017).

During inflammation, damage-associated molecular patterns (DAMPs) can be actively secreted by epithelial cells, monocytes, macrophages, and other cells, or passively released by damaged and dying cells (Fig. 3). Extracellularly, they can stimulate pattern recognition receptors (PRRs) such as the Toll-like receptor 4 (TLR4). TLR4 signaling leads to the activation of transcription factors like the nuclear factor 'kappa-light-chain-enhancer' of activated B-cells (NF- κ B), which is a key activator of inflammatory cascades (Liu et al., 2014; Takeda and Akira, 2004). NF- κ B induces the expression and modulates the secretion of pro-inflammatory cytokines such as TNF- α , IL-1 β , and IL-8, which can stimulate the corresponding cytokine receptors leading to further activation of NF- κ B and other signaling pathways (Liu et al., 2014, 2016; Varfolomeev and Vucic, 2018; Weber et al., 2010). This positive feedback can amplify and propagate inflammatory processes in autocrine or paracrine fashion (Gan et al., 2014; Lucas and Maes, 2013; Ziegler et al., 2020).

1.3 Research Objectives

This dissertation addresses two aspects of current atmospheric and biomedical research related to proteins: (i) Quantification and characterization of biological ice nuclei in laboratory and field samples, and (ii) studies on the chemical modification of allergenic and pro-inflammatory proteins. The specific objectives can be outlined as follows:

1. Development of a droplet freezing assay for efficient analysis and characterization of biological ice nuclei in laboratory and field samples.
2. Investigation of bacterial and fungal ice nuclei, their stability under atmospherically relevant conditions, and related protein interactions.
3. Elucidation of the reaction kinetics and products of allergenic proteins interacting with ozone, nitrogen dioxide, and peroxyxynitrite.
4. Investigation of the inflammatory potential of peroxyxynitrite-modified proteins acting as damage-associated molecular patterns.

2 Results and Conclusions

2.1 Overview

In the course of my dissertation, I studied protein interactions involved in environmental and physiological processes. The results obtained are described in a total of 16 manuscripts for publication in peer-reviewed scientific journals. The main results and conclusions are summarized below.

A droplet freezing assay was developed, which was applied in several biological IN studies with laboratory and field samples, leading to nine manuscripts including two first-author and three second-author manuscripts. Two of the papers have already been published in and two more are submitted to internationally leading journals of atmospheric science and physical chemistry and are attached in Appendix C.

In the field of allergenic and inflammatory processes, I performed two studies on the chemical modification of different proteins, one as first-author and the other one as co-first-author. I contributed to several other studies related to chemically modified proteins and their effects on human health, leading to five additional manuscripts including one second-author manuscript. Three of the papers have already been published in and one more is submitted to internationally leading journals of environmental science and biology and are attached in Appendix C.

2.2 Biological Ice Nucleation

2.2.1 Twin-plate Ice Nucleation Assay (TINA)

A new high-throughput droplet freezing instrument was developed for efficient analysis and characterization of biological IN in laboratory and field samples under immersion freezing conditions. The Twin-plate Ice Nucleation Assay (TINA) is a fully-automated setup to study ice nucleation and freezing events simultaneously in hundreds of microliter-range droplets using infrared detectors. The instrument was tested and characterized in experiments with bacterial and fungal IN from *Pseudomonas syringae* and *Mortierella alpina*, and it was applied to investigate the influence of chemical processing on the activity of biological IN, in particular the effects of oxidation and nitration reactions. To show

its applicability for field samples, TINA was used to study aqueous extracts of atmospheric aerosols, simultaneously analyzing a multitude of differently pre-treated samples to distinguish several types of IN. The results confirmed that TINA is suitable for high-throughput experiments and efficient analysis of biological IN in laboratory and field samples. For details see Appendix C.1, Kunert *et al.*, Atmos. Meas. Tech., 2018.

2.2.2 Macromolecular Fungal Ice Nuclei in *Fusarium*

Ice nucleation activity in fungi was first discovered in the cosmopolitan genus *Fusarium*, which can be regarded as the best studied IN-active fungus today, but the frequency and distribution of IN activity within *Fusarium* are still insufficiently investigated. More than 100 strains from 65 different *Fusarium* species were tested for IN activity, and in total, ~11 % of all tested species included IN-active strains, and ~16 % of all tested strains showed IN activity above 12 °C. Besides species with known IN activity, *F. armeniacum*, *F. begoniae*, *F. concentricum*, and *F. langsethiae* were newly identified as IN-active. Filtration experiments indicate that cell-free ice-nucleating macromolecules (INMs) from *Fusarium* are smaller than 100 kDa and that molecular aggregates can be formed in solution. Long-term storage and freeze-thaw cycle experiments as well as exposure to atmospherically relevant concentrations of ozone and nitrogen dioxide did not affect the activity of *Fusarium* IN. Heat treatments, however, strongly reduced the IN activity, confirming earlier hypotheses that the INM in *Fusarium* largely consists of a proteinaceous compound. The frequency and the wide distribution of IN activity within the genus *Fusarium*, combined with the stability of the IN under atmospherically relevant conditions, suggest that fungal IN may have a stronger influence on the Earth's water cycle and climate than previously assumed. Additional research is necessary to characterize the INMs in *Fusarium* and processes that can result in their agglomeration to larger protein complexes. To evaluate the implication of these IN on Earth's climate, additional work is required to study the abundance of *Fusarium* IN in environmental samples on a global scale. For details see Appendix C.2, Kunert *et al.*, Biogeosciences, 2019.

2.2.3 Related Studies

TINA was applied to related further IN studies investigating the interactions of bacterial IN from *Pseudomonas syringae* with antifreeze proteins (AFPs), salts, pH changes, and electrostatic interactions.

AFPs have been reported to be able to inhibit the activity of IN, but the generality of this effect is not understood, and for the few known examples of IN inhibition by AFPs, the molecular mechanisms remain unclear. A comprehensive evaluation of the effects of all major classes of AFPs on the activity of bacterial IN was performed. Certain AFPs inhibited the activity of bacterial IN, while other AFPs

showed no effect. Thus, the ability to inhibit the IN activity is not an intrinsic property of AFPs, and the interactions of IN and different AFPs proceed through protein-specific rather than universal molecular mechanisms. For details see Appendix C.3, Schwidetzky *et al.*, submitted to J. Phys. Chem. Lett., 2020a.

The activity of bacterial IN is also reduced by lowering the pH to acidic values, whereas a pH change to alkaline values does not affect the IN activity. The molecular origin for this pH sensitivity is unknown, but a better understanding would provide necessary insights into the driving forces of functional IN aggregation, which is required for the high activity of bacterial IN. Electrostatic interactions are important for the IN activity of *Pseudomonas syringae* as the net charge strongly correlates with the IN activity of the larger IN aggregates. In contrast, the activity of IN monomers is less affected by pH changes. Thus, electrostatic interactions play an essential role in the formation of the highly efficient IN aggregates. For details see Appendix C.4, Lukas *et al.*, submitted to J. Am. Chem. Soc., 2019.

Furthermore, the interactions of different salts with bacterial IN from *Pseudomonas syringae* was investigated. The IN activity was influenced by different salts following the Hofmeister series. Weakly hydrated ions like thiocyanate strongly inhibit bacterial ice nucleation, whereas enhancement was observed for strongly hydrated ions like sulfate. Ongoing experiments with surface-specific sum-frequency generation spectroscopy and molecular dynamic simulations investigate the ability of the different salts to interact with the IN proteins (Schwidetzky *et al.*, in preparation, 2020b).

2.3 Chemical Modification and Inflammatory Potential of Proteins

2.3.1 Timothy Grass Pollen Allergen Phl p 5

The allergenic potential of proteins may be enhanced via posttranslational modifications, but the molecular mechanisms and kinetics of the chemical modifications by O_3/NO_2 and $ONOO^-$ are still not fully understood. Tyrosine nitration and oligomerization of the grass pollen allergen Phl p 5 was investigated upon simultaneous exposure of O_3 and NO_2 and reaction with $ONOO^-$. The nitration degree and the oligomer formation was determined by reversed-phase and size-exclusion liquid chromatography. Both, exposure to O_3/NO_2 and reaction with $ONOO^-$, led to nitration and oligomerization of the allergen. The nitration degree and oligomer formation were dependent on the nitrating agent and the reaction conditions (concentration, exposure time), and they were higher for $ONOO^-$ than for O_3/NO_2 . For the reaction with $ONOO^-$, a maximum nitration degree of 25 % was found, which corresponds to maximal ~ 6 nitrated tyrosine residues per Phl p 5 monomer, whereas exposure to O_3/NO_2 resulted in a maximum nitration degree of 8 % corresponding to ~ 1 nitrotyrosine per Phl p 5 monomer. The obtained results were similar to earlier studies investigating other

allergenic and reference proteins (Bet v 1, BSA). The data will be further interpreted with regard to protein structure, and kinetic modeling will be included in the future. For details see Appendix C.5, Kunert *et al.*, in preparation, 2020.

2.3.2 Pro-inflammatory Proteins α -Syn, HSP60, HMGB1

In the course of inflammatory processes, reactive oxygen and nitrogen species such as peroxynitrite can chemically modify proteins, but the effects of such modifications on human health are not well understood. It was investigated how the innate immune responses via TLR4 and NF- κ B and the pro-inflammatory potential of the proteinous DAMPs α -Syn, HSP60, and HMGB1 changes upon chemical modification with peroxynitrite. For the peroxynitrite-modified proteins, a strongly enhanced activation of TLR4 and NF- κ B was found as well as increased expression and secretion of the pro-inflammatory cytokines TNF- α , IL-1 β , and IL-8. The results show that peroxynitrite-modified DAMPs more potently amplify inflammation via TLR4 activation than the native DAMPs, and provide first evidence that such modifications can directly enhance innate immune responses via a defined receptor. For details see Appendix C.6, Ziegler *et al.*, submitted to Redox Biol., 2020.

2.3.3 Related Studies

Air pollution and climate change are potential drivers for the increasing burden of allergic diseases as the allergenic potential of airborne proteins may be enhanced via posttranslational modification induced by air pollutants like O₃ and NO₂. The molecular mechanisms by which air pollutants and climate parameters may influence allergic diseases, however, are complex and elusive.

The review article provides an overview of physical, chemical, and biological interactions between air pollution, climate change, allergens, adjuvants, and the immune system, addressing how these interactions may promote the development of allergies. Key findings from atmospheric, climate, and biomedical research were reviewed and synthesized. The current state of knowledge, open questions, and future research perspectives are outlined and discussed. For details see Appendix C.7, Reinmuth-Selzle *et al.*, Environ. Sci. Technol., 2017.

In another study, protein tyrosine nitration and oligomerization upon exposure to O₃ and NO₂ were studied, using bovine serum albumin as a model protein. Generally, more tyrosine residues were found to react via the nitration pathway than via the oligomerization pathway, and the experimental results were well reproduced by the kinetic multilayer model of aerosol surface and bulk chemistry. The extent of nitration and oligomerization strongly depends on relative humidity due to moisture-induced phase transition of proteins, highlighting the importance of cloud processing conditions for accelerated protein chemistry. For details see Appendix C.7, Liu *et al.*, Faraday Discuss., 2017.

2.4 Summary and Outlook

In the first part of this dissertation, a high-throughput droplet freezing assay was developed for efficient analysis and characterization of biological ice nuclei in laboratory and field samples. It was used to investigate bacterial, fungal, and chemically modified IN as well as air particulate matter. We found that proteinaceous, nanometer-sized ice nuclei in *Fusarium* are highly stable under atmospherically relevant conditions, suggesting that fungal IN may have a stronger influence on the Earth's water cycle and climate than previously assumed. Further studies revealed how the activity of bacterial IN from *Pseudomonas syringae* is affected by interactions with antifreeze proteins, salts, air pollutants, pH changes, and electric charges.

In the second part, the chemical reactivity and inflammatory potential of proteins acting as allergens or DAMPs was investigated. Chemical modification of the timothy grass pollen allergen Phl p 5 by O_3/NO_2 and $ONOO^-$ led to nitrated and oligomerized products, whereby higher reactivity was observed for the physiological reactant $ONOO^-$ than for the air pollutant mixture of O_3/NO_2 . The results extend the mechanistic insights gained in related studies with other allergenic proteins and reference substances (Bet v 1, BSA). The chemical modification of the proteinous DAMPs α -Syn, HSP60, and HMGB1 enhanced innate immune responses via TLR4 and NF- κ B. The results show that peroxynitrite-modified DAMPs more potently amplify inflammation via TLR4 activation than the native DAMPs, and provide first evidence that such modifications can directly enhance innate immune responses via a defined receptor.

The instrument developed and the measurement results obtained in this dissertation provide new insights and a basis for efficient further investigations of protein interactions relevant for climate and public health. Future challenges in these areas of research comprise the full elucidation of molecular mechanisms of biological ice nucleation as well as identifying and quantifying the immunochemical reactions and underlying mechanisms leading to allergies and inflammation in response to environmental pollution. A comprehensive understanding of the protein interactions and their effects in environmental and physiological processes will help to assess and mitigate the consequences of the steeply increasing and globally pervasive human influence on air quality, climate, and public health in the Anthropocene (Pöschl and Shiraiwa, 2015; Reinmuth-Selzle et al., 2017; Ziegler et al., 2020).

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A List of Abbreviations

α -Syn	α -Synuclein
AFP	Antifreeze protein
BSA	Bovine serum albumin
DAMPs	Damage-associated molecular patterns
HSP60	Heat shock protein 60
HMGB1	High-mobility group box 1 protein
IN	Ice nuclei
INM	Ice-nucleating macromolecule
NF- κ B	nuclear factor 'kappa-light-chain-enhancer' of activated B-cells
PRR	Pattern recognition receptor
RNS	Reactive nitrogen species
ROI	Reactive oxygen intermediate
ROS	Reactive oxygen species
TINA	Twin-plate Ice Nucleation Assay
TLR4	Toll-like receptor 4

B Personal List of Publications

Journal Articles

Published

1. Kunert, A. T., Pöhlker, M. L., Tang, K., Krevert, C. S., Wieder, C., Speth, K. R., Hanson, L. E., Morris, C. E., Schmale III, D. G., Pöschl, U., and Fröhlich-Nowoisky, J.: Macromolecular fungal ice nuclei in *Fusarium*: effects of physical and chemical processing, *Biogeosciences* **2019**, *16*, 4647–4659.
2. Kunert, A. T., Lamneck, M., Helleis, F., Pöschl, U., Pöhlker, M. L., Fröhlich-Nowoisky, J.: Twin-plate Ice Nucleation Assay (TINA) with infrared detection for high-throughput droplet freezing experiments with biological ice nuclei in laboratory and field samples, *Atmos. Meas. Tech.* **2018**, *11*, 6327–6337.
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Submitted

6. Ziegler, K., Kunert, A. T., Reinmuth-Selzle, K., Leifke, A. L., Widera, D., Weller, M. G., Schuppan, D., Fröhlich-Nowoisky, J., Lucas, K., Pöschl, U.: Chemical modification of pro-inflammatory proteins by peroxynitrite increases activation of TLR4 and NF- κ B: Implications for the health effects of air pollution and oxidative stress, *submitted to Redox Biol.* **2020**.
7. Schwidetzky, R., Kunert, A. T., Bonn, M., Pöschl, U., Ramlov, H., DeVries, A. L., Fröhlich-Nowoisky, J., Meister, K.: To Freeze or Not to Freeze: Inhibition of Bacterial Ice Nucleators by Antifreeze Proteins, *submitted to J. Phys. Chem. Lett.* **2020**.
8. Lukas, M., Schwidetzky, R., Kunert, A. T., Pöschl, U., Fröhlich-Nowoisky, J., Bonn, M., and Meister, K.: Electrostatic Interactions Control the Functionality of Bacterial Ice Nucleators, *submitted to J. Am. Chem. Soc.* **2019**.

In Preparation

9. Kunert, A. T., Reinmuth-Selzle, K., Ziegler, K., Krevert, C. S., Wieder, C., Weller, M. G., Lucas, K., Pöschl, U., and Fröhlich-Nowoisky, J.: Reaction kinetics and products of nitration and oligomerization of the major grass pollen allergen Phl p 5 by peroxynitrite and nitrogen dioxide with ozone, *in preparation*.
10. Tang, K., Kunert, A. T., Sánchez-Parra, B., Barbosa, C., Campos, D., Kremper, L., Pickersgill, D. A., Prass, M., Artaxo, P., Pöhlker, C., Pöschl, U., Fröhlich-Nowoisky, J.: Characterization of ice nuclei in ATTO air filter samples, *in preparation*.
11. Lukas, M., Kunert, A. T., Kutus, B., Fröhlich-Nowoisky, J., Abdelmonem, A., Pöschl, U., Bonn, M., Backus, E. H. G.: Ice-nucleating bacteria and fungi: Increased order of water molecules at low temperatures, *in preparation*.
12. Sosso, G. C., Sudera, P., Kunert, A. T., Fröhlich-Nowoisky, J., Pöschl, U., Backus, E. H. G., Bonn, M., and Michaelides, A.: Same Chemistry, Different Structure: Understanding the Key Ingredients of Ice Formation, *in preparation*.
13. Reinmuth-Selzle, K., Tchipilov, T., Kunert, A. T., Ziegler, K., Lucas, K., Pöschl, U., Fröhlich-Nowoisky, J., Weller, M. G.: Quantitative protein analysis of challenging and modified protein samples, *in preparation*.
14. Schwidetzky, R., Yar, A. Y., Lukas, M., Kunert, A. T., Pöschl, U., Koop, T., Bonn, M., Fröhlich-Nowoisky, J., Nagata, Y., Meister, K.: Specific Ion Effects on Bacterial Ice Nucleation, *in preparation*.

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15. Reinmuth-Selzle, K., Bellinghausen, I., Ziegler, K., Kunert, A. T., Leifke, A. L., Lucas, K., Fröhlich-Nowoisky, J., Pöschl, U.: Chemical modification of aeroallergens Bet v 1 and Phl p 5 by air pollutants and peroxyne nitrite affects serum IgE binding and TLR4 signaling, *in preparation*.
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Oral Presentations

1. Kunert, A. T., Ziegler, K., Krevert, C. S., Wieder, C., Reinmuth-Selzle, K., Lucas, K., Pöschl, U., Fröhlich-Nowoisky, J.: Nitration and oligomerization of allergens: efficiency and effects, Lindau Alumni Retreat, Heidelberg, Germany (19th - 22nd September 2019).
2. Kunert, A. T., Lamneck, M., Helleis, F., Pöschl, U., Pöhlker, M. L., Fröhlich-Nowoisky, J.: Twin-plate Ice Nucleation Assay (TINA) with infrared detection for high-throughput droplet freezing experiments with biological ice nuclei in laboratory and field samples, European Aerosol Conference, Goteborg, Sweden (25th - 30th August 2019).
3. Kunert, A. T., Lamneck, M., Helleis, F., Pöschl, U., Pöhlker, M. L., Fröhlich-Nowoisky, J.: Twin-plate Ice Nucleation Assay (TINA) with infrared detection for high-throughput droplet freezing experiments with biological ice nuclei in laboratory and field samples, International Conference on Carbonaceous Particles in the Atmosphere, Vienna, Austria (3rd - 6th April 2019).
4. Kunert, A. T., Ziegler, K., Krevert, C. S., Wieder, C., Reinmuth-Selzle, K., Lucas, K., Pöschl, U., Fröhlich-Nowoisky, J.: Nitration and Oligomerization of Allergens upon Exposure to Ozone and Nitrogen Dioxide as well as Peroxyne nitrite Treatment, 31st Mainzer Allergy Workshop, Mainz, Germany (29th - 30th March 2019).
5. Kunert, A. T., Lamneck, M., Helleis, F., Pöschl, U., Pöhlker, M. L., Fröhlich-Nowoisky, J.: High-Throughput Measurements of Biological Ice Nuclei, Lindau Alumni Retreat, Frankfurt, Germany (17th - 18th November 2018).
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7. Kunert, A. T., Krevert, C. S., Speth, K. R., Scheel, J. F., Kopper, G. M., Kampf, C. J., Liu, F., Hanson, L. E., Franc, G. D., Pöschl, U., Morris, C., Schmale, D. G., Fröhlich-Nowoisky, J.:

Highly active, stable, and cell-free fungal ice nuclei are widespread among *Fusarium* spp, INUIT Final Conference and Second Atmospheric Ice Nucleation Conference, Grasellenbach, Germany (26th February - 1st March 2018).

8. Kunert, A. T., Lamneck, M., Gurk, C., Helleis, F., Klimach, T., Scheel, J., Pöschl, U., Fröhlich-Nowoisky, J.: TINA, a new fully automated high-performance droplet freezing assay coupled to a customized infrared detection system, 5th Workshop - Microphysics of Ice Clouds, Vienna, Austria (22nd - 23rd April 2017).

Poster Presentations

1. Kunert, A. T., Lamneck, M., Helleis, F., Pöschl, U., Pöhlker, M. L., Fröhlich-Nowoisky, J.: Twin-plate Ice Nucleation Assay (TINA) with infrared detection for high-throughput droplet freezing experiments with biological ice nuclei in laboratory and field samples, General Assembly of the European Geoscience Union, Vienna, Austria (7th - 12nd April 2019).
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4. Kunert, A. T., Maurus, I., Liu, F., Reinmuth-Selzle, K., Kampf, C. J., Lang-Yona, N., Ziegler, K., Lucas, K., Fröhlich-Nowoisky, J., Pöschl, U.: Chemical Modification of Proteins by Anthropogenic Air Pollutants, WE-Heraeus-Seminar on Aerosol, Climate and Health, Bad Honnef, Germany (27th - 31st March 2017).
5. Kunert, A. T., Lamneck, M., Gurk, C., Helleis, F., Klimach, T., Scheel, J., Pöschl, U., Fröhlich-Nowoisky, J.: TINA: A New High-Performance Droplet Freezing Assay for the Analysis of Ice Nuclei with Complex Composition, General Assembly of the European Geoscience Union, Vienna, Austria (17th - 22nd April 2016).
6. Kunert, A. T., Lamneck, M., Gurk, C., Helleis, F., Klimach, T., Scheel, J., Pöschl, U., Fröhlich-Nowoisky, J.: TINA: A New High-Performance Droplet Freezing Assay for the Analysis of Ice Nuclei with Complex Composition, Spring Symposium of the Youth Organization of the German Chemical Society, Kiel, Germany (16th - 19th March 2016).

C Selected Publications

1. Kunert, A. T., Lamneck, M., Helleis, F., Pöschl, U., Pöhlker, M. L., Fröhlich-Nowoisky, J.: Twin-plate Ice Nucleation Assay (TINA) with infrared detection for high-throughput droplet freezing experiments with biological ice nuclei in laboratory and field samples, *Atmos. Meas. Tech.* **2018**, *11*, 6327-6337.
2. Kunert, A. T., Pöhlker, M. L., Tang, K., Krevert, C. S., Wieder, C., Speth, K. R., Hanson, L. E., Morris, C. E., Schmale III, D. G., Pöschl, U., and Fröhlich-Nowoisky, J.: Macromolecular fungal ice nuclei in *Fusarium*: effects of physical and chemical processing, *Biogeosciences* **2019**, *16*, 4647–4659.
3. Schwidetzky, R., Kunert, A. T., Bonn, M., Pöschl, U., Ramlov, H., DeVries, A. L., Fröhlich-Nowoisky, J., Meister, K.: To Freeze or Not to Freeze: Inhibition of Bacterial Ice Nucleators by Antifreeze Proteins, *submitted to J. Phys. Chem. Lett.* **2020**.
4. Lukas, M., Schwidetzky, R., Kunert, A. T., Pöschl, U., Fröhlich-Nowoisky, J., Bonn, M., and Meister, K.: Electrostatic Interactions Control the Functionality of Bacterial Ice Nucleators, *submitted to J. Am. Chem. Soc.* **2019**.
5. Kunert, A. T., Reinmuth-Selzle, K., Ziegler, K., Krevert, C. S., Wieder, C., Weller, M. G., Lucas, K., Pöschl, U., and Fröhlich-Nowoisky, J.: Reaction kinetics and products of nitration and oligomerization of the major grass pollen allergen Phl p 5 by peroxyxynitrite and nitrogen dioxide with ozone, *in preparation*.
6. Ziegler, K., Kunert, A. T., Reinmuth-Selzle, K., Leifke, A. L., Widera, D., Weller, M. G., Schuppan, D., Fröhlich-Nowoisky, J., Lucas, K., Pöschl, U.: Chemical modification of pro-inflammatory proteins by peroxyxynitrite increases activation of TLR4 and NF- κ B: Implications for the health effects of air pollution and oxidative stress, *submitted to Redox Biol.* **2020**.
7. Reinmuth-Selzle, K., Kampf, C. J., Lucas, K., Lang-Yona, N., Fröhlich-Nowoisky, J., Shiraiwa, M., Lakey, P. S. J., Lai, S., Liu, F., Kunert, A. T., Ziegler, K., Shen, F., Sgarbanti, R., Weber, B., Bellinghausen, I., Saloga, J., Weller, M. G., Duschl, A., Schuppan, D., Pöschl, U.: Air Pollution and Climate Change Effects on Allergies in the Anthropocene: Abundance, Interaction, and Modification of Allergens and Adjuvants, *Environ. Sci. Technol.* **2017**, *51*, 4119-4141.

8. Liu, F., Lakey, P. S. J., Berkemeier, T., Tong, H., Kunert, A. T., Meusel, H., Cheng, Y., Su, H., Fröhlich-Nowoisky, J., Lai, S., Weller, M. G., Shiraiwa, M., Pöschl, U., Kampf, C. J.: Atmospheric protein chemistry influenced by anthropogenic air pollutants: nitration and oligomerization upon exposure to ozone and nitrogen dioxide, *Faraday Discuss.* **2017**, *200*, 413-427.

Twin-plate Ice Nucleation Assay (TINA) with infrared detection for high-throughput droplet freezing experiments with biological ice nuclei in laboratory and field samples

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Atmospheric Measurement Techniques **11**, 6327-6337, (2018)

Online supplement available at:

<https://doi.org/10.5194/amt-11-6327-2018-supplement>

Author contributions

ATK, ML, and FH developed the instrument. ATK, UP, JFN conceived and designed the experiments.

ATK performed the experiments. MLP wrote the code to process the data and did the error calculation. All authors discussed the data and contributed to the writing of the manuscript.

Atmos. Meas. Tech., 11, 6327–6337, 2018
<https://doi.org/10.5194/amt-11-6327-2018>
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Twin-plate Ice Nucleation Assay (TINA) with infrared detection for high-throughput droplet freezing experiments with biological ice nuclei in laboratory and field samples

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Received: 11 July 2018 – Discussion started: 24 July 2018

Revised: 7 November 2018 – Accepted: 11 November 2018 – Published: 23 November 2018

Abstract. For efficient analysis and characterization of biological ice nuclei under immersion freezing conditions, we developed the Twin-plate Ice Nucleation Assay (TINA) for high-throughput droplet freezing experiments, in which the temperature profile and freezing of each droplet is tracked by an infrared detector. In the fully automated setup, a couple of independently cooled aluminum blocks carrying two 96-well plates and two 384-well plates, respectively, are available to study ice nucleation and freezing events simultaneously in hundreds of microliter-range droplets (0.1–40 μL). A cooling system with two refrigerant circulation loops is used for high-precision temperature control (uncertainty $< 0.2\text{ K}$), enabling measurements over a wide range of temperatures ($\sim 272\text{--}233\text{ K}$) at variable cooling rates (up to 10 K min^{-1}).

The TINA instrument was tested and characterized in experiments with bacterial and fungal ice nuclei (IN) from *Pseudomonas syringae* (Snomax[®]) and *Mortierella alpina*, exhibiting freezing curves in good agreement with literature data. Moreover, TINA was applied to investigate the influence of chemical processing on the activity of biological IN, in particular the effects of oxidation and nitration reactions. Upon exposure of Snomax[®] to O_3 and NO_2 , the cumulative number of IN active at $270\text{--}266\text{ K}$ decreased by more than 1 order of magnitude. Furthermore, TINA was used to study aqueous extracts of atmospheric aerosols, simultaneously investigating a multitude of samples that were pre-treated in different ways to distinguish different kinds of IN. For example, heat treatment and filtration indicated that most biological IN were larger than $5\text{ }\mu\text{m}$. The results confirm that TINA is suitable for high-throughput experiments and effi-

cient analysis of biological IN in laboratory and field samples.

1 Introduction

Clouds and aerosols still contribute the largest uncertainty to the evaluation of the Earth's changing energy budget (Boucher et al., 2013). Thus, the understanding of the contribution of atmospheric aerosols in cloud processes is of fundamental importance. Atmospheric ice nucleation is essential for cloud glaciation and precipitation, thereby influencing the hydrological cycle and climate. Ice can be formed via homogeneous nucleation in liquid water droplets or heterogeneous nucleation triggered by particles serving as atmospheric ice nuclei (IN) (Pruppacher and Klett, 1997).

A wide range of droplet freezing assays and instruments have been developed and applied for the analysis of IN in immersion freezing experiments (e.g., Budke and Koop, 2015; Fröhlich-Nowoisky et al., 2015; Häusler et al., 2018; Murray et al., 2010; O'Sullivan et al., 2014; Stopelli et al., 2014; Tobo, 2016; Vali, 1971b; Whale et al., 2015; Wright and Peters, 2013; Zarogatos et al., 2016). Most of the available assays and instruments, however, are limited to the investigation of small droplet numbers and use optical detection systems in the UV–Vis wavelength range.

Infrared (IR) detectors enable efficient detection of droplet freezing (Harrison et al., 2018; Zarogatos et al., 2016). Upon the phase change of water from liquid to solid, latent heat is released resulting in a sudden temperature change of the

droplet, which can be detected by IR video thermography. In 1995, Ceccardi et al. (1995) used IR video thermography as a new technique to non-destructively study ice formation on plants by visualizing the changes in surface temperature. Wisniewski et al. (1997) evaluated the IR video thermography under controlled conditions and determined it as an excellent method for directly observing ice nucleation and propagation in plants. Since then, IR video thermography has been used in a range of studies investigating freezing in plants (e.g., Ball et al., 2002; Carter et al., 1999; Charrier et al., 2017; Fuller and Wisniewski, 1998; Hacker and Neuner, 2007; Pearce and Fuller, 2001; Sekozawa et al., 2004; Stier et al., 2003; Wisniewski et al., 2008; Workmaster, 1999). Further applications of IR video thermography are investigations of cold thermal stress in insects (Gallego et al., 2016), monitoring of freeze-drying processes (Emteborg et al., 2014), as well as detection of ice in wind turbine blades (Gómez Muñoz et al., 2016) and helicopter rotor blades (Hansman and Dershowitz, 1994). Freezing of single water droplets in an acoustic levitator has also been successfully observed by IR video thermography (Bauerecker et al., 2008).

Here, we introduce the Twin-plate Ice Nucleation Assay (TINA) for high-throughput droplet freezing experiments, in which the temperature profile and freezing of each droplet is tracked by an infrared detector. In the fully automated setup, a couple of independently cooled aluminum blocks are available to study ice nucleation and freezing events in nearly 1000 microliter-range droplets simultaneously. The instrument was developed in the course of the INUIT project over the last three years, in which it has been presented and discussed at several conferences and workshops (Kunert et al., 2016a,b, 2017a,b, 2018). We use the bacterial IN Snomax® and the IN-active fungus *Mortierella alpina* as biological test substances to investigate heterogeneous ice nucleation. Moreover, TINA is applied to investigate the effect of O₃ and NO₂ exposure on the IN activity of Snomax®. Furthermore, aqueous extracts of atmospheric aerosols are treated in different ways and are analyzed for different kinds of IN.

2 Experimental setup

2.1 Technical details

The core of TINA is composed of two independently cooled, customized sample holder aluminum blocks, which have been shaped for multiwell plates with 96 and 384 wells, respectively. In each cooling block, two multiwell plates can be analyzed simultaneously. The maximal droplet volume in the 96-well block is 250 µL, and the minimal droplet volume is 0.1 µL, which is the limit of our liquid handling station (epMotion ep5073, Eppendorf, Hamburg, Germany). For each experiment, new sterile multiwell plates are used (96-well: Axon Labortechnik, Kaiserslautern, Germany, 384-well: Eppendorf, Hamburg, Germany). As shown in Fig. 1, the design

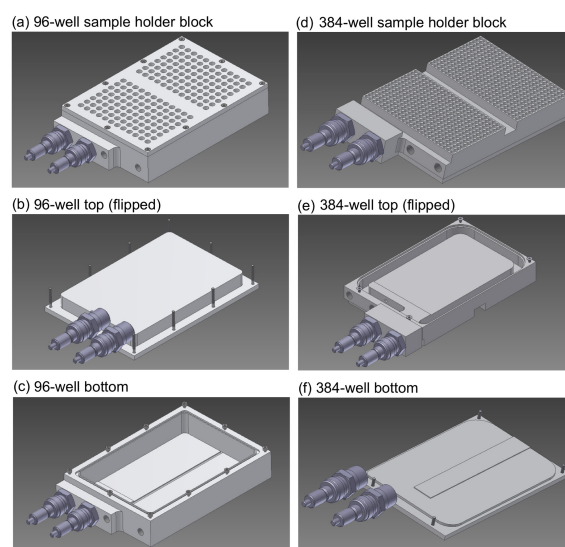


Figure 1. Sample holder and cooling blocks of the Twin-plate Ice Nucleation Assay (TINA) with (a–c) 96-well plates and (d–f) 384-well plates (CAD drawings).

of the two sample holder blocks is basically identical, but the detailed construction varies slightly. Both blocks consist of two parts, a trough and a cap, which are screwed together and sealed with an O-ring. But, for the 96-well block (Fig. 1a), the cap is at the top (Fig. 1b), and the trough is at the bottom (Fig. 1c), whereas, for the 384-well block (Fig. 1d), the trough is at the top (Fig. 1e) and the cap is at the bottom (Fig. 1f). Two openings with Swagelok® adapters for cooling liquid are placed next to each other, and the cooling liquid flows in a small passage around an elevation in the middle of the trough.

The customized sample holder blocks are cooled with a silicon-based cooling liquid (SilOil M80.055.03, Peter Huber Kältemaschinenbau AG, Offenburg, Germany) tempered by an external high-performance refrigeration bath circulator (CC-508 with Pilot ONE, Peter Huber Kältemaschinenbau AG), which can supply temperatures down to 218 K (−55 °C). Both sample holder blocks can be operated in parallel and independently of each other by use of two self-developed mixing valves and cooling loops (Fig. 2). This allows either the cooling of two different droplet freezing assays at the same time or the observation of 960 droplets in one experiment. The mixing of a cold and a warm loop of cooling liquid for each block enables a fast and precise adjustment of the sample holder block temperatures without being dependent on the cooling rate of the refrigeration bath circulator itself. In each experiment, the refrigeration bath circulator is cooled down 5 K below the coldest temperature, which is projected for the experiment, while no mixing of warm and cold cooling liquid occurs. By changing the posi-

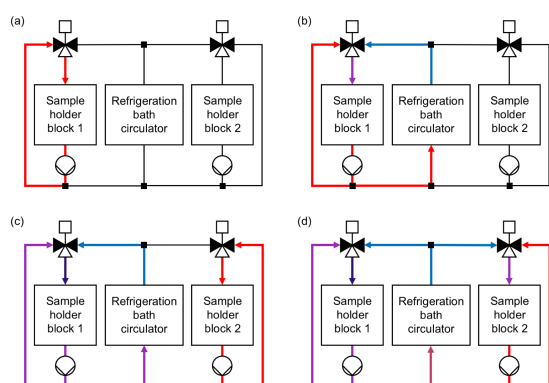


Figure 2. Cooling system layout and operating principle of the Twin-plate Ice Nucleation Assay (TINA). (a) Cooling liquid is pumped in warm cooling loop of sample holder block 1 without connection to colder cooling liquid provided by refrigeration bath circulator. (b) Mixing valve is opened for both warm cooling liquid of warm cooling loop and cold cooling liquid of refrigeration bath circulator. Position of mixing valve defines temperature within sample holder block 1. (c) Sample holder block 1 is cooled further down, while cooling liquid is pumped in warm cooling loop of sample holder block 2. (d) Sample holder block 2 can be run in parallel independently of the temperature in sample holder block 1.

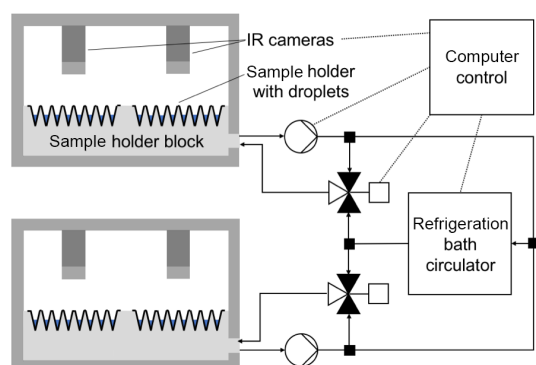


Figure 3. Schematic illustration of the overall setup: sample holder blocks, sample holders with droplets, IR cameras, cooling system with refrigeration bath circulator, pumps and mixing valves, computer control.

tion of the mixing valves for a defined period of time, cold and warm cooling liquids are mixed together so that the desired temperatures within the two blocks are reached. Two pumps (VPP-655 PWM Single Version, Alphacool International GmbH, Braunschweig, Germany) ensure the continuous circulation of cooling liquid through each block independently of the position of the mixing valves. Figure 3 is a schematic illustration of the overall setup of TINA.

2.2 Temperature control and calibration

Within each sample holder block, the temperature is measured with two temperature sensors, an NTC thermistor in the cooling liquid stream (TH-44033, resistance: $2255\Omega/298\text{ K}$, interchangeability: $\pm 0.1\text{ K}$, Omega Engineering GmbH, Deckenpfronn, Germany) and a customized sensor with an NTC thermistor (10K3MRBD1, resistance: $10000\Omega/298\text{ K}$, interchangeability: $\pm 0.2\text{ K}$, TE Connectivity Company, Galway, Ireland) and a thermocouple (K type, 0.08 mm diameter, Omega), which were glued together in a $1/8\text{ in.}$ Swagelok® pipe, placed inside the elevated central part of the block. With further thermocouples connected to this reference, this offers the possibility of measuring temperature differences between the NTC thermistor and arbitrary points simultaneously. Another NTC thermistor (10K3MRBD1, resistance: $10000\Omega/298\text{ K}$, interchangeability: $\pm 0.2\text{ K}$, TE Connectivity Company) monitors the temperature behind each mixing valve. Temperature control within the entire system is achieved by a self-developed microcontroller-based electronic system. The analog input unit is equipped with a low-noise, 24 bit ADC (ADS1256, Texas Instruments Incorporated, Dallas, TX, USA), which assures the required accuracy to process the resolution of the used thermistors. All thermistors had been calibrated with a reference thermometer (2180A, Fluke Deutschland GmbH, Glottertal, Germany; 0.01 K resolution, system uncertainty $\delta_{\text{Fluke}}: \pm 0.08\text{ K}$ at 223 K and $\pm 0.07\text{ K}$ at 273 K). Therefore, all thermistors were bound together with a PT100 sensor of the reference thermometer, and the bundle was placed inside a brass cylinder filled with cooling liquid. The cylinder was placed inside the cooling bath of the refrigeration bath circulator. The temperature within the bath was cooled down from 303.2 to 218.2 K (30.0 to -55.0°C) in 5 K steps, warmed to 220.7 K (-52.5°C), and raised again from 220.7 to 300.7 K (-52.5 to 27.5°C) in 5 K steps. Each step was kept for 30 min to equilibrate the temperature, while the resistance of all thermistors and the temperature measured by the reference thermometer were monitored. For the conversion of the measured resistance of the thermistors into temperature, cubic spline interpolation was used ($\delta_{\text{Ipol}} < 0.01\text{ K}$). We obtained the thermistor calibration uncertainty $\delta_{\text{Thermistor}} < 0.09\text{ K}$ ($\delta_{\text{Thermistor}} = \delta_{\text{Fluke}} + \delta_{\text{Ipol}}$).

To determine a potential temperature gradient of the sample holder blocks, two thermocouples (K type, 0.08 mm diameter, Omega) were positioned in various wells of multi-well plates (Fig. S1a and b in the Supplement), each filled with $30\text{ }\mu\text{L}$ pure water (see Sect. 3.1). These thermocouples were connected to the thermocouple in the elevation of each sample holder block, and the temperature offset between sample holder block and wells was measured for a continuous cooling rate of 1 K min^{-1} (Fig. S1c). Below -2°C , the temperature offset between sample holder block and wells is nearly constant, in this example ~ 0.16 and $\sim 0.19\text{ K}$. The measurement was performed in duplicate for all observed

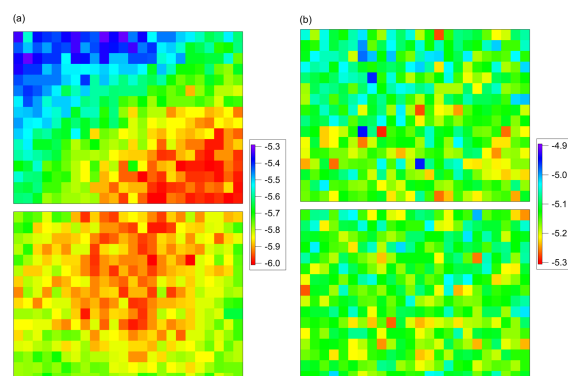


Figure 4. Measurement of temperature gradient of 384-well sample holder block using *Mortierella alpina* 13A as calibration substance. A correction matrix was calculated to compensate for temperature gradient and offset. (a) Data before correction. (b) Data after correction.

wells. Figure S2 shows the temperature gradient exemplarily for the 384-well sample holder block in a 2-D interpolation based on all measurements.

To characterize the uncertainty of this measurement, the two thermocouples were placed in an ice water bath, and the sample holder block was cooled down to 2, 1, 0, -1 , and -2 °C (T_{block}), while the difference between the ice water and the block temperature was monitored by the thermocouples (T_{diffTC}) (Fig. S3). From these experiments, we obtained thermocouple uncertainties $\delta_{\text{TC}} < 0.05$ K ($\delta_{\text{TC}} = T_{\text{block}} + T_{\text{diffTC}}$).

Additionally, we used undiluted IN filtrate of *Mortierella alpina* 13A (see Sect. 3.2) as calibration substance, and a freezing experiment was performed as described for the biological reference materials (see Sect. 3.2). These results were used to compensate for the temperature gradient, and the thermocouple measurements were used to correct the temperature offset between gradient-corrected wells and thermistors. A correction matrix was calculated, and this matrix was used to correct subsequent freezing experiments. Figure 4 shows the results of the fungal IN filtrate measurement (a) before and (b) after correction. After correction, all fungal IN filtrate measurements showed a standard deviation of < 0.06 K (δ_{Morti}). From the calibration measurements, we obtained a total uncertainty estimate of $\delta_{\text{total}} < 0.2$ K ($\delta_{\text{total}} = \delta_{\text{Thermistor}} + \delta_{\text{TC}} + \delta_{\text{Morti}}$).

2.3 Infrared video thermography

Droplet freezing is determined by a distinct detection system, where the temperature profile of each single droplet is tracked by infrared cameras (Seek Thermal Compact XR, Seek Thermal Inc., Santa Barbara, CA, USA) coupled to a self-written software. The camera has a resolution of $206 \times$

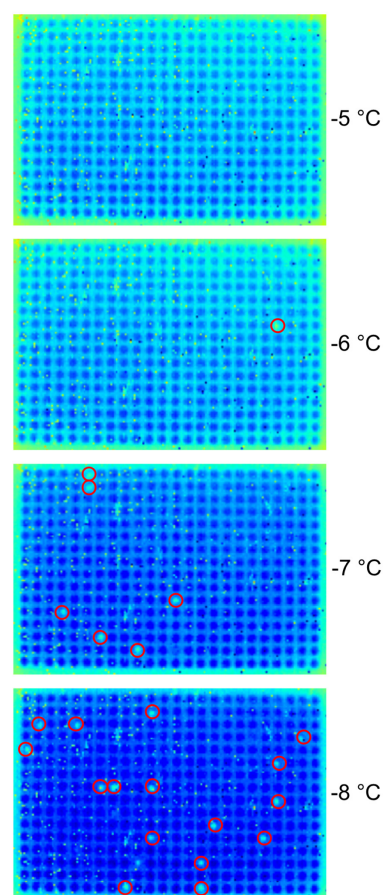


Figure 5. Sequence of infrared camera images showing 384 droplets during cooling. Red circles indicate freezing droplets.

156 pixels, and it takes 10 pictures per second. These pictures are averaged to one picture per second. The concept enables a doubtless determination of freezing events because freezing of supercooled liquid releases energy, which leads to an abrupt rise in the detected temperature of the observed droplet, as discussed earlier (Sect. 1). This detection system uses the IR video thermography only to determine freezing events, while the proper temperature is monitored by thermistors. Figure 5 is a sequence of infrared camera images showing 384 droplets during cooling and freezing (red circles). Software analysis uses a grid of 96 and 384 points, respectively, where the grid point is set to the center of each well enabling one to fit the dimensions of each plate under different perspective angles. The temperature is tracked for each well during the experiment. A self-written algorithm detects a local maximum shortly followed by a local minimum in the derivative of the temperature profile, which is caused

by the release of latent heat during freezing. The software exports the data for each droplet in CSV format.

2.4 Data analysis

Assuming ice nucleation as a time-independent (singular) process, the number concentration of IN ($\frac{\Delta N_m}{\Delta T}$) active at a certain temperature (T) per unit mass of material is given by Eq. (1) (Vali, 1971a).

$$\frac{\Delta N_m}{\Delta T}(T) = -\ln\left(1 - \frac{s}{a - \sum_{i=0}^j s}\right) \cdot \frac{c}{\Delta T}; 0 \leq j \leq a \quad (1)$$

$$\text{with } c = \frac{V_{\text{wash}}}{V_{\text{drop}}} \cdot \frac{d}{m}, \quad (2)$$

where s is the number of freezing events in 0.1 K bins (ΔT), a is the number of all droplets, m is the mass of the particles in the initial suspension, V_{wash} is the volume of the initial suspension, V_{drop} is the droplet volume, and d is the dilution factor of the droplets relative to m . The measurement uncertainty ($\delta \frac{\Delta N_m}{\Delta T}(T)$) was calculated using the counting error of s plus one digit and the Gaussian error propagation (Eq. 3).

$$\delta \frac{\Delta N_m}{\Delta T}(T) = \sqrt{\left(\frac{1}{1 - \frac{s}{a - \sum_{i=0}^j s}} \cdot \frac{c}{\Delta T} \cdot \frac{\sqrt{s+1}}{a - \sum_{i=0}^j s}\right)^2 + \left(\frac{1}{1 - \frac{s}{a - \sum_{i=0}^j s}} \cdot \frac{c}{\Delta T} \cdot \frac{s \cdot \sqrt{\sum_{i=0}^j s + 1}}{(a - \sum_{i=0}^j s)^2}\right)^2} \quad (3)$$

The cumulative IN number concentration ($N_m(T)$) is given by Eq. (4).

$$N_m(T) = -\ln\left(1 - \frac{\sum_{i=0}^j s}{a}\right) \cdot c; 0 \leq j \leq a \quad (4)$$

The error of the cumulative IN number concentration ($\delta N_m(T)$) was calculated using Eq. (5).

$$\delta N_m(T) = \sqrt{\left(\frac{c}{1 - \frac{\sum_{i=0}^j s}{a}} \cdot \frac{\sqrt{\sum_{i=0}^j s + 1}}{a}\right)^2} \quad (5)$$

According to the above equations, the uncertainty is proportional to the number of frozen droplets per temperature bin. In the freezing experiments described below, the lowest number of freezing events and largest uncertainties were obtained at the lower and higher end of each dilution series (Poisson distribution). Data points with uncertainties $\geq 100\%$ were excluded (overall less than 6 % of the measurement data).

3 Freezing experiments

The fully automated TINA setup was tested and characterized for immersion freezing experiments with pure-water

droplets, as well as Snomax[®] and IN filtrate of the fungus *Mortierella alpina* as biological reference substances. Moreover, TINA was used to study the effect of O₃ and NO₂ exposure on the IN activity of Snomax[®]. Furthermore, TINA was applied to atmospheric aerosol samples.

3.1 Pure water

Pure water was obtained from a Barnstead[™] GenPure[™] xCAD Plus water purification system (Thermo Scientific, Braunschweig, Germany). The water was autoclaved at 394 K (121 °C) for 20 min, filtered three times through a sterile 0.1 µm pore diameter sterile polyethersulfone (PES) vacuum filter unit (VWR International, Radnor, PA, USA), and autoclaved again.

For background measurements, 3 µL aliquots of autoclaved and filtered pure water were pipetted into new sterile multiwell plates by a liquid handling station. Therefore, four (96-well plate) and eight (384-well plate) different water samples were pipetted column-wise distributed into the plates. In total, six columns per sample were apportioned over the two twin plates, i.e., 48 droplets per sample in 96-well plates, and 96 droplets per sample in 384-well plates. The plates were placed in the sample holder blocks and were cooled down quickly to 273 K (0 °C) and, as soon as the temperature was stable for 1 min, in a continuous cooling rate of 1 K min^{−1} further down to 238 K (−35 °C).

As the phase transition from liquid water to ice is kinetically hindered, supercooled water can stay liquid at temperatures down to 235 K (−38 °C), where homogeneous ice nucleation takes place. This is only true for nanometer-sized droplets because the freezing temperature is dependent on droplet volume and cooling rate, and the classical nucleation theory predicts a homogeneous freezing temperature of about 240 K (−33 °C) for microliter-volume droplets using a cooling rate of 1 K min^{−1} (Fornea et al., 2009; Murray et al., 2010; Pruppacher and Klett, 1997; Tobo, 2016). However, several studies reported average freezing temperatures for microliter-volume droplets of pure water at significantly higher temperatures because of possible artifacts (e.g., Conen et al., 2011; Fröhlich-Nowoisky et al., 2015; Hill et al., 2016; Whale et al., 2015). To our knowledge, only two studies reported an average homogeneous freezing temperature of 240 K (−33 °C) for microliter-volume droplets and a cooling rate of 1 K min^{−1}, using hydrophobic surfaces as a contact area for the droplets (Fornea et al., 2009; Tobo, 2016). Providing microliter droplets free of suspended IN and surfaces free of contaminants is difficult, so the temperature limit below which freezing cannot be traced back to heterogeneous IN needs to be determined individually for each setup.

Our results showed that most pure-water droplets froze around 248 K (−25 °C) in 96-well plates (Fig. 6a) and around 245 K (−28 °C) in 384-well plates (Fig. 6b). The 96-well plates were obtained from a different manufacturer than the 384-well plates. All in all, these freezing temperatures are

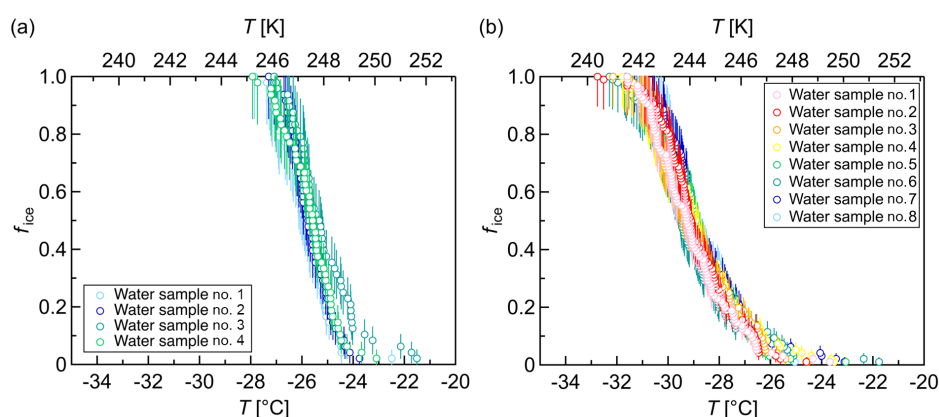


Figure 6. Freezing experiments with pure-water droplets. Fraction of frozen droplets (f_{ice}) vs. temperature (T) obtained with a continuous cooling rate of 1 K min^{-1} and a droplet volume of $3 \mu\text{L}$. (a) Four different samples with 48 droplets each apportioned over two 96-well plates. (b) Eight different samples with 96 droplets each apportioned over two 384-well plates. The error bars were calculated using the counting error and the Gaussian error propagation. The temperature error is 0.2 K .

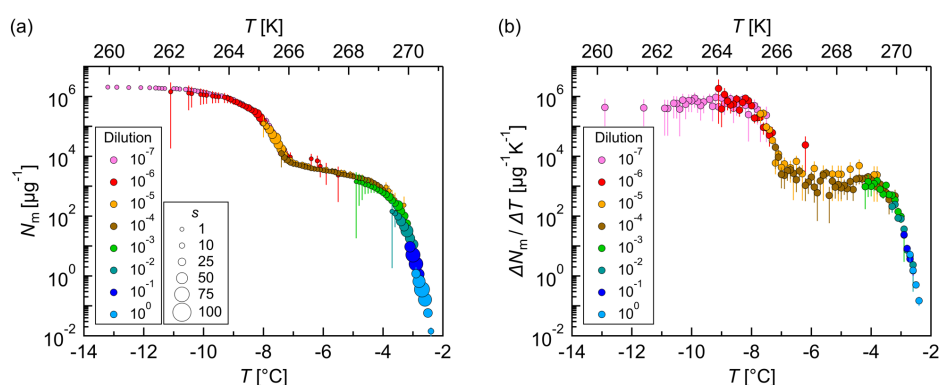


Figure 7. Measurements of dilution series of bacterial IN (Snomax®). (a) Cumulative number of IN (N_m) and (b) differential number of IN ($\Delta N_m / \Delta T$) per unit mass of Snomax® vs. temperature (T). Droplets of the same dilution of three independent measurements were added to a total droplet number of 288 (3×96 droplets). Symbol colors indicate different dilutions; symbol size indicates the number of frozen droplets per 0.1 K bin (s). The error bars were calculated using the counting error and the Gaussian error propagation. The temperature error is 0.2 K .

substantially above the expected temperatures for homogeneous nucleation of microliter droplets, but they are in accord with the results of Whale et al. (2015).

3.2 Biological reference materials

The performance of TINA was further assessed using Snomax® as a bacterial IN-active reference substance (e.g., Budke and Koop, 2015; Hartmann et al., 2013; Möhler et al., 2008; Turner et al., 1990; Ward and DeMott, 1989) and IN filtrate of the well-studied IN fungus *Mortierella alpina* (Fröhlich-Nowoisky et al., 2015; Pummer et al., 2015).

Snomax® was obtained from SMI Snow Makers AG (Thun, Switzerland), and a stock solution was prepared in pure water with an initial mass concentration of 1 mg mL^{-1} . This suspension was then serially diluted 10-fold with pure water by the liquid handling station. The resulting Snomax® concentrations varied between 1 mg mL^{-1} and 0.1 ng mL^{-1} , equivalent to a total mass of Snomax® between $3 \mu\text{g}$ and 0.3 pg , respectively, per $3 \mu\text{L}$ droplet.

Each dilution was pipetted column-wise distributed over the twin plates as described before in 96 droplets into 384-well plates by the liquid handling station. Two plates at a time were placed inside the 384-well sample holder block, and the plates were cooled down quickly to 273 K (0°C) and, as

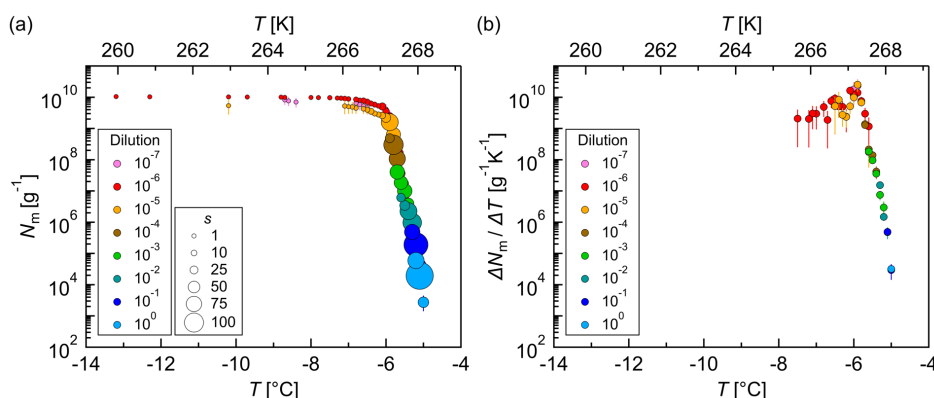


Figure 8. Measurements of dilution series of fungal IN (*Mortierella alpina* 13A). **(a)** Cumulative number of IN (N_m) and **(b)** differential number of IN ($\Delta N_m / \Delta T$) per unit mass of mycelium vs. temperature (T). Droplets of the same dilution of three independent measurements were added to a total droplet number of 288 (3×96 droplets). Symbol colors indicate different dilutions; symbol size indicates the number of frozen droplets per 0.1 K bin (s). The error bars were calculated using the counting error and the Gaussian error propagation. The temperature error is 0.2 K.

soon as the temperature was stable for 1 min, in a continuous cooling rate of 1 K min^{-1} further down to 253 K (-20°C).

Three independent experiments with Snomax[®] showed reproducible results (Fig. S4), and, therefore, droplets of the same dilution were added to a total droplet number of 288. The obtained results were plotted in a cumulative and a differential IN spectrum (Fig. 7). The cumulative IN number concentration represents the total number of IN active above a certain temperature. The cumulative IN spectrum showed two strong increases: around 270 K (-3°C) and around 265 K (-8°C). These findings are in good agreement with the results of Budke and Koop (2015). The differential IN number concentration was calculated according to Vali (1971a), and it represents the number of IN active in a particular temperature interval. The differential IN spectrum showed a similar shape as the cumulative IN spectrum with a distinct plateau between 268 and 266 K (-5 and -7°C) and two slight maxima, around 269 K (-4°C) and around 264 K (-9°C). This indicates the presence of highly efficient IN, active at a temperature of approximately 269 K (-4°C), and less-efficient IN, active around 264 K (-9°C). The fact that the less-efficient IN appeared in higher dilutions implies that they occur in higher concentrations than the highly efficient IN. The presence of further IN with lower freezing temperatures and low concentrations cannot be excluded.

The analysis of different IN active within a wide temperature range was only possible with the measurement of a dilution series. TINA enables the simultaneous measurement of such a dilution series with high statistics in a short period of time.

Mortierella alpina 13A was grown on full-strength PDA (VWR International GmbH, Darmstadt, Germany) at 277 K (4°C) for 7 months. Fungal IN filtrate was prepared as de-

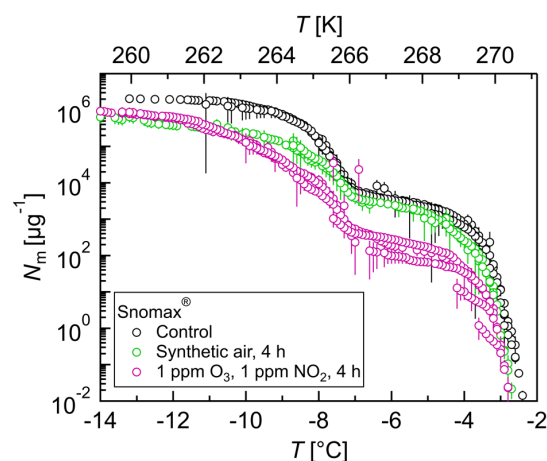


Figure 9. Freezing experiments with ozonized and nitrated bacterial IN. Cumulative number of IN (N_m) per unit mass of Snomax[®] vs. temperature (T). Droplets of the same dilution of three independent measurements were added to a total droplet number of 288 (3×96 droplets). Symbol colors indicate different exposure conditions. The error bars were calculated using the counting error and the Gaussian error propagation. The temperature error is 0.2 K.

scribed previously (Fröhlich-Nowoisky et al., 2015; Pummer et al., 2015) and contained IN from spores and mycelial surfaces. It was serially diluted 10-fold with pure water by the liquid handling station. The experiment was performed as described above.

For test measurements with fungal IN, IN filtrate of three different culture plates from *Mortierella alpina* 13A was

measured, and the results were reproducible (Fig. S5). The cumulative number of IN per gram mycelium only varied by 1 order of magnitude, which is a good achievement for a biological sample, and droplets of the same dilutions were added to a total droplet number of 288. A cumulative IN spectrum (Fig. 8a) and a differential IN spectrum (Fig. 8b) were plotted. The cumulative number of IN and the initial freezing temperature of 268 K (-5°C) are in good agreement with the literature (Fröhlich-Nowoisky et al., 2015; Pummer et al., 2015). The cumulative and the differential IN spectra showed similar shapes with one maximum around 267 K (-6°C), indicating the presence of one type of IN, which is highly efficient.

3.3 Ozonized and nitrated samples

To study the effect of O_3 and NO_2 exposure on the IN activity of Snomax[®], an aliquot of 1 mL of a 1 mg mL^{-1} suspension of Snomax[®] in pure water was exposed in liquid phase to gases with or without O_3 and NO_2 as described in Liu et al. (2017).

Briefly, O_3 was produced by exposing synthetic air to UV light (L.O.T.-Oriel GmbH & Co. KG, Darmstadt, Germany), and the O_3 concentration was adjusted by tuning the amount of UV light. The gas flow was $\sim 1.9\text{ L min}^{-1}$, and it was mixed with N_2 containing $\sim 5\text{ ppmV NO}_2$ (AIR LIQUIDE Deutschland GmbH, Düsseldorf, Germany). The NO_2 concentration was regulated by the addition of the amount of the $\sim 5\text{ ppmV NO}_2$ gas. The O_3 and NO_2 concentrations were monitored with commercial monitoring instruments (ozone analyzer: 49i, Thermo Scientific, Braunschweig, Germany; NO_x analyzer: 42i-TL, Thermo Scientific). The gas mixture was directly bubbled through 1 mL of the Snomax[®] solution at a flow rate of 60 mL min^{-1} using a Teflon tube (ID: 1.59 mm). The Snomax[®] solution was exposed to a mixture of 1 ppm O_3 and 1 ppm NO_2 for 4 h, representing the exposure to an atmospherically relevant amount of about 200 ppb each for about 20 h. The exposure experiments were performed in triplicate. After exposure, the treated samples were serially diluted and the IN activity was measured as described for the Snomax[®] reference measurements.

The results showed that gas exposure affected the IN activity of Snomax[®] (Fig. 9). High concentrations of O_3 and NO_2 reduced the cumulative number of IN from Snomax[®] between 1 and 2 orders of magnitude, while exposure to synthetic air showed smaller effects.

Snomax[®] contains IN proteins of the bacterium *Pseudomonas syringae*. Attard et al. (2012) found no significant or only weak effects of exposure to $\sim 100\text{ ppb O}_3$ and $\sim 100\text{ ppb NO}_2$ on the IN activity of two strains of *P. syringae*, and a variable response of a third strain, suggesting a strain-specific response.

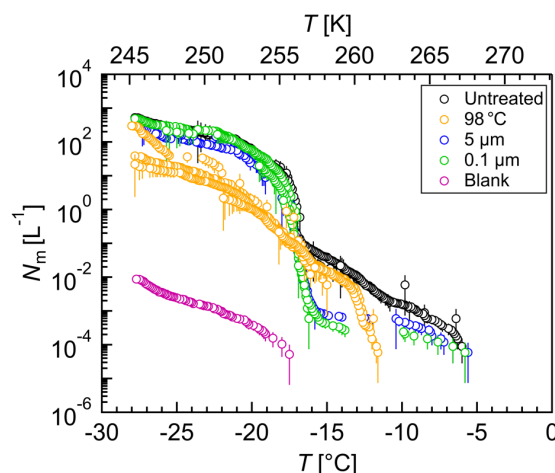


Figure 10. Freezing experiments with aqueous extracts of atmospheric aerosols. Cumulative number of IN per liter air (N_m) vs. temperature (T) for untreated (black), heated (yellow), $5\text{ }\mu\text{m}$ filtered (blue), $0.1\text{ }\mu\text{m}$ filtered (green), and blank (magenta) filter extracts. Droplets of the same dilution of two aliquots were added to a total droplet number of 192 (2×96 droplets). The error bars were calculated using the counting error and the Gaussian error propagation. The temperature error is 0.2 K .

3.4 Air filter samples

Total suspended particle samples were collected onto 150 mm glass fiber filters (Type MN 85/90, Macherey-Nagel GmbH, Düren, Germany) using a high-volume sampler (DHA-80, Digitel Elektronik AG, Hegnau, Switzerland) operated at 1000 L min^{-1} , which was placed on the roof of the Max Planck Institute for Chemistry (Mainz, Germany). There, a mix of urban and rural continental boundary layer air can be sampled in central Europe. The filter was taken in April 2018, and the sampling period was seven days, corresponding to a total air volume of approximately $10\,000\text{ m}^3$. Filters were pre-baked at 603 K (330°C) for 10 h to remove any biological material, and blank samples were taken to detect possible contaminations. All filters were packed in pre-baked aluminum bags, and loaded filters were stored at 193 K (-80°C) until analysis.

An aerosol and a blank filter were cut with a sterilized scissor into aliquots ($\sim 1/16$), and the exact percentage was determined gravimetrically. For reproducibility, two filter sample aliquots of each filter were extracted. Each filter sample aliquot was transferred into a sterile 50 mL tube (Greiner Bio-One, Kremsmünster, Austria), and 10 mL of pure water was added. The tubes were shaken horizontally at 200 rpm for 15 min. Afterwards, the filter was removed, and the aqueous extract was tested for IN activity. To further characterize the IN, the effects of filtration and heat treatment were investigated. Therefore, aliquots of the extract were treated as

follows: (i) 1 h at 371 K (98 °C), (ii) filtration through a 5 µm pore diameter filter (Acrodisc®, PES, Pall GmbH, Dreieich, Germany), (iii) filtration through a 5 and a 0.1 µm pore diameter filter (Acrodisc®).

Each solution (96 aliquots of 3 µL) was pipetted columnwise into 384-well plates by the liquid handling station. The plates were cooled down quickly to 273 K (0 °C) and, as soon as the temperature was stable for 1 min, at a continuous cooling rate of 1 K min⁻¹ further down to 243 K (-30 °C).

Each solution of the two aliquots of each filter was measured separately, and droplets of the same solution were added to a total droplet number of 192 (2 × 96 droplets) (Figs. 10 and S6). All IN concentrations were calculated per liter air.

The untreated filter extract showed IN activity at relatively high temperatures with an initial freezing temperature of 267 K (-6 °C). The concentration of IN active at temperatures above 263 K (-10 °C) was about 0.001 L⁻¹, but heat treatment led to a loss of IN activity above 263 K (-10 °C). Because the activity of known biological IN results from proteins or proteinaceous compounds (Green and Warren, 1985; Kieft and Ruscetti, 1990; Pouleur et al., 1992; Tsumuki and Konno, 1994) and proteins are known to be heat-sensitive, the results suggest the presence of biological IN. The concentration of IN between 263 K (-10 °C) and 257 K (-16 °C) increased about 2 orders of magnitude and in a sudden increase another 2 orders between 257 K (-16 °C) and 256 K (-17 °C). The IN concentration below 256 K (-17 °C) increased continuously up to about 500 L⁻¹, but heat treatment reduced the IN concentration of up to 1 order of magnitude below 256 K (-17 °C). Filtration experiments did not affect the initial freezing temperature, but the concentration of biological IN decreased significantly. The results suggest the presence of many biological IN or agglomerates larger than 5 µm and of a few biological IN smaller than 0.1 µm. The cumulative number of IN active between 263 K (-10 °C) and 257 K (-16 °C) decreased up to 2 orders of magnitude upon filtration, but the IN concentration below 256 K (-17 °C) was not affected. The findings show that many IN active between 263 K (-10 °C) and 257 K (-16 °C) were larger than 5 µm, whereas IN active below 256 K (-17 °C) were smaller than 0.1 µm.

4 Conclusions

The new high-throughput droplet freezing assay TINA was introduced to study heterogeneous ice nucleation of microliter-range droplets in the immersion mode. TINA provides the analysis of 960 droplets simultaneously or 192 and 768 droplets in two independent experiments at the same time, enabling the analysis of many samples with high statistics in a short period of time. Moreover, an infrared camera-based detection system allows the reliable determination of droplet freezing. The setup was tested with Snomax® as bac-

terial IN and IN filtrate of *Mortierella alpina* as fungal IN. For these reference materials, both the initial freezing temperature and the cumulative number of IN per gram unit mass were in good agreement with the literature, which demonstrates the functionality of the new setup.

TINA was applied to study the effect of O₃ and NO₂ exposure on the IN activity of Snomax®, where high concentrations of O₃ and NO₂ reduced the IN activity significantly. Atmospheric aerosol samples from Mainz (Germany) were analyzed for IN activity to show the applicability of TINA for field samples. Here, the results suggest that most of the biological IN were larger than 5 µm. Moreover, many IN active between 263 K (-10 °C) and 257 K (-16 °C) were larger than 5 µm, whereas IN active below 256 K (-17 °C) were smaller than 0.1 µm. The results confirm that TINA is suitable for high-throughput experiments and efficient analysis of biological IN in laboratory and field samples.

Data availability. All data are available from the corresponding authors upon request.

Supplement. The supplement related to this article is available online at: <https://doi.org/10.5194/amt-11-6327-2018-supplement>.

Author contributions. ATK, ML, and FH developed the instrument. ATK, UP, JFN conceived and designed the experiments. ATK performed the experiments. MLP wrote the code to process the data and did the error calculation. All authors discussed the data and contributed to the writing of the manuscript.

Competing interests. The authors declare that they have no conflict of interest.

Special issue statement. This article is part of the special issue “Results from the ice nucleation research unit (INUIT) (ACP/AMT inter-journal SI)”. It is not associated with a conference.

Acknowledgements. The authors thank Christian Gurk, Thomas Klimach, Florian Rubach, Frank Kunz, and the workshop team for supporting the experimental setup, Nina-Maria Kropf, Carola Sophie Krevert, Isabel Maurus, Gila Maren Kopper, and Petya Yordanova for technical support, and Hinrich Grothe, Thomas Koop, Thomas Berkemeier, Alex Huffman, Daniel Andrew Pickersgill, Naama Lang-Yona, and Jan Frederik Scheel for helpful discussions. The Max Planck Society (MPG) and the Ice Nuclei Research Unit of the Deutsche Forschungsgemeinschaft (DFG FR 3641/1-2, FOR 1525 INUIT) are acknowledged for financial support.

The article processing charges for this open-access publication were covered by the Max Planck Society.

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A. T. Kunert *et al.*: Twin-plate Ice Nucleation Assay (TINA)

Edited by: Mingjin Tang

Reviewed by: two anonymous referees

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Macromolecular fungal ice nuclei in *Fusarium*: effects of physical and chemical processing

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Biogeosciences **16**, 4647-4659, (2019)

Online supplement available at:

<https://doi.org/10.5194/bg-16-4647-2019-supplement>

Author contributions

CEM, JFN, and UP designed the experiments. DGS III and LEH provided fungal cultures. CEM, DGS III, and JFN performed the initial screenings. ATK, KT, CSK, CW, and KRS performed the experiments. ATK, JFN, MLP, and UP discussed the results. ATK and JFN wrote the manuscript with contributions of all coauthors.

Biogeosciences, 16, 4647–4659, 2019
<https://doi.org/10.5194/bg-16-4647-2019>
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Macromolecular fungal ice nuclei in *Fusarium*: effects of physical and chemical processing

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Received: 9 July 2019 – Discussion started: 31 July 2019

Revised: 4 November 2019 – Accepted: 7 November 2019 – Published: 9 December 2019

Abstract. Some biological particles and macromolecules are particularly efficient ice nuclei (IN), triggering ice formation at temperatures close to 0 °C. The impact of biological particles on cloud glaciation and the formation of precipitation is still poorly understood and constitutes a large gap in the scientific understanding of the interactions and coevolution of life and climate. Ice nucleation activity in fungi was first discovered in the cosmopolitan genus *Fusarium*, which is widespread in soil and plants, has been found in atmospheric aerosol and cloud water samples, and can be regarded as the best studied ice-nucleation-active (IN-active) fungus. The frequency and distribution of ice nucleation activity within *Fusarium*, however, remains elusive. Here, we tested more than 100 strains from 65 different *Fusarium* species for ice nucleation activity. In total, ~ 11 % of all tested species included IN-active strains, and ~ 16 % of all tested strains showed ice nucleation activity above –12 °C. Besides *Fusarium* species with known ice nucleation activity, *F. armeniacum*, *F. begoniae*, *F. concentricum*, and *F. langsethiae* were newly identified as IN-active. The cumulative number of IN per gram of mycelium for all tested *Fusarium* species was comparable to other biological IN like *Sarocladium implicatum*, *Mortierella alpina*, and Snomax®. Filtration experiments indicate that cell-free ice-nucleating macromolecules (INMs) from *Fusarium* are smaller than 100 kDa and that molecular aggregates can be formed in solution. Long-term storage and freeze–thaw cycle experiments revealed that the fungal IN in aqueous solution remain active over several months and in the course of repeated freezing and thawing.

Exposure to ozone and nitrogen dioxide at atmospherically relevant concentration levels also did not affect the ice nucleation activity. Heat treatments at 40 to 98 °C, however, strongly reduced the observed IN concentrations, confirming earlier hypotheses that the INM in *Fusarium* largely consists of a proteinaceous compound. The frequency and the wide distribution of ice nucleation activity within the genus *Fusarium*, combined with the stability of the IN under atmospherically relevant conditions, suggest a larger implication of fungal IN on Earth's water cycle and climate than previously assumed.

1 Introduction

Ice particles in the atmosphere are formed either by homogeneous nucleation at temperatures below –38 °C or by heterogeneous nucleation catalyzed by particles or macromolecules serving as ice nuclei (IN) at warmer temperatures (Pruppacher and Klett, 1997; reviewed in detail in Fröhlich-Nowoisky et al., 2016 and Knopf et al., 2018). Biological particles in particular are expected to play an important role as IN in the temperature range from –15 to 0 °C, but the impact of biological particles on cloud glaciation and the formation of precipitation is still poorly understood (Coluzza et al., 2017). Several studies suggest a triggering effect of biological IN for cloud glaciation and formation of precipitation (Creamean et al., 2013; DeMott and Prenni, 2010; Failor et al., 2017; Hanlon et al., 2017; Joly et al., 2014;

Petters and Wright, 2015; Pratt et al., 2009; Stopelli et al., 2015, 2017), and former studies have shown that biological particles are more efficient than mineral IN (DeMott and Prenni, 2010; Després et al., 2012; Hill et al., 2014; Hoose and Möhler, 2012; Huffman et al., 2013; Möhler et al., 2007; Morris et al., 2014; Murray et al., 2012; Pratt et al., 2009).

The best characterized biological IN are common plant-associated bacteria of the genera *Pseudomonas*, *Pantoea*, and *Xanthomonas* (Garnham et al., 2011; Govindarajan and Lindow, 1988; Graether and Jia, 2001; Green and Warren, 1985; Hill et al., 2014; Kim et al., 1987; Ling et al., 2018; Šantl-Temkiv et al., 2015; Schmid et al., 1997; Wolber et al., 1986), and, recently, an ice-nucleation-active (IN-active) *Lysinibacillus* was found (Failor et al., 2017). The first identified IN-active fungi were strains of the genus *Fusarium* (Hasegawa et al., 1994; Pouleur et al., 1992; Richard et al., 1996; Tsumuki et al., 1992). To date, a few more fungal genera with varying initial freezing temperatures such as *Isaria farinosa* ($\sim -4^{\circ}\text{C}$), *Mortierella alpina* ($\sim -5^{\circ}\text{C}$), *Puccinia* species (-4 to -8°C), and *Sarocladium* (formerly named *Acremonium*) *implicatum* ($\sim -9^{\circ}\text{C}$) have been identified as IN-active (Fröhlich-Nowoisky et al., 2015; Huffman et al., 2013; Morris et al., 2013; Richard et al., 1996).

The genus *Fusarium* is cosmopolitan and includes saprophytes and pathogens of plants and animals (Leslie and Summerell, 2006; Nelson et al., 1994). Although they are considered to be primarily soilborne fungi, many species of *Fusarium* are airborne (Prussin et al., 2014; Schmale et al., 2012; Schmale and Ross, 2015), and they were found in atmospheric and cloud water samples (e.g., Amato et al., 2007; Fröhlich-Nowoisky et al., 2009; Fulton, 1966). Some species can cause wilts, blights, root rots, and cankers in agriculturally important crops worldwide (e.g., Schmale and Gordon, 2003; Wang and Jeffers, 2000). Other species can produce secondary metabolites known as mycotoxins that can cause a variety of acute and chronic health effects in humans and animals (e.g., Bush et al., 2004; Ichinoe et al., 1983).

While the factors for a positive selective pressure for ice nucleation activity in *Fusarium* and other fungi have not been directly identified, an ecological advantage of initiating ice formation is easily conceivable. Indeed, most IN-active bacteria and fungi are isolated from regions with seasonal temperatures below 0°C (Diehl et al., 2002; Schnell and Vali, 1972). Ice nucleation activity at temperatures close to 0°C could be beneficial for pathogens or might provide an ecological advantage for saprophytic *Fusarium* species by facilitating in the acquisition of nutrients liberated during cell rupture of the host (Lindow et al., 1982). Furthermore, IN on the surface of the mycelium could avoid physical damage of the fungus by protective extracellular freezing (Fröhlich-Nowoisky et al., 2015; Zachariassen and Kristiansen, 2000) or by binding moisture as ice in cold and dry seasons (Pouleur et al., 1992). With increasing temperatures, the retained water can be of advantage in early vegetative periods and for bacterial movement on the mycelial water film known as the fun-

gal highway (Kohlmeier et al., 2005; Warmink et al., 2011). Moreover, ice nucleation activity might be beneficial for airborne *Fusarium* and for their return to Earth's surface under advantageous conditions in a feedback cycle known as bioprecipitation (Després et al., 2012; Morris et al., 2013, 2014; Sands et al., 1982). In addition, once the IN are released into the environment, they can adsorb to clay and might also be available in the atmosphere associated with soil dust particles (Conen et al., 2011; Fröhlich-Nowoisky et al., 2015, 2016; Hill et al., 2016; O'Sullivan et al., 2014, 2015, 2016; Sing and Sing, 2010).

The sources, abundance, and identity of biological IN are not well characterized (Coluzza et al., 2017), and it has been proposed that systematic surveys will likely increase the number of IN-active fungal species discovered (Fröhlich-Nowoisky et al., 2015). *Fusarium* is the best-known IN-active fungus, but the frequency and distribution of ice nucleation activity within *Fusarium* is not well known. In this study, more than 100 strains from 65 different *Fusarium* species were tested for ice nucleation activity in three laboratories with different freezing methods. A high-throughput droplet freezing assay was used to quantify the IN of selected *Fusarium* species, and filtration experiments were performed to estimate the size of the *Fusarium* IN. Furthermore, the stability of *Fusarium* IN upon exposure to ozone and nitrogen dioxide, under high and low or quickly changing temperatures, and after short- and long-term storage under various conditions was investigated.

2 Materials and methods

2.1 Origin and growth conditions of fungal cultures

Thirty *Fusarium* strains from USDA-ARS, Michigan State University (Linda E. Hanson, East Lansing, MI, USA), 13 strains from the Schmale Laboratory at Virginia Tech (David G. Schmale, Blacksburg, VA, USA), and 69 strains from the Kansas State University Teaching Collection (John F. Leslie, Manhattan, KS, USA) were screened for ice nucleation activity (Table S1 in the Supplement).

The strains from the USDA-ARS, Michigan State University, were collected from crop tissue (sugar beet). All isolates were from field-grown beets and were obtained by hyphal tip transfer. The strains from the Schmale Laboratory at Virginia Tech were collected with unmanned aircraft systems (UASs or drones) equipped with remotely operated sampling devices containing a *Fusarium* selective medium (e.g., Lin et al., 2013, 2014). All of the Schmale Laboratory strains were collected 100 m above ground level at the Kentland Farm in Blacksburg, Virginia, USA. Detailed information is not available for the sources of the strains for the Kansas State University Teaching collection. However, some of these strains are holotype strains referenced in Leslie and Summerell (2006).

The strains from the USDA-ARS, Michigan State University, were cultivated on dextrose peptone yeast extract agar, containing 10 g L⁻¹ dextrose (VWR, Radnor, PA, USA), 3 g L⁻¹ peptone (Difco Proteose Peptone No. 3, Becton, Dickinson and Company, Franklin Lakes, NY, USA), and 0.3 g L⁻¹ yeast extract (Merck, Kenilworth, NJ, USA), and were filtered through a 0.2 µm pore diameter filter (PES disposable filter units, Life Science Products, Frederick, CO, USA). After filtration, 12 g L⁻¹ agarose (Certified Molecular Biology Agarose, Bio-Rad, Hercules, CA, USA) was added, and the medium was sterilized by autoclaving at 121 °C for 20 min. The colonies were grown at 22 to 24 °C for 7 to 19 d. The strains from the Schmale Laboratory at Virginia Tech and the Kansas State University Teaching Collection were maintained in cryogenic storage at -80 °C and were grown on quarter-strength potato dextrose agar (Difco Laboratories, Detroit, USA) on 100 mm petri plates at ambient room temperature for 4 d prior to ice nucleation assays.

For quantitative analysis, exposure experiments, heat treatments, freeze-thaw cycles, and short- and long-term storage tests, a selection of IN-active tested strains was grown on full-strength potato dextrose agar (VWR International GmbH, Darmstadt, Germany) first at room temperature for 4 to 6 d and then at 6 °C for about 4 weeks. For filtration experiments, the fungal cultures were grown at 6 °C for up to 6 months.

2.2 Preparation and treatments of aqueous extracts

For LED-based Ice Nucleation Detection Apparatus (LINDA) (Stopelli et al., 2014) experiments (see Sect. 2.3), 4 mL of sterile 0.9 % NaCl was added to each of eight petri plates, and the fungal cultures were scraped with the flat end of a sterile bamboo skewer. The resulting suspension of mycelium and spores was filtered through a 100 µm filter (Corning Life Sciences, Reims, France).

For Twin-plate Ice Nucleation Assay (TINA) (Kunert et al., 2018) experiments (see Sect. 2.3) the fungal mycelium was scraped off the agar plate and transferred into a 15 mL tube (Greiner Bio One, Kremsmünster, Austria). The fresh weight of the mycelium was determined gravimetrically. Pure water was prepared as described in Kunert et al. (2018). Aliquots of 10 mL pure water were added before vortexing three times at 2700 rpm for 30 s (Vortex-Genie 2, Scientific Industries, Inc., Bohemia, NY, USA) and centrifugation at 4500 g for 10 min (Heraeus Megafuge 40, Thermo Scientific, Braunschweig, Germany). For all experiments, the aqueous extract were filtered successively through a 5 and a 0.1 µm PES syringe filter (Acrodisc®, Sigma-Aldrich, Taufkirchen, Germany), and the aqueous extracts contained IN from spores and mycelial surfaces.

For filtration experiments, the 0.1 µm filtrate was further filtered successively through 300 000 and 100 000 MWCO PES ultrafiltration units (Vivaspin®, Sartorius AG, Göttingen, Germany).

After each filtration step, the IN concentration was determined using TINA.

For exposure experiments, aqueous extracts of *F. acuminatum* 3–68 and *F. avenaceum* 2–106 were exposed to high concentrations of O₃ and NO₂ as described in Liu et al. (2017). Briefly, a mixture of 1 ppm O₃ and 1 ppm NO₂ was bubbled through 1 mL aliquots of aqueous extract for 4 h, and the IN concentration was determined using TINA.

For heat treatment experiments, aliquots of aqueous extracts were incubated at 40, 70, and 98 °C for 1 h for each of *F. acuminatum* 3–68, *F. armeniacum* 20 970, *F. avenaceum* 2–106, and *F. langsethiae* 19 084. The IN concentration was determined using TINA.

For freeze-thaw cycles, the ice nucleation activity of *F. acuminatum* 3–68 was determined shortly after preparation of the aqueous extract and after storage at 6 °C for 24 h using TINA. Then, the aqueous extract was stored at -20 °C for 24 h and thawed again. The ice nucleation activity was tested before storage at -20 °C for an additional 24 h. After thawing, the ice nucleation activity was determined again.

For long-term storage experiments, the aqueous extracts of various *Fusarium* species were stored at 6 °C for about 4 months or at -20 °C for about 8 months, and the ice nucleation activity was determined using TINA.

2.3 Ice nucleation assays

Two independent droplet freezing assays conducted in two laboratories were used to investigate the distribution of ice nucleation activity within *Fusarium* in an initial screening.

First, a thermal cycler (PTC200, MJ Research, Hercules, CA, USA) was used as described in Fröhlich-Nowoisky et al. (2015) to screen 30 *Fusarium* strains from seven species from USDA-ARS, Michigan State University, in the temperature range from -2 to -9 °C. Mycelium was picked with sterile pipette tips (Eppendorf, Westbury, NY, USA) into 80 µL aliquots of 0.2 µm pore diameter filtered dextrose peptone yeast (DPY) broth in sterile 96-well polypropylene PCR plates (VWR International, LLC, Radnor, PA, USA). Up to seven droplets were measured for each sample, and the mean freezing temperature was calculated. Aliquots of uninoculated DPY broth were used as negative controls, which did not freeze in the investigated temperature interval.

Second, the LED-based Ice Nucleation Detection Apparatus (LINDA) was used as described by Stopelli et al. (2014) to screen 13 strains from the Schmale Laboratory at Virginia Tech and 69 strains from the Kansas State University Teaching Collection. Aliquots of 200 µL of each aqueous extract were transferred to three separate 500 µL tubes and placed on ice for 1 h prior to the LINDA experiments. LINDA was run from -1 to -20 °C, and images of the samples were recorded every 6 s. The mean freezing temperature for three droplets was calculated. Note that the aqueous extracts were prepared in 0.9 % NaCl solution, which could reduce the freezing temperatures by 0.5 °C based on theoret-

ical calculations. We cannot exclude, however, that the high concentration of IN compensates for the effect of NaCl on the freezing temperature. This is supported by the investigations of Stopelli et al. (2014), who did not find a systematic suppression of freezing at this salt concentration in LINDA experiments. As a negative control, a 0.9 % NaCl solution was added to three uninoculated agar plates, and the freezing started below -14°C . As positive control, aqueous suspensions of *Pseudomonas syringae* CC94 from the collection of INRA (Avignon, France) (Berge et al., 2014), with a final OD_{580} of 0.5 to 0.7, i.e. $\sim 10^9$ bacteria mL^{-1} , were used for each experiment. The bacteria were grown on King's medium B (King et al., 1954) at 22 to 25°C for 48 h, and aqueous suspensions were equilibrated at 4°C for 1 to 4 h before LINDA experiments. The freezing temperatures of *P. syringae* CC94 ranged from -3.5 to -4.6°C .

Ice nuclei of selected *Fusarium* species, which were long known for ice nucleation activity (*F. acuminatum*, *F. avenaceum*), as well as all the newly identified species, were further analyzed in immersion freezing mode using the high-throughput Twin-plate Ice Nucleation Assay (TINA) (Kunert et al., 2018). The aqueous extracts were serially diluted 10-fold with pure water by a liquid handling station (epMotion ep5073, Eppendorf, Hamburg, Germany) to a dilution at which droplets remained liquid in the investigated temperature interval. Of each dilution, 96 droplets ($3\ \mu\text{L}$) were tested with a continuous cooling rate of $1^{\circ}\text{C}\ \text{min}^{-1}$ from 0 to -20°C . Pure water samples ($0.1\ \mu\text{m}$ filtrated) served as a negative control for each experiment. These did not freeze in the observed temperature interval. The temperature was measured with an accuracy of $0.2\ \text{K}$ (Kunert et al., 2018). The obtained fraction of frozen droplets (f_{ice}) and the counting error were used to calculate the cumulative number of IN (N_{m}) with the associated error using the Vali formula and the Gaussian error propagation (Kunert et al., 2018; Vali, 1971). For each experiment, the cumulative number of IN was averaged over all dilutions. If the experiment was repeated, the cumulative number of IN was averaged over all experiments, and the standard error was calculated. Three independent experiments with aqueous extracts from three individual fungal culture plates of the same strain showed similar results with only slight variation. An example of results is presented for *F. armeniacum* 20970 (Fig. S1 in the Supplement).

3 Results and discussion

3.1 IN-active *Fusarium* species

Although several IN-active *Fusarium* species are known, the frequency and distribution of ice nucleation activity within the fungal genus *Fusarium* are still not well studied (Hasegawa et al., 1994; Humphreys et al., 2001; Pouleur et al., 1992; Richard et al., 1996; Tsumuki and Konno, 1994). Two initial screenings in the temperature range from -1 to

Table 1. Ice-nucleation-active *Fusarium* strains with corresponding mean freezing temperatures of the initial screening. The newly identified IN-active *Fusarium* species are marked with an asterisk (*).

Species	Strain	T ($^{\circ}\text{C}$)
<i>F. acuminatum</i>	1–3	-5.6
<i>F. acuminatum</i>	1–4	-5.0
<i>F. acuminatum</i>	1–5	-5.6
<i>F. acuminatum</i>	1–24	-3.5
<i>F. acuminatum</i>	2–38	-5.0
<i>F. acuminatum</i>	2–48	-5.6
<i>F. acuminatum</i>	2–109	-5.6
<i>F. acuminatum</i>	3–48	-5.0
<i>F. acuminatum</i>	3–68	-3.5
<i>F. acuminatum</i>	20964	-6.2
<i>F. armeniacum</i> *	20970	-5.3
<i>F. avenaceum</i>	2–106	-5.0
<i>F. avenaceum</i>	11 440	-7.6
<i>F. begoniae</i> *	10 767	-11.2
<i>F. concentricum</i> *	10 765	-4.6
<i>F. langsethiae</i> *	19 084	-9.4
<i>F. tricinctum</i>	20 990	-7.3

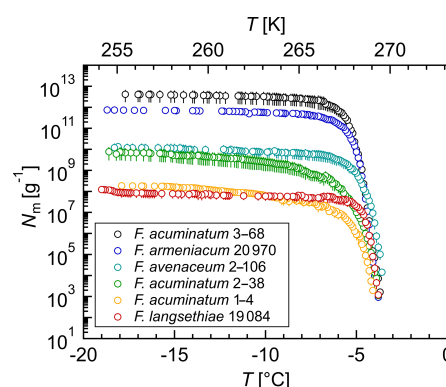


Figure 1. Overview of ice nucleation activity for selected *Fusarium* species and strains: cumulative number of IN (N_{m}) per gram of mycelium plotted against the temperature (T); arithmetic mean values and standard error of two independent experiments with aqueous extracts from two individual fungal culture plates of the same species.

-20°C were performed to better evaluate the frequency of ice nucleation activity within *Fusarium*. A strain was defined as IN-active, when it initiated ice formation above -9°C (thermal cycler) and -12°C (LINDA), respectively.

In total, $\sim 16\%$ (18/112) of the tested strains showed ice nucleation activity with mean freezing temperatures of -3.5 to -11.2°C (Table 1) in the typical range known for *Fusarium* (-1 and -9°C) (Hasegawa et al., 1994; Humphreys et al., 2001; Pouleur et al., 1992; Richard et al., 1996; Tsumuki et al., 1992; Tsumuki and Konno, 1994). Most

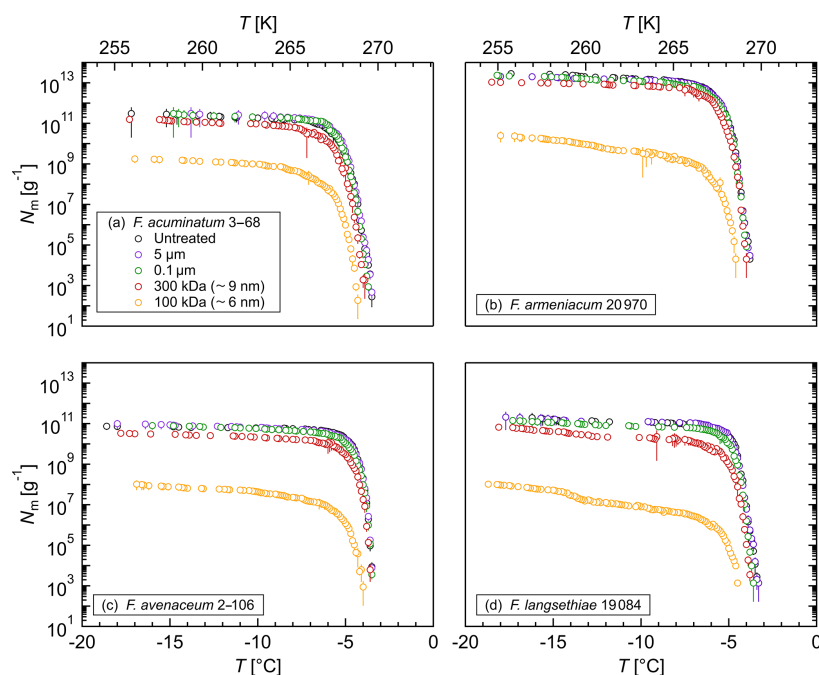


Figure 2. Size determination of the *Fusarium* IN upon filtration: cumulative number of IN (N_m) per gram of mycelium plotted against the temperature (T) for (a) *F. acuminatum* 3–68, (b) *F. armeniacum* 20970, (c) *F. avenaceum* 2–106, and (d) *F. langsethiae* 19084. The error bars were calculated using the counting error and the Gaussian error propagation.

formerly reported initial freezing temperatures were obtained with different *Fusarium* strains, growth conditions, and freezing assays, which might explain differences compared to our results. The high proportion of IN-active strains within *F. acuminatum* is consistent with previous reports (Pouleur et al., 1992; Tsumuki et al., 1995). Overall, $\sim 11\%$ (7/65) of the tested species included IN-active strains. In addition to strains from *Fusarium* species with known ice nucleation activity, four *Fusarium* species were newly identified as IN-active: *F. armeniacum*, *F. begoniae*, *F. concentricum*, and *F. langsethiae*. In further experiments, the ice nucleation activity of *F. begoniae* and *F. concentricum* could not be verified.

The newly identified IN-active species are cosmopolitan. *Fusarium armeniacum* is a toxigenic saprophyte (Burgess et al., 1993), causing seed and root rot on soybeans (Ellis et al., 2012). The geographical distribution has been reported as tropical and subtropical (Leslie and Summerell, 2006), but it was also found in Minnesota, USA (Kommedahl et al., 1979), and Australia (Burgess et al., 1993). *Fusarium begoniae* is a plant pathogen of *Begonia* found in Germany with a potential wider distribution (Nirenberg and O'Donnell, 1998). *Fusarium concentricum* is a plant pathogen, which is frequently found in Central America and isolated from bananas (Aoki et al., 2001; Leslie and Summerell, 2006), and *F. langsethiae* is a broadly distributed cereal pathogen (Torpe

and Nirenberg, 2004). Some strains of these newly identified IN-active species are known to produce mycotoxins, which can threaten the health of humans and animals (Fotso et al., 2012; Kokkonen et al., 2012; Wing et al., 1993a, b).

The results suggest that the ice nucleation activity within *Fusarium* is more widespread than previously known. Not all *Fusarium* species include IN-active strains and not all strains within one species show ice nucleation activity. Earlier studies including experiments suggested that *Fusarium* IN are proteins or at least contain a proteinaceous compound (Hasegawa et al., 1994; Pouleur et al., 1992; Tsumuki and Konno, 1994). Their production requires energy, and we might assume that this trait would not be expressed or maintained unless there was an ecological advantage. It is known that *Fusarium* can regulate the gene expression for IN production, depending on environmental conditions such as nutrient availability (Richard et al., 1996), and some *Fusarium* species reduce or lose their ice nucleation activity after several subcultures (Pummer et al., 2013; Tsumuki et al., 1995). Thus, we cannot exclude that all *Fusarium* strains have the ability to produce IN. From the phylogenetic distribution of ice nucleation activity across the genus *Fusarium*, we can speculate that ice nucleation activity is a very old trait, but either the gene expression requires a trigger, which is not yet identified, or the trait might be in the process of being lost. It is unlikely, however, that the age of the genetic determi-

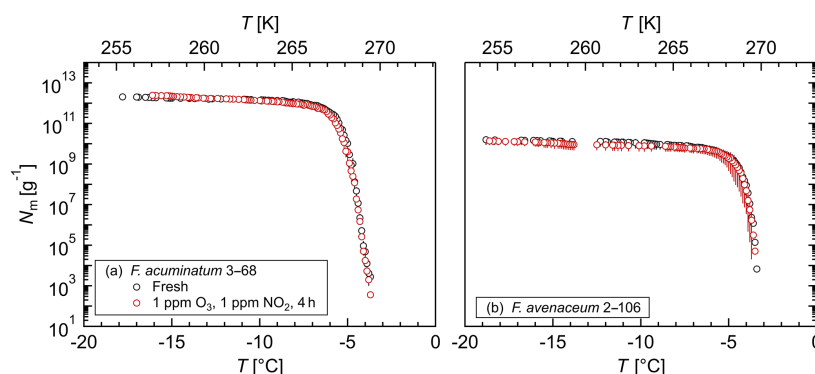


Figure 3. Exposure of aqueous extract from *Fusarium* to ozone and nitrogen dioxide: cumulative number of IN (N_m) per mass of mycelium plotted against the temperature (T) for (a) *F. acuminatum* 3–68 and (b) *F. avenaceum* 2–106; arithmetic mean values and standard error of two independent experiments with aqueous extracts from two individual fungal culture plates of the same species.

nants of fungal ice nucleation activity is older than that in bacteria, since fungi diverged well after the age that has been attributed to the bacterial IN gene (Morris et al., 2014), and the genetic determinants are not the same as those in bacteria.

3.2 Quantification and size determination of IN from selected *Fusarium* species

A selection of IN-active *Fusarium* species was further investigated by extensive droplet freezing assay analysis using TINA. All tested *Fusarium* strains initiated ice nucleation between -3 and -4 °C (Fig. 1). Differences in the freezing temperatures between the initial screening and the quantitative analysis can be due to different growth conditions and freezing assays. The cumulative number of IN (N_m) per gram of mycelium was in the range between 10^8 g⁻¹ and 10^{13} g⁻¹. *Fusarium acuminatum* 3–68 showed the highest ice nucleation activity and *F. langsethiae* the lowest per gram of mycelium. The results are comparable to other IN-active microorganisms like *Sarocladium implicatum* (10^8 g⁻¹, Pummer et al., 2015), *Mortierella alpina* (10^9 g⁻¹, Fröhlich-Nowoisky et al., 2015; 10^{10} g⁻¹, Kunert et al., 2018), and the bacterial IN-active substance Snomax[®] containing *Pseudomonas syringae* (10^{12} g⁻¹, Budke and Koop, 2015; Kunert et al., 2018).

The size of the *Fusarium* IN was investigated by filtration experiments. Filtration through a 5 and a 0.1 µm filter did not affect the ice nucleation activity (Fig. 2), revealing that *Fusarium* IN are smaller than 100 nm, cell-free, easily removed from the fungus, and stay active in solution. This is in agreement with other *Fusarium* studies (O’Sullivan et al., 2015; Pouleur et al., 1992; Tsumuki and Konno, 1994). Moreover, biological ice-nucleating macromolecules (INMs) smaller than 200 nm were also found in various organisms, e.g., other fungi (Fröhlich-Nowoisky et al., 2015; Pummer et al., 2015); leaves, bark, and pollen from birch trees (*Betula* spp.) (Felgitsch et al., 2018; Pummer et al., 2012); leaf

litter (Schnell and Vali, 1973); some microalgae (Tesson and Šantl-Temkiv, 2018); strains of *Lysinibacillus* (Failor et al., 2017); and biological particles in the sea surface microlayer (Irish et al., 2019; Wilson et al., 2015). Filtration through a 300 000 MWCO filter unit decreased the cumulative number of IN per gram of mycelium by about 50 % to 75 % depending on the *Fusarium* species, but a tremendous number of IN (10^{10} – 10^{13} g⁻¹) still passed through the filter. The initial freezing temperature was slightly shifted towards lower temperatures. Further filtration through a 100 000 MWCO filter unit reduced the IN number to 10^8 – 10^{10} g⁻¹, which is less than 1 % of the initial IN concentration. Additionally, the initial freezing temperature was shifted about 1 °C towards lower temperatures.

As ice nucleation activity was found in all filtrates, the aqueous extract of *Fusarium* consists of a mixture of IN-active proteins with different sizes. We hypothesize that *Fusarium* IN are macromolecules (INMs) smaller than 100 kDa, which agglomerate to large protein complexes in solution. Some of these complexes fall apart upon filtration, so that the INMs can pass through the filter. The small shift in the initial freezing temperature suggests that these INMs reassemble again to aggregates after filtration, as larger IN nucleate at warmer temperatures (Govindarajan and Lindow, 1988; Pummer et al., 2015). Erickson (2009) determined the size of proteins based on theoretical calculations. As the interior of proteins is closely packed with no substantial holes and almost no water molecules inside, proteins are rigid structures with approximately the same density (~ 1.37 g cm⁻³). Assuming the protein as a smooth spherical particle, the minimum diameter of the INM would be smaller than 6.1 nm. Our results are in accordance with Lagzian et al. (2014), who cloned and expressed a 49 kDa IN-active protein from *F. acuminatum*.

As *Fusarium* IN are cell-free and can easily be washed off the fungal surface, they can be released in high num-

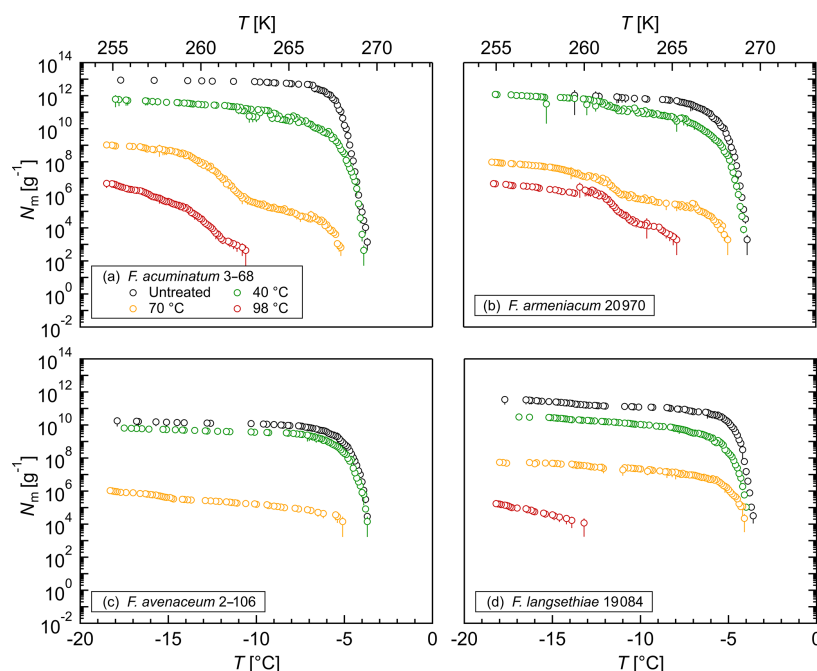


Figure 4. Effects of high temperatures on the ice nucleation activity of *Fusarium*: cumulative number of IN (N_m) per gram of mycelium plotted against the temperature (T) for (a) *F. acuminatum* 3–68, (b) *F. armeniacum* 20970, (c) *F. avenaceum* 2–106, and (d) *F. langsethiae* 19084. The error bars were calculated using the counting error and the Gaussian error propagation.

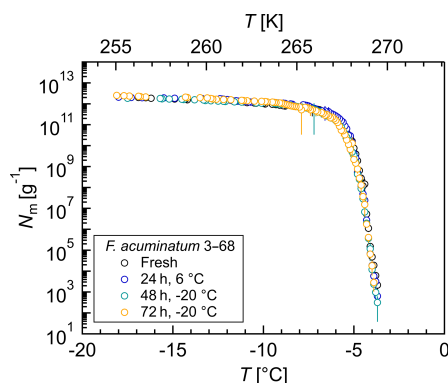


Figure 5. Effects of short-term storage and freeze–thaw cycles on the ice nucleation activity of *Fusarium acuminatum* 3–68: cumulative number of IN (N_m) per gram of mycelium plotted against the temperature (T). The same aqueous extract was measured immediately after preparation (black), after storage at 6 °C for 24 h (blue), after another 24 h stored at –20 °C (total 48 h; turquoise), and after another 24 h stored at –20 °C (total 72 h; yellow). The error bars were calculated using the counting error and the Gaussian error propagation.

bers into the environment. If they are not degraded by microorganisms before, the IN can adsorb to soil dust and be aerosolized while attached to these particles (Conen et al., 2011; Fröhlich-Nowoisky et al., 2015; Hill et al., 2016; O’Sullivan et al., 2014, 2015, 2016; Sing and Sing, 2010). This is in good agreement with Pruppacher and Klett (1997), who found a positive correlation between IN number concentration and particles in the coarse mode. Other releasing processes cannot be excluded; however, it is unlikely that the INMs are present in the atmosphere as individual aerosol particles. Individual proteins with a diameter of ~ 6 nm, which may enter the atmosphere, would be in the nucleation mode size range, where particles tend to uptake gaseous compounds and grow to Aitken mode particles, which themselves tend to coagulate to larger agglomerates (Seinfeld and Pandis, 1998).

3.3 Stability of *Fusarium* IN

In the atmosphere, IN can interact with other aerosol particles or gases. They can be exposed to chemically modifying agents like ozone and nitrogen dioxide, as well as physical stressors like high and low or quickly changing temperatures. To investigate the stability of *Fusarium* IN, we performed exposure experiments, heat treatments, freeze–thaw cycles, and long-term storage tests.

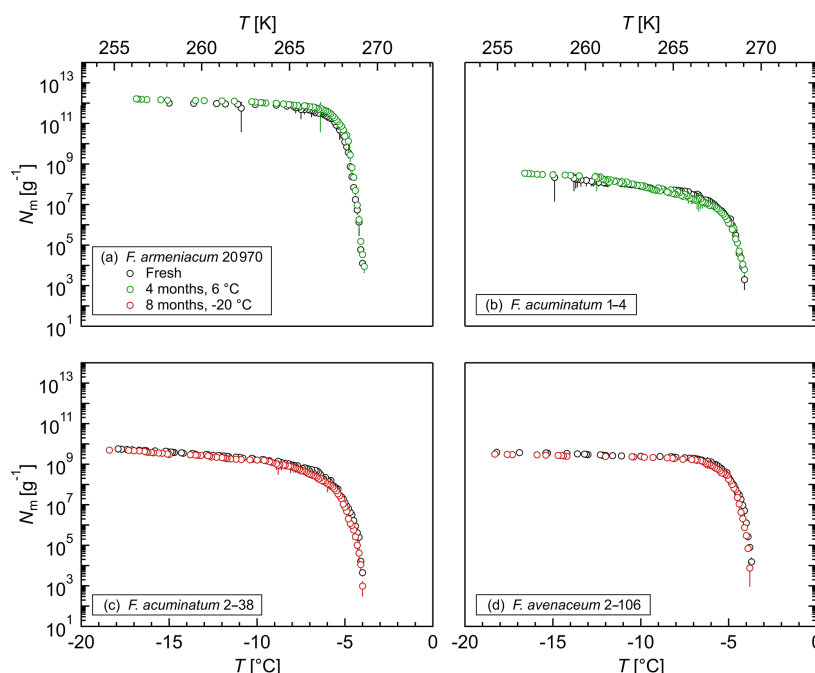


Figure 6. Effect of long-term storage on the ice nucleation activity of (a) *F. armeniacum* 20970, (b) *F. acuminatum* 1–4, (c) *F. acuminatum* 2–38, and (d) *F. avenaceum* 2–106: cumulative number of IN (N_m) per gram of mycelium plotted against the temperature (T). The error bars were calculated using the counting error and the Gaussian error propagation.

The influence of chemical processing on the *Fusarium* IN, in particular oxidation and nitration reactions as occurring during atmospheric aging, was investigated by exposing aqueous extracts from *F. acuminatum* 3–68 and *F. avenaceum* 2–106 to high concentrations of ozone and nitrogen dioxide in liquid phase. Figure 3 shows that for both species neither the initial freezing temperature nor the cumulative number of IN per gram of mycelium was affected by exposure. These results demonstrate a high stability of *Fusarium* IN under oxidizing and nitrating conditions. This is in contrast to other biological IN, e.g., bacterial IN (Snomax®) (Kunert et al., 2018), birch and alder pollen (Gute and Abbatt, 2018), and dissolved organic matter (Borduas-Dedekind et al., 2019), where exposure to oxidizing agents reduced the IN activity.

The stability of the INM in *Fusarium* was investigated in heat treatment experiments. The ice nucleation activity was reduced significantly at a 40 °C treatment (Fig. 4). Between 40 % and 90 % of IN were lost at this temperature depending on the species, which supports the hypothesis that the INM in *Fusarium* consists of a proteinaceous compound. A heat treatment at 70 °C reduced the ice nucleation activity to less than 0.01 % compared to the initial level. Moreover, the initial freezing temperature was shifted to lower temperatures, indicating a breakdown of the large protein aggregates. After a 98 °C treatment, we still found ice nucleation activity for all investigated species except for *F. avenaceum* 2–106. The

results are in agreement with previous studies, which also reported a reduction in ice nucleation activity with increasing temperature in heat treatment experiments (Hasegawa et al., 1994; Pouleur et al., 1992; Tsumuki and Konno, 1994). The remaining activity after the 98 °C treatment, however, could indicate that posttranslational modifications like glycosylation and therefore polysaccharides could play a role in the ice nucleation activity of *Fusarium*. Further systematic studies including chemical analyses are needed for elucidation.

To study the effects of short-term storage and freeze–thaw cycles on the ice nucleation activity of *F. acuminatum* 3–68, IN measurements of the same aqueous extract were performed at different time points (Fig. 5). The results of freshly prepared aqueous extract revealed that the highest activity of fungal IN was already developed during preparation of the filtrate and no time for equilibration was required. Storage of aqueous extract at 6 °C for 24 h did not affect the ice nucleation activity. Also, further storage at –20 °C for another 24 h and repeated freeze–thaw cycles had no impact on the ice nucleation activity. This means that, once in the atmosphere, the IN can undergo several freeze–thaw cycles without losing their activity and are still able to influence cloud glaciation and the formation of precipitation. This could be an explanation for why not all fungi are always IN-active as their IN are highly stable and quasi-recyclable. Ice nuclei might influence the availability of moisture over long time

periods, and if enough moisture is available in the environment, the necessity of IN production would be omitted, and the fungus could save energy.

In addition, the stability of the INM in *Fusarium* was studied in long-term storage tests, where aqueous extracts of various *Fusarium* species were stored at different temperatures for a long period of time. Figure 6 shows that storage at 6 °C for 4 months and at −20 °C for 8 months did not influence the ice nucleation activity of *F. armeniacum* 20970, *F. acuminatum* 1–4, *F. avenaceum* 2–106, or *F. acuminatum* 2–38. The results demonstrate the high stability of the INMs in *Fusarium* in liquid and frozen solutions over long time periods, which makes *Fusarium* well applicable for laboratory IN studies. Moreover, the high stability is likely an advantage for these fungi to be linked to atmospheric processes.

4 Conclusions

The frequency and distribution of ice nucleation activity within the fungal genus *Fusarium* was investigated in a screening of more than 100 strains from 65 different *Fusarium* species. In total, ~ 11 % (7/65) of all tested species included IN-active strains, and ~ 16 % (18/112) of all tested strains showed ice nucleation activity, demonstrating the wide distribution of ice nucleation activity within *Fusarium*. Filtration experiments suggest that *Fusarium* IN form aggregates consisting of INMs smaller than 100 kDa (~ 6 nm). Exposure experiments, freeze–thaw cycles, and long-term storage tests revealed a high stability of the INMs in *Fusarium*, demonstrating the suitability of *Fusarium* in laboratory IN studies. Heat treatments at 40 to 98 °C reduced the IN concentration significantly, supporting the hypothesis that the INM in *Fusarium* largely consists of a proteinaceous compound. An involvement of polysaccharides, however, cannot be excluded. The wide distribution of ice nucleation activity within the genus *Fusarium*, together with the stability of the INM in *Fusarium* under atmospherically relevant conditions, suggests that the implication of these IN on Earth's water cycle and climate might be more significant than previously assumed. Additional research is necessary to characterize the INMs in *Fusarium* and processes which can result in their agglomeration to larger protein complexes. To evaluate the implication of these IN on Earth's climate, additional work is required to study the abundance of *Fusarium* IN in environmental samples on a global scale.

Data availability. All data are available from the corresponding author upon request.

Supplement. The supplement related to this article is available online at: <https://doi.org/10.5194/bg-16-4647-2019-supplement>.

Author contributions. CEM, JFN, and UP designed the experiments. DGS III and LEH provided fungal cultures. CEM, DGS III, and JFN performed the initial screenings. ATK, KT, CSK, CW, and KRS performed the experiments. ATK, JFN, MLP, and UP discussed the results. ATK and JFN wrote the manuscript with contributions of all coauthors.

Competing interests. The authors declare that they have no conflict of interest.

Acknowledgements. We thank Claudia Bartoli, Jan-David Förster, Tom Godwill, Nina-Maria Kropf, and Emiliano Stopelli for technical assistance; Gary D. Franc, Thomas C. J. Hill, Kathrin Reinmuth-Selzle, Beatriz Sánchez-Parra, Jan F. Scheel, and Michael G. Weller for helpful discussions; and the Max Planck Society (MPG) and the Deutsche Forschungsgemeinschaft (DFG, FR3641/1-2, FOR 1525 INUIT) for financial support. This work is dedicated to the memory of Gary D. Franc, whose pioneering work in atmospheric microbiology has been an inspiration for this work.

Financial support. The article processing charges for this open-access publication were covered by the Max Planck Society.

Review statement. This paper was edited by Carol Robinson and reviewed by three anonymous referees.

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To Freeze or Not to Freeze: Inhibition of Bacterial Ice Nucleators by Antifreeze Proteins

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submitted to the Journal of Physical Chemistry Letters (2020)

Contribution

Aided in freezing experiments, data analysis, and writing of the manuscript.

Title:

**To Freeze or Not to Freeze: Inhibition of Bacterial Ice Nucleators by
Antifreeze Proteins**

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Keywords:

Antifreeze Proteins, Ice-Nucleating Proteins, Protein Interactions, Heterogeneous Ice Nucleation, Snomax[®]

Abstract:

Cold-adapted organisms use antifreeze proteins (AFPs) or ice-nucleating proteins (INPs) for the survival in freezing habitats. AFPs have been reported to be able to inhibit the activity of INPs, a property that would be of great physiological relevance. The generality of this effect is not understood, and for the few known examples of INP inhibition by AFPs, the molecular mechanisms remain unclear. Here, we report a comprehensive evaluation of the effects of all major classes of AFPs on the activity of bacterial ice nucleators using a high-throughput ice nucleation assay. We find that bacterial INPs are inhibited by certain AFPs, while others show no effect. Thus, the ability to inhibit the activity of INPs is not an intrinsic property of AFPs, and the interactions of INPs and different AFPs proceed through protein-specific rather than universal molecular mechanisms.

Ice formation is the most important liquid-to-solid phase transition on earth and is critical for fields as diverse as cryobiology, geology, and climate science ¹⁻². The crystallization of water into ice and the process of ice recrystallization upon thawing are further major contributors to cell death and lethal to most organisms ³. Ice nucleation-active bacteria are a primary cause of frost damage to plants, and in the earth's atmosphere, they can contribute to the formation of snow and rain ⁴⁻⁶. The formation of ice is thermodynamically favored in water at temperatures below 0 °C, but the initial crystallization is kinetically hindered ⁷. As a result, pure water can be supercooled to temperatures as low as -46 °C, below which homogenous ice nucleation occurs ⁸. Freezing of water in biological systems is a heterogeneous process, facilitated by the presence of ice-nucleating agents of biological and abiotic origins ⁹⁻¹¹. The most efficient biological ice nucleators are ice-nucleating proteins (INPs) from bacteria such as *Pseudomonas syringae* and *Erwinia herbicola*, which can initiate ice formation at temperatures close to 0 °C ¹²⁻¹³.

Most known bacterial ice nucleators consist of large proteins with an estimated weight of 150-180 kDa that are anchored in the outer cell membranes. INPs are typically present as monomers but have repeatedly been shown to aggregate in the bacterial outer membranes ¹⁴⁻¹⁷. The ice nucleation induced by bacteria generally occurs in the ranges of -2 °C to -4 °C, -5 °C to -7 °C and below -7 °C and is associated with INP subpopulations of different sizes with the monomers being the least and the multimer aggregates being the most efficient ¹⁸⁻¹⁹. Based on their activity in droplet freezing experiments, the bacterial INP aggregates are usually grouped into classes A, B, and C ¹⁹. Class A represents large protein oligomers of up to a hundred INPs ¹⁴ and class C consists of a few single proteins ^{14, 20}. However, up to now, the details of the structure and functionality of the INP aggregates remain the object of active research. On the molecular scale, the INPs are believed to function by organizing water into ice-like patterns, which increase in size as the temperature decreases until they are large enough to form a stable embryonic crystal, which leads to ice growth ²¹⁻²². Potent biological ice nucleators have also been identified in freeze-tolerant organisms, i.e., insects that survive the freezing of a fraction of their body fluids, pollen, and fungi, as integral parts of their freeze-tolerance strategy in nature ^{11, 23-26}.

Freeze-avoiding organisms have evolved an opposite approach to ensure survival in subzero environments. They produce antifreeze proteins (AFPs) or antifreeze glycoproteins (AFGPs) that are able to bind to embryonic ice crystals and arrest their macroscopic growth ^{11, 27-28}. The success of AF(G)Ps as efficient protection against freezing can be witnessed by their wide distribution among organisms of different kingdoms, phyla, and species ¹¹. The AF(G)Ps found

in different organisms show a remarkable diversity in structures²⁹, but share the same capability of binding to ice and the lowering of the temperatures of ice growth²⁸. Despite having diametrically opposite functions, the INPs resemble AF(G)Ps in many respects. Both classes of proteins can interact with ice, and molecular simulations suggest that the active ice-binding surfaces and adjacent interfacial water of INPs are similar to AF(G)Ps, just on a much larger scale^{20,30}. In fact, there is increasing experimental evidence that AF(G)Ps have ice-nucleating activity³¹⁻³³, but at temperatures much lower than those of INPs.

Interestingly, AF(G)Ps have frequently been reported to also inhibit the activity of biological ice nucleators³⁴⁻³⁵. Such an effect would be of physiological importance and could explain why some cold-adapted organisms produce both AFPs and INPs^{11,23}. However, the mechanism of the inhibition of INPs by AFPs remains largely unknown and is highly debated³⁴⁻³⁶. For instance, Parody-Morreale *et al.* reported that antifreeze glycoproteins (AFGPs) from *Dissostichus mawsoni* inhibited the ice-nucleating activity of the bacteria *E. herbicola* and *P. syringae*³⁷, whereas Holt *et al.* found that in the presence of AFGPs the ice-nucleating activity of *P. syringae* was slightly enhanced³⁶. Olsen and Duman reported that AFPs from the beetle *Dendroides canadensis* inhibited the activity of the bacteria *Pseudomonas fluorescens*, but not the INPs from the crane fly *Tipula trivittata*³⁸. While the above studies clearly demonstrate the importance of the subject, the reported experiments focused on a limited number of antifreeze proteins and left many open questions. Here, we investigate the effect of all major classes of AFPs and the non-AFP bovine serum albumin (BSA) on bacterial ice nucleators using the recently developed high-throughput twin-plate ice nucleation assay (TINA)³⁹. TINA enables the simultaneous measurement of a complete dilution series with high statistics, enabling the analysis and characterization of the efficiency of biological ice nuclei and their inhibitors with high accuracy³⁹.

Figure 1 shows the results of TINA measurements of a dilution series of bacterial ice nucleators from *P. syringae* (Snomax®). Snomax® is widely used as a model and reference system for biological and atmospheric ice nucleation studies³⁹⁻⁴¹ and contains a preparation of freeze-dried irradiated bacteria cells of *P. syringae*. The initial mass concentration was 0.1 mg/mL and was then serially diluted 10-fold, over six orders of magnitude, with pure water using an automated liquid handling station. The resulting ice nucleator concentration ranged from 0.1 mg/mL to 0.1 ng/mL, per 3 µL droplet. The cumulative ice nucleator number concentration (N_m) was calculated using Vali's formula and represents the total number of ice nucleators that are active above a certain temperature⁴².

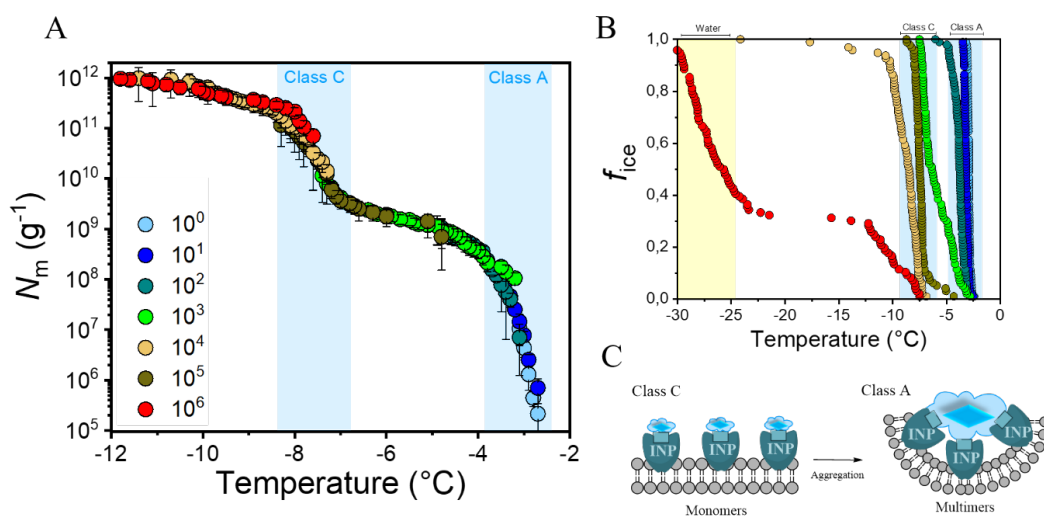


Figure 1. Freezing experiments of aqueous solutions of Snomax[®] containing bacterial ice nucleators from *P. syringae*. **A)** Cumulative number of ice nucleators (N_m) per unit mass of *P. syringae* vs. temperature for various dilutions indicated in the legend, starting with 0.1 mg/mL. The error bars were calculated using the counting error and the Gaussian error propagation. **B)** Fraction of frozen droplets (f_{ice}) for different *P. syringae* dilutions. Symbol colors indicate data from droplets with different concentrations and are identical to A. The temperature ranges for class A and C nucleators according to the definition by Turner *et al.* are shaded in light blue¹⁹. The yellow shaded region presents the temperature range in which pure water freezes in our system³⁹. **C)** Schematic structure of class C and A nucleators. Class C consists of monomeric INPs, which aggregate to form the highly-efficient class A. The INPs are shown with their active ice-binding site and preordered water patterns in their vicinity.

For the pure bacterial ice nucleator solution, the spectrum shows two strong increases in $N_m(T)$ around ~ -2.9 °C and ~ -7.5 °C with distinct plateaus between ~ -4.5 °C and ~ -7 °C and above ~ -9.5 °C (Figure 1). These values are reproducible (Supporting Figure 1), and consistent with previous studies³⁹⁻⁴¹. The two distinct rises in the spectrum indicate that the ice nucleation activity of *P. syringae* stems from two distinct subpopulations of ice nucleators with different activation temperatures. These subpopulations have previously been assigned to isolated (~ -7.5 °C) and aggregated (~ -2.5 °C) INPs¹⁹. Aggregation of INPs occurs in the cell membrane of *P. syringae* under conditions of stress as shown in Figure 1C¹³. The aggregation of the INPs effectively increases the size of the ice-binding surface, which increases the ice nucleation activity through cooperative effects⁴³. The plateaus at temperatures below the two increases of $N_m(T)$ indicate that there are no or few additional ice nucleators active at different temperatures in the investigated droplets⁴¹. The presence of two predominant ice nucleators becomes apparent when comparing the droplet freezing statistics of the different dilutions as shown in

Figure 1B. The two main bacterial ice nucleators with activities around ~ -2.5 °C and ~ -7.5 °C are apparent, as well as a third rise at ~ -25 °C. We assign the observed rise at ~ -2.5 °C to the highly efficient nucleators of class A and the less-efficient ice nucleators that cause nucleation around ~ -7.5 °C to class C¹⁹. Protein complexes of class B were not clearly observed in our measurements which is in agreement with others^{39,41}. The third rise at ~ -25 °C corresponds to the freezing point of pure water in our system³⁹. At high Snomax® concentrations, intermolecular interactions between INPs and different bacterial fragments of Snomax® (*P. syringae*) occur, which results in aggregation and the formation of class A. Diluting the samples reduces the probability of such interactions and hence the formation of larger aggregates (Supporting Figure 2). At very high dilution, there are no more ice nucleators present and eventually, the curve resembles that of pure water.

Figure 2 shows the effect of BSA, the fish antifreeze proteins type 1 (AFP-1), type 3 (AFP-3), the insect antifreeze protein from *Rhagium mordax* (RmAFP), and the small antifreeze glycoproteins (AFGP₇₋₈) and larger AFGP₁₋₅ isoforms on the ice-nucleating activity of *P. syringae*. The investigated proteins vary significantly in terms of three-dimensional structures, putative ice-binding planes, and antifreeze activities²⁹. The freezing assay experiments were performed at fixed AF(G)P concentration of 0.1 mg/mL, and the same dilution range for INP as that shown in Figure 1, which results in varying AF(G)P: INP ratios spanning from 1:1 (weight %) to a vast excess of AF(G)Ps.

The most efficient bacterial class A ice nucleators are predominately present in concentrated Snomax® (*P. syringae*) solutions. The addition of different proteins to the concentrated *P. syringae* solutions gives rise to three types of effect: (i) the addition of AFGP₁₋₅ has no effect on the freezing curve, and the freezing statistics are indistinguishable from that of Snomax® (*P. syringae*) in water. AFGP₁₋₅ is known to stabilize membranes⁴⁴⁻⁴⁵, so that the aggregation of proteins in membrane structures (Figure 1B) is preserved and IN activity retained; (ii) AFP-3, AFGP₇₋₈, RmAFP, and BSA have similar and rather small effects on the class A ice nucleators: The resulting response is comparable to that of the pure bacteria, only shifted by ~ 0.7 °C to lower temperatures (Figure 2A). This minor inhibition of the class A nucleators is statistically insignificant, but it should be noted that the f_{ice} at 0.5 was found to always be at lower temperatures compared to Snomax®. This could indicate a slight inhibition that originates from non-specific interactions of the different proteins with the outer cell membrane of *P. syringae*. Such interactions would perturb the formation of the precisely aligned INP aggregates that give rise to the high freezing temperatures, and, as a consequence, freezing occurs at lower

temperatures. This inhibition mechanism is independent of antifreeze activities, explaining why AFPs and non-AFPs show similar activities. (iii) The addition of AFP-1 shifts the freezing point by ~ 3.5 °C to lower temperatures. The resulting freezing temperature around ~ -7 °C closely resembles that of class C nucleators. AFP-1 is known to insert and disturb model cell membranes⁴⁵⁻⁴⁷ and evidently has the ability to prevent the formation of highly efficient class A aggregates, reducing the activity of class A to that of class C.

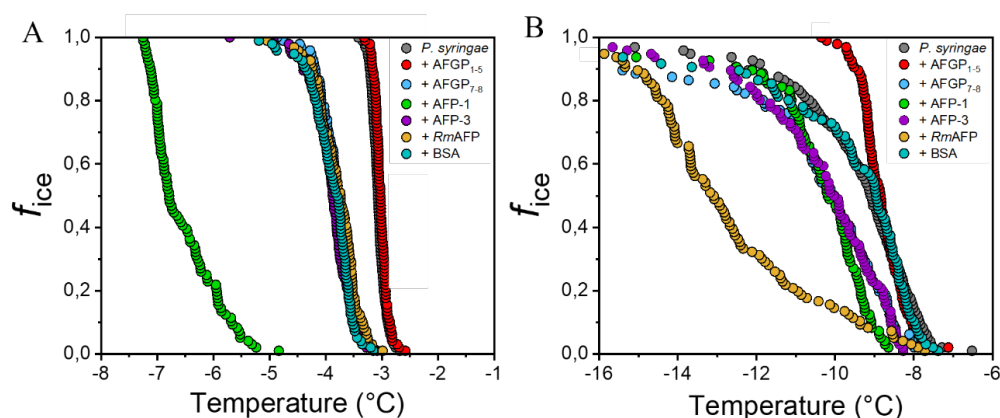


Figure 2: Freezing experiments of aqueous solutions of Snomax® containing ice nucleators from *P. syringae* in the presence of a variety of AF(G)Ps. Fraction of frozen droplets (f_{ice}) for concentrated (0.1 mg/mL, A) and diluted (1 ng/mL, B) Snomax® (*P. syringae*) solutions in the presence of different AF(G)Ps at fixed 0.1 mg/mL concentration. The high concentration Snomax® (*P. syringae*) solutions predominately contain INP aggregates (class A), while lower concentration solutions are mostly monomeric INPs (class C).

Additionally, we examined the effect of the AF(G)Ps on the class C ice nucleators of *P. syringae*, where the INPs are predominately present as smaller aggregates and monomers. We find that some AF(G)Ps inhibit ice nucleation activity and shift the freezing point to lower temperatures. Interestingly, the degree of the inhibition is markedly different from that observed for class A. For class C, *RmAFP* is the most efficient inhibitor and AFP-1, the most efficient class A inhibitor, only shows moderate counter-activity. We further find that BSA shows no effect on class C, which is in line with previous studies that observed no inhibitory effect of non-AFPs on class C nucleators³⁴.

Interestingly, we find that the larger isoform AFGP₁₋₅ shows no inhibition but a moderate enhancement of the activity of class C nucleators, while the smaller isoform AFGP₇₋₈ inhibits the INPs. The enhancement of ice nucleation by the addition of AFGP₁₋₅ is interesting, since bacterial nucleators are very potent ice nucleators on their own. We rank the inhibition

efficiency in the order $RmAFP \gg \gg AFP-1 = AFGP_{7-8} = AFP-3 = BSA > AFGP_{1-5}$. This ranking does not reflect the efficiency of the antifreeze activity of the proteins, except for $RmAFP$, which is the most active AFP known.

To further explore the relationship between the inhibition of the INPs and antifreeze activity, we studied inactive mutants of AFP-3 and AFGP₁₋₅. In the AFP-3* variant, the threonine 18 residue was replaced by asparagine, which results in the elimination of antifreeze activity²⁹. In the AFGP₁₋₅* variant, we oxidized the C-6 hydroxyls of the galactose moieties to carboxylic acids, which results in the elimination of >80 % of the antifreeze activity⁴⁸. In Figure 3, we compare the true supercooling of all the AFPs and different variants at high and low Snomax® (*P. syringae*) concentrations. Supercooling values ΔT represent the temperature difference at $f_{ice} = 0.5$ between Snomax® (*P. syringae*) with and without the added protein at high (left panel) and low (right panel) Snomax® (*P. syringae*) concentrations. All proteins were measured at least three times. For both AFP-3s, we find no significant difference between active and inactive variants on the class A or C ice nucleators. For the modified AFGP-Carboxyl variant, we likewise observe no significant change relative to the natural AFGP₁₋₅. AFP-1 and $RmAFP$ both inhibit ice nucleation, yet each with marked and opposite efficiencies for class A and class C INPs.

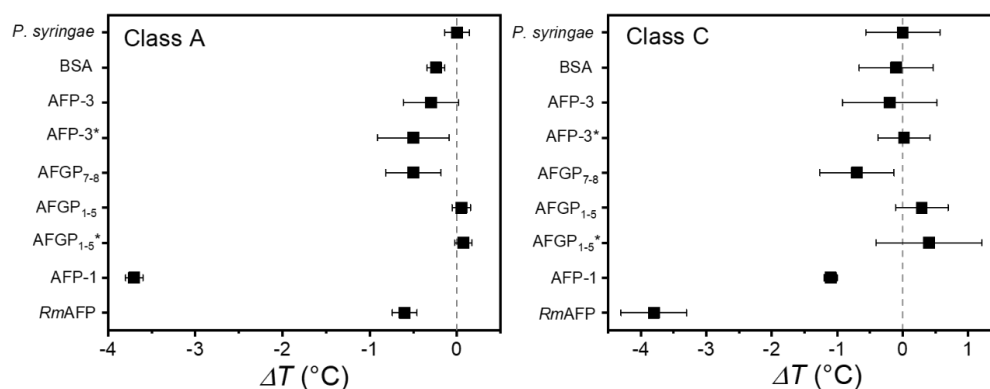


Figure 3: Ice nucleation inhibition efficiency of AF(G)Ps and non-AFP on class A (aggregates) and C (monomer) bacterial ice nucleators solutions. Shifts represent the difference of f_{ice} at 0.5 between *P. syringae* and added AF(G))P. Each experiment was performed at least three times, and the error bars represent the standard deviation between the individual measurements.

Freeze-avoiding organisms extend their supercooling abilities by the accumulation of colligative solutes, by the removal of ice nucleators, by inhibiting the activity of ice nucleators present, or alternately by inactivating embryonic ice crystals¹¹. The importance of biological macromolecules that inhibit ice nucleators has been recognized, but there have been no systematic studies with sufficiently robust analytical methods⁴⁹. AF(G)Ps were among the most studied systems due to their known properties to recognize and bind ice crystals inhibiting their growth³⁴⁻³⁵. However, the generality of such inhibition of INPs by AF(G)P is not well supported, and for the few reported examples of inhibition by AF(G)Ps, the mechanism is unknown. Our results provide clear evidence that some AF(G)Ps can inhibit the activity of bacterial INPs, while others show no effect. Bacterial INPs are known to form preordered “ice-like” interfacial water domains at their putative ice-binding sites²². Likewise, AFPs like *RmAFP* are known to use preordered “ice-like” interfacial water domains for the recognition and subsequent attachment to ice^{32, 50-53}. We speculate that the preordered water domains of the AFPs will have a high affinity for similar “ice-like” water domains and, therefore, for the preordered water of the INPs⁵⁰. Consequently, upon diffusion of AFPs into the vicinity of INPs, they will remain longer in the hydration shell of INPs and thereby disturb the interfacial water arrangement required for a nucleation event. This form of inhibition would be dependent on the similarity of the structure and hydration motifs of AFPs and INPs and on the antifreeze activity of the AFPs, which has been related to the extent of their preordered interfacial water domains⁵². Bacterial INPs from *P. syringae* and hyperactive insect AFPs like *RmAFP* are proposed to bind ice through a flat array of TxT repeats, where T is threonine and x a non-conserved amino acid³⁰. Hence, the hyperactive *RmAFP* shows strong inhibition. Non-AFPs and inactive AFP variants will also randomly diffuse through the preordered water domains of INPs, but lack the affinity for preordered interfacial water and therefore fail to mask the INPs. This hypothetical mechanism is schematically shown in Figure 4. AFGPs are unique because of their highly flexible solution structure, multiple ice-binding sites, and capability to stabilize membranes^{44, 54-55}. Our observations that the larger AFGP₁₋₅ isoforms do not inhibit bacterial INPs but rather show no effect or even a slight enhancement are in agreement with Holt *et al.*³⁶. We speculate that larger AFGPs can stabilize INPs in the membranes and potentially link them to form larger aggregates that nucleate at higher temperatures. Our results further demonstrate that the highly efficient class A nucleators are more vulnerable to inhibition. For the most active inhibitor AFP-1, we propose a mechanism in which the proteins interact with the cell membrane of the bacteria to prevent the aggregation of the INP monomers, thereby preventing the formation of the highly efficient class A nucleators. This mechanism is in line with findings that AFP-1 directly inserts

into model cell membranes, whereas AFGPs and other AFPs interact weaker and show less effects on the membrane⁴⁵⁻⁴⁷. The different mechanisms underlying the inhibition of bacterial INPs by AF(G)Ps are likely highly specific. Thus, we conclude that no universal INP inhibition mechanism through AFPs exists. This is further supported by the findings that AFP-1, the best inhibitor of bacterial INPs, had no effect on fungal INPs (Supporting Figure 3). Ice nucleation can be promoted by a range of properties, including charge⁵⁶, lattice matching⁵⁷, hydrophobicity⁵⁸ or morphology⁵⁹. In such cases, it is not apparent that AF(G)Ps will show inhibitory activity.

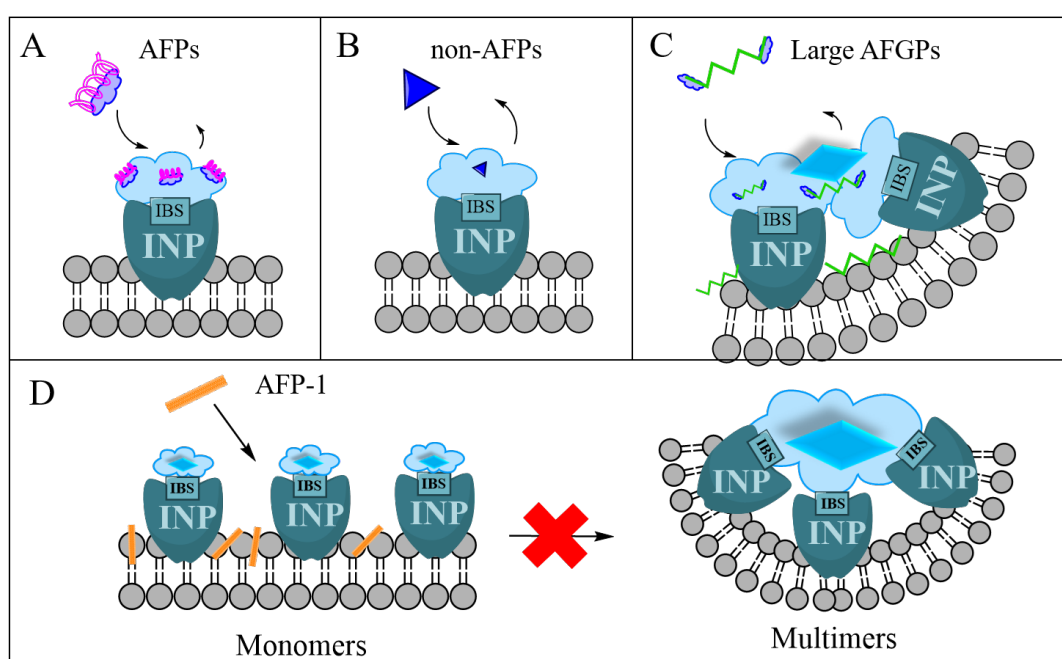


Figure 4: Proposed mechanism of the inhibition of ice nucleators from *P. syringae* by AF(G)Ps. INPs are shown with their active ice-binding sites (IBS) and preordered ice-like water patterns in their vicinity (blue clouds). **A)** Hyperactive AFPs like *RmAFP* use preordered water domains to bind to ice. They also have structural similarity to the INPs of *P. syringae*³⁰. Hence, preordered water domains of *RmAFP* will have an affinity for the preordered interfacial water domains of the bacterial INPs. As a result, *RmAFP* will disturb the preordering of the interfacial water that is needed for ice nucleation and masks the bacterial ice-nucleating sites. **B)** Non-AFPs have no preordered water domains (no blue cloud) and therefore no affinity for the interfacial water domains of INPs. **C)** Large AFGP isoforms can stabilize and link INPs to form larger aggregates that nucleate at higher temperatures. **D)** AFP-1 interacts with the cell membranes of the bacteria to prevent the functional aggregation of the INP monomers, which then give rise to the highly efficient class A nucleators.

Associated Content

Experimental details of the setups and the measurement procedure, Supporting Figures 1-3 (Ice nucleation measurements of Snomax®; Dynamic Light Scattering measurements; Ice nucleation measurements of fungal ice nucleators.)

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Competing interests

The authors declare no competing financial interests.

Acknowledgement

We thank L. E. Hanson for providing *Fusarium acuminatum* and N. Bothen, N.-M. Kropf and T. Goodwill for technical support. We are grateful to the MaxWater initiative from the Max Planck Society for financial support.

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Supplementary Material

To Freeze or Not to Freeze: Inhibition of Bacterial Ice Nucleators by Antifreeze Proteins

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submitted to the Journal of Physical Chemistry Letters (2020)

Supporting Information:

To Freeze or Not to Freeze: Inhibition of Bacterial Ice Nucleators by Antifreeze Proteins

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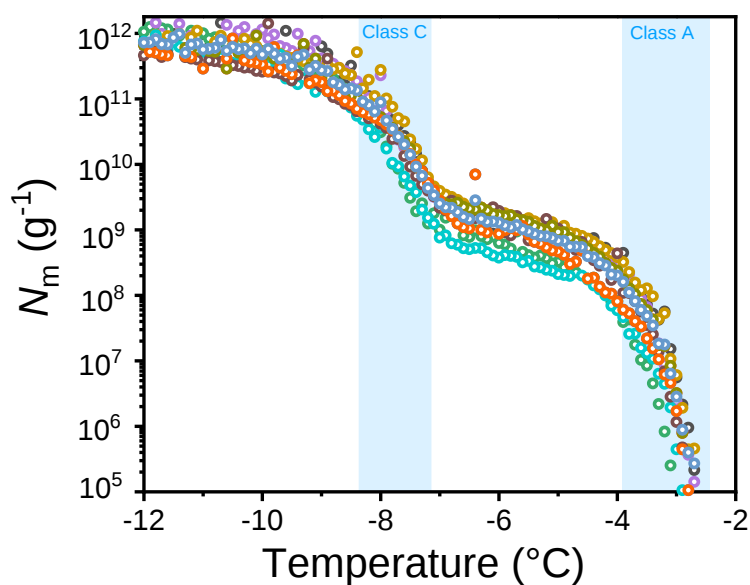
Experimental Methods

Protein Samples. Snomax[®] was obtained from SMI Snow Makers AG (Thun, Switzerland) and contains a preparation of fragmented freeze-dried, irradiated bacteria cells of *P. syringae*. AFP-1 was purified from winter flounder *Pseudopleuronectes americanus*¹. AFGP₁₋₅ and AFGP₇₋₈ were purified from the Antarctic toothfish *Dissostichus mawsoni*². RmAFP was obtained by recombinant protein expression as described elsewhere³. We investigated AFP-3 purified from the Antarctic eelpout (*Lycodichthys dearborni*) and obtained by recombinant protein expression, including the mutant T18N⁴. In this mutant, the threonine residue at position 18 is replaced by asparagine which causes the complete loss of antifreeze activity⁴. For the AFGP₁₋₅ variant, we oxidized the C-6 hydroxyl groups of the galactose moieties to carboxylic acids as described previously⁵. Only AF(G)P samples with proven antifreeze activity were used, and the activity was determined prior to the experiments using nanoliter cryoscopy. BSA was obtained from Sigma Aldrich and was used without further purification. Pure water was prepared as described elsewhere⁶.

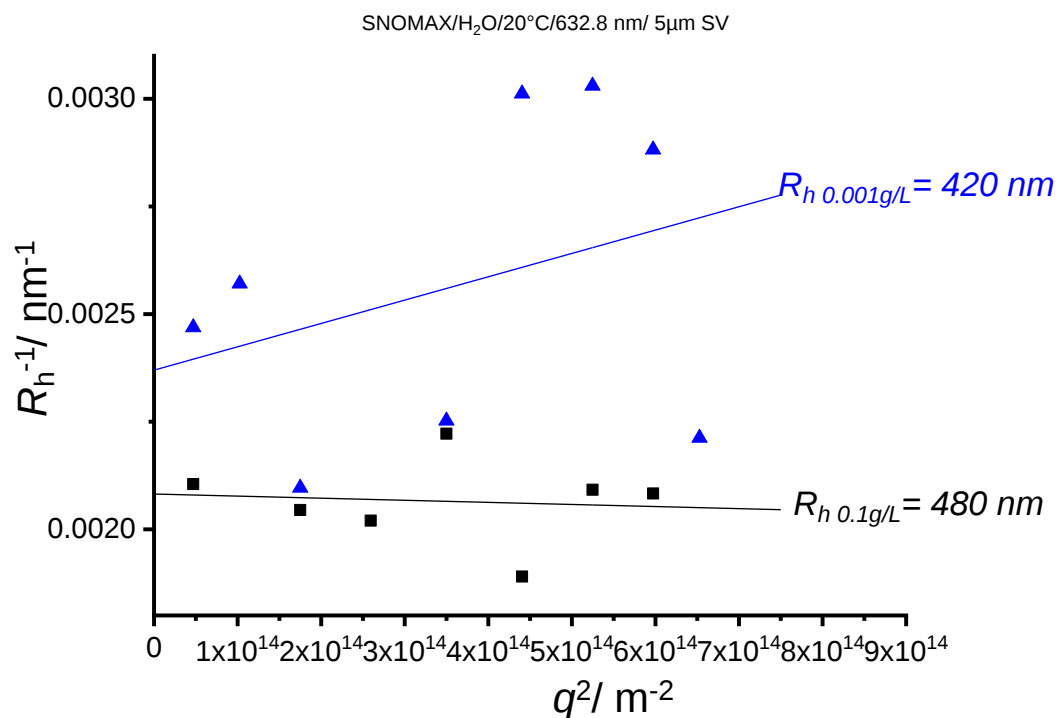
Fusarium acuminatum from USDA-ARS, Michigan State University (Linda E. Hanson, East Lansing, MI, USA) was cultivated on full-strength potato dextrose agar (VWR International GmbH, Darmstadt, Germany) first at room temperature for one week and then at 6 °C for about 4 weeks. Aqueous extract containing proteinaceous IN from spores and mycelial surfaces was prepared as described elsewhere⁷.

TINA Experiments. Ice nucleation experiments were performed using a high-throughput droplet freezing assay. The details of the instrument have been described previously⁶. In a typical experiment, a 0.1 mg/mL solution of Snomax[®] (*P. syringae*) in pure water was serially diluted ten-fold by a liquid handling station (epMotion ep5073, Eppendorf, Hamburg, Germany) with either a solution of AF(G)Ps (0.1 mg/mL) or pure water. For each dilution, 96 droplets (3 µL) per dilution were placed on two 384-well-plates and tested with a continuous cooling-rate of 1 °C/min from 0 °C to -30 °C. The freezing of droplets was determined by two infrared cameras (Seek Therman Compact XR, Seek Thermal Inc., Santa Barbara, CA, USA). The uncertainty in the temperature of the setup was ±0.2 °C. The obtained fraction of frozen droplets and the counting error were used to calculate the cumulative number of ice nucleators with the associated error using the Vali formula and the Gaussian error propagation^{6,8}. The pH values of all samples were controlled before TINA measurements and were ~6-7. Samples were measured at least three times in independent experiments except for the variants which were measured twice.

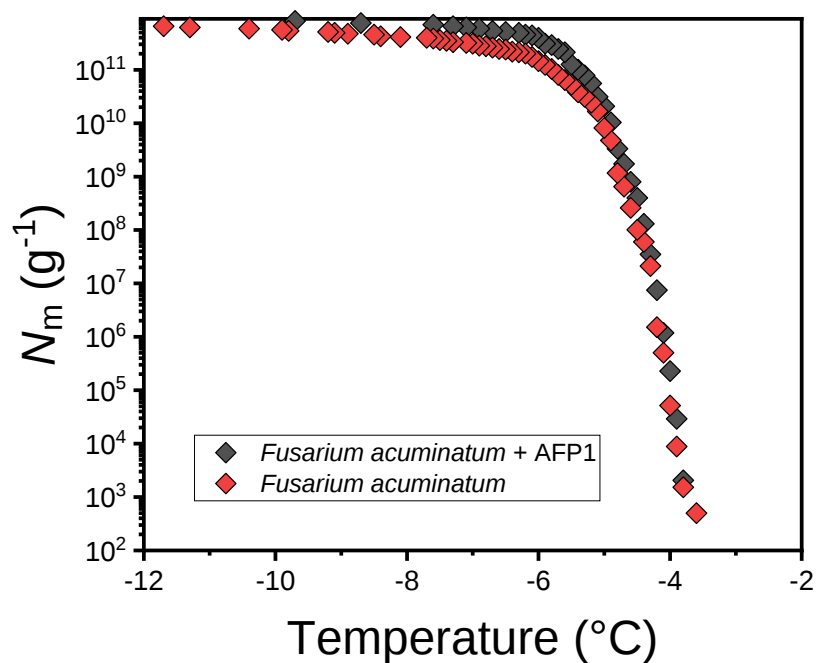
DLS measurements. The hydrodynamic radii (R_h) of 0.1 g/L and 0.001 g/L Snomax[®] in ultrapure water were determined using dynamic light scattering (DLS). Light scattering measurements were performed on an ALV spectrometer consisting of a goniometer and an ALV-5004 multiple-tau full-digital correlator (320 channels), which allows measurements over an angular range from 30° to 150°. A He-Ne laser (wavelength of 632.8 nm) was used as light source. Measurements were performed at 20 °C at 9 angles ranging from 30° to 150°.



Supporting Figure 1: Freezing experiments of aqueous solutions of Snomax[®] containing bacterial ice nucleators from *P. syringae*. Shown are the cumulative number of ice nucleators per unit mass of *P. syringae* vs. temperature for eleven independent experiments. The temperature ranges for class A and class C INPs according to the definition by Turner *et al.* are shaded in light blue⁹.



Supporting Figure 2: Hydrodynamic radii of Snomax[®] containing bacterial ice nucleators from *P. syringae* at 0.1 g/L and at 0.001 g/L. The radii at higher concentration are larger, which suggests that larger aggregates are present in the solution.



Supporting Figure 3: Freezing experiments of aqueous extracts from *Fusarium acuminatum* containing fungal ice nucleating proteins and in the presence of AFP type 1. Shown is the cumulative number of ice nucleators (N_m) per gram of mycelium plotted against the temperature (T).

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C.4 Lucas *et al.*, J. Am. Chem. Soc., 2019

Electrostatic Interactions Control the Functionality of Bacterial Ice Nucleators

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submitted to the Journal of the American Chemical Society (2019)

Contribution

Aided in freezing experiments, data analysis, and writing of the manuscript.

Title: Electrostatic Interactions Control the Functionality of Bacterial Ice Nucleators

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Abstract:

Bacterial ice-nucleating proteins (INPs) promote heterogeneous ice nucleation more efficiently than any other known material. The details of their working mechanism remain elusive, but their high activity has been shown to involve the formation of functional INP aggregates. Here we reveal the importance of electrostatic interactions for the activity of INPs from the bacterium *Pseudomonas syringae* by combining a high-throughput ice nucleation assay with surface-specific sum-frequency generation spectroscopy. We determined the charge state of non-viable *P. syringae* as a function of pH by monitoring the degree of alignment of the interfacial water molecules and the corresponding ice nucleation activity. The net charge correlates with the ice nucleation activity of the INP aggregates, which is minimal at the isoelectric point. In contrast, the activity of INP monomers is less affected by pH changes. We conclude that electrostatic interactions play an essential role in the formation of the highly-efficient functionally aligned INP aggregates, providing a mechanism for promoting aggregation under conditions of stress that prompt the bacteria to nucleate ice.

Ice formation is the most important liquid-to-solid phase transition on earth and is strongly affected by the presence of nucleators that initiate heterogeneous ice nucleation at temperatures above $-40\text{ }^{\circ}\text{C}$. There is a large variety of compounds of biological and abiotic origin that can act as ice nucleators, and their efficiency strongly differs¹⁻³. The most efficient ice nucleators are bacteria from *Pseudomonas syringae*, which can initiate the crystallization of water at temperatures as high as $-2\text{ }^{\circ}\text{C}$ ⁴⁻⁵. The ability of bacteria to nucleate ice is caused by specialized ice-nucleating proteins (INPs) that are exposed on the bacterial surface and are anchored in the outer membrane on the bacterial cell wall⁶. Bacterial INPs contain a large central-repeat domain that has been proposed to be the active site and which is responsible for ice nucleation through a mechanism that likely involves the preordering of water molecules^{7,8}. Apart from the specific ice-binding site, the high ice nucleation activity of INPs has been shown to crucially depend on the ability to aggregate into larger functional protein clusters⁹⁻¹¹. Based on their activity in droplet freezing experiments, the bacterial INP aggregates are typically divided into classes as shown in Figure 1¹⁰. Class A consists of large aggregates (>50 INPs) that are responsible for freezing at high subzero temperatures (-2 to $-4\text{ }^{\circ}\text{C}$)¹⁰. Class B consists of smaller aggregates that induce freezing at $-5\text{ }^{\circ}\text{C}$ to $-6.5\text{ }^{\circ}\text{C}$, and class C consists of mostly monomeric INPs that induce ice formation between $-7\text{ }^{\circ}\text{C}$ to $-12\text{ }^{\circ}\text{C}$ ¹⁰. In nature, the aggregation of INPs occurs under conditions of stress, which require the bacteria to nucleate ice⁵. The INP aggregation mechanism and whether INP aggregation in cell membranes is promoted by a change in chemistry is unknown. Notably, lowering the pH to acidic values has repeatedly been shown to reduce the ice nucleation activity of bacterial INPs^{10, 12-13}. In contrast, changing the pH to alkaline values did not affect the nucleation activity¹⁰. The molecular origin for this pH sensitivity is unknown, but a better understanding would provide needed insights into the driving forces of INP aggregation. Moreover, it would have direct implications for understanding biological ice nucleation in the atmosphere, where the pH levels are oftentimes acidic due to anthropogenic activities¹².

Figure 1 shows the results of ice nucleation measurements of the bacterial ice nucleator Snomax[®] at three pH values. Snomax[®] is a commonly used model system for biological and atmospheric ice nucleation studies¹⁴⁻¹⁶, and it consists of a preparation of inactivated bacteria cells of *P. syringae*. The initial Snomax[®] solutions in water had a concentration of 0.1 mg/mL and a pH value of ~ 6.2 . The samples were then serially diluted, resulting in concentrations from 1 ng/mL to 1 mg/mL . The cumulative ice nucleator number concentration (N_m) was

calculated using Vali's formula, and it represents the number of ice nucleators per unit weight that are active above a certain temperature¹⁷.

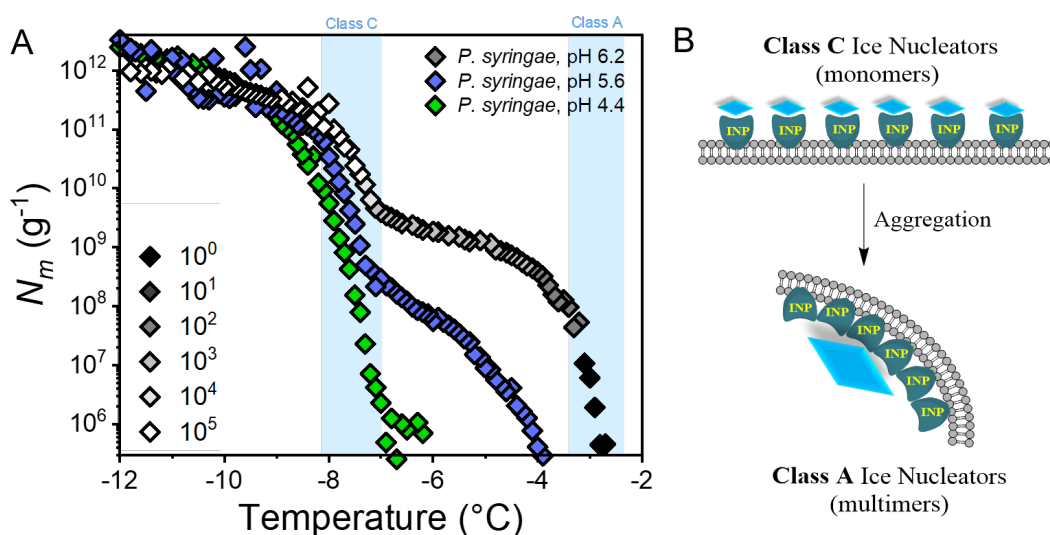


Figure 1: Freezing experiments of aqueous solutions of Snomax[®] containing bacterial ice nucleators from *P. syringae* in water and at different pH values. A) Shown is the cumulative number of ice nucleators (N_m) per unit mass of Snomax[®] vs. temperature. Numbers in the legend denote dilution factors. The temperature ranges for class A and C bacterial ice nucleators are shaded in blue¹⁰. B) Schematic structure of class A and C nucleators. Class C refers to monomeric INPs, which can aggregate to form the highly efficient class A nucleators.

For the bacterial ice nucleator solution in water (pH $\sim$ 6.2), the spectrum shows two strong increases in $N_m(T)$ around $\sim$ -2.9 °C and $\sim$ -7.5 °C with plateaus between $\sim$ -4.5 °C and $\sim$ -7 °C and above $\sim$ -9.5 °C. The two rises in the spectrum reveal that the ice nucleation activity of *P. syringae* stems from two classes of ice nucleators with different activation temperatures. The plateaus at temperatures T below each increase of $N_m(T)$ arise when fewer ice nucleators at these temperatures are present¹⁵. We attribute the observed rises at $\sim$ -2.9 °C and $\sim$ -7.5 °C to class A and C ice nucleators, respectively.

For lower pH solutions, the trend looks markedly different. At pH 5.5 the rise at $\sim$ -2.9 °C is absent; instead, we observe a rise at $\sim$ -4.5 °C. Further, the second rise at $\sim$ -7.5 °C is slightly shifted by $\sim$ -0.5 °C. Evidently, lowering the pH influences the ability of the ice-nucleating proteins to form the more efficient class A aggregates. Further lowering the pH fortifies this effect and at a pH of $\sim$ 4.4, the class C nucleators have disappeared; apparently, class A

nucleators were converted into class C. While class A nucleators are very pH-sensitive, class C nucleators are only weakly affected by pH variations.

To obtain a more complete picture of the effect of the pH on the ice nucleation activity of the class A bacterial INPs we conducted a comprehensive evaluation of pH values from 2–10.5 as shown in Figure 2. We find that the change of pH gives rise to different types of effects as shown in Figure 2B: (i) Lowering the pH to acidic values shifts the freezing point of class A to lower temperatures. This trend increases until pH ~ 4 , where the resulting freezing temperature $\sim -7^\circ\text{C}$ closely resembles that of class C nucleators. The more acidic conditions clearly prevent the formation of highly efficient class A aggregates, reducing the activity of class A to that of class C. (ii) Raising the pH to basic conditions has a small effect on the class A ice nucleators and the resulting response looks similar to that of bacteria in water, only shifted by $\sim 0.6^\circ\text{C}$ to lower temperatures (Figure S1, Figure 2B). (iii) Raising the pH to extreme basic conditions shifts the freezing point by $\sim 3.5^\circ\text{C}$ to lower temperatures. The resulting freezing temperature $\sim -7^\circ\text{C}$ resembles that of class C nucleators.

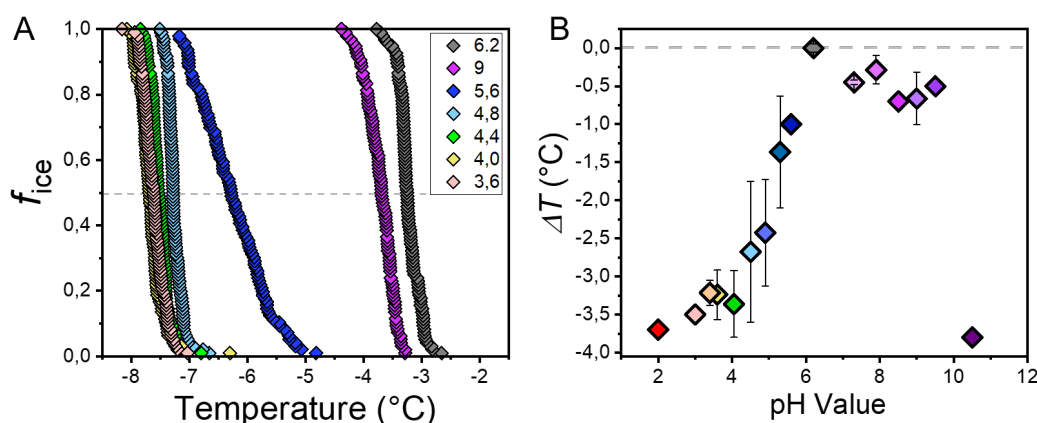


Figure 2: Freezing experiments of aqueous solutions of Snomax® containing bacterial ice nucleators from *P. syringae* at different pH values. A) Fraction of frozen droplets (f_{ice}) for highly concentrated (0.1 mg/mL) Snomax® solutions. B) Temperature shifts ΔT induced by different pH values. Shifts represent the difference at $f_{ice} = 0.5$ between Snomax® in water (pH 6.2) and at different pH values. Error bars represent the standard deviation for multiple independent measurements.

To investigate the molecular origin of the strong pH dependence of bacterial INP's ice nucleation efficiency, we conducted sum-frequency generation (SFG) spectroscopic experiments of Snomax® adsorbed to the air-water interface at different pH values. SFG is a

surface-specific method that can be used to probe the interfacial water of biomolecules¹⁸⁻¹⁹. In this technique, an infrared light pulse and a visible pulse are combined at a surface to generate light at the sum-frequency of the two incident fields. The technique is bulk-forbidden in isotropic media, and only ensembles of molecules with a net orientation, e.g. at an interface, can generate a detectable signal. The SFG signal intensity depends on the number of aligned molecules at the interface.

At charged surfaces, the surface field can align the water dipoles. Such charge-induced enhanced ordering of the interfacial water molecules causes the signal intensity in the O-H stretching region ($\sim 3150\text{--}3600\text{ cm}^{-1}$) to increase, and, inversely, the SFG signal intensity can be used to quantify the amount of charge at the electrified surface. This concept has previously been applied to determine the isoelectric point (IEP) of proteins²⁰⁻²³. Figure 3A shows pH-dependent SFG spectra of aqueous solutions of 0.1 mg/mL Snomax[®] adsorbed to the air-water interface. In the frequency region from $2800\text{--}3100\text{ cm}^{-1}$, the SFG spectra show strong signals from C-H stretching vibrations. We assign these bands to the methyl symmetric stretch (2880 cm^{-1}), a Fermi resonance (2940 cm^{-1}), and aromatic C-H (3050 cm^{-1}) vibrations. At frequencies above 3100 cm^{-1} , the spectrum shows a broad response from the O-H stretching band of interfacial water molecules. The SFG intensity of the CH and OH groups shows substantial changes when the solution pH is altered. We find that at pH values ~ 4.2 the intensity of the O-H stretching signals is close to zero, whereas, at values below and above, the intensity of the O-H bands increases markedly and dominates the SFG spectrum of *P. syringae*. The observed changes in the C-H region can be explained with interferences with the O-H resonances²¹⁻²². To more accurately determine the IEP of *P. syringae*, we integrated the SFG signal in the frequency region of $\sim 3100\text{--}3600\text{ cm}^{-1}$. Figure 3B shows the pH dependence of the integrated intensity of the O-H stretching region. The lowest value for *P. syringae* was found at pH ~ 4.2 , which corresponds to the point of no net charge or IEP of the bacteria. This value is similar to the bulk IEP of 4.0 that was previously reported for *P. syringae*²⁴.

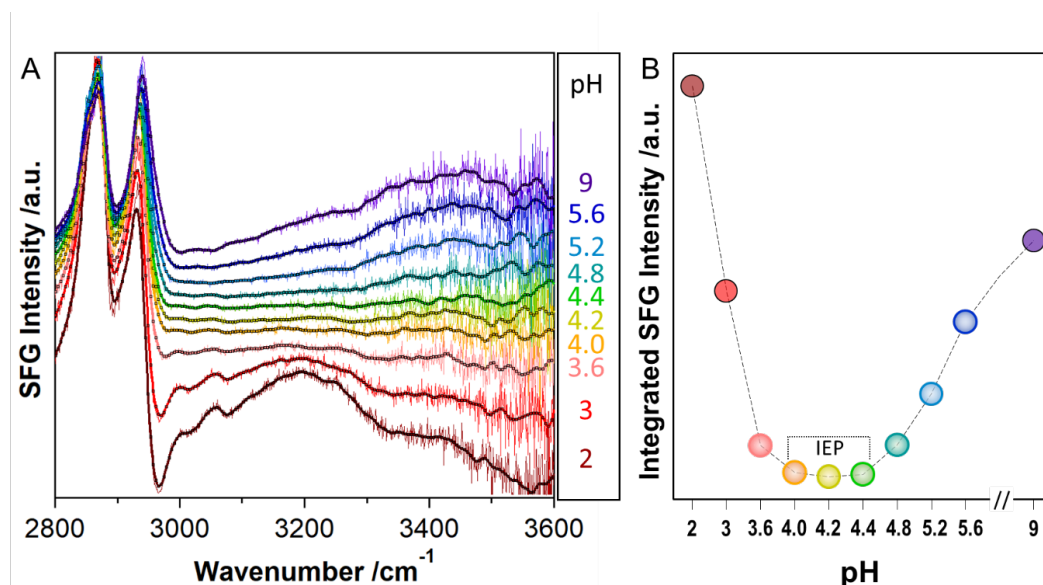


Figure 3: A) SFG spectra of aqueous solutions of Snomax[®] containing bacterial ice nucleators from *P. syringae* at the air-water interface as a function of bulk pH. The bulk concentration of Snomax[®] was 0.1 mg/mL. B) Integrated SFG intensity of the frequency region from 3100–3600 cm⁻¹ for Snomax[®] (*P. syringae*) at different bulk pH values.

Ice nucleation activities of bacteria are being studied extensively due to their important roles in precipitation and frost injury of plants²⁵. Elucidating the impact of environmental factors such as the pH is essential for understanding not only atmospheric ice nucleation processes, but also the functionality of INP aggregates. Electrostatic properties are governed by the distribution and ratio of charged and polar residues within protein structures and are among the most important factors that determine the functionality, stability, and interactions of proteins. We demonstrate that acidic pH values reduce the ice-nucleating activity of non-viable *P. syringae* bacteria and that the effect arises from the inactivation of the highly efficient class A aggregates active at high subzero temperatures. In contrast, we find no significant influence of basic pH values on the ice nucleation activity. The effect of acidic pH on the ice nucleation activity of *P. syringae* has previously been observed and those results are consistent with our findings^{4, 10, 12-13}. Turner *et al.* proposed that the acidic pH denatures the larger class A INP complexes and that this process is irreversible¹⁰. Unlike Turner *et al.*, we observe that the ice nucleation activity of the class A INP aggregates can be partially recovered upon raising the pH back to neutral values (Figure S2-S4). Combined, the pH-dependent SFG and droplet freezing experiments revealed that eliminating the net negative charge of *P. syringae* correlates with the decrease of the ice nucleation activity from the large class A INP aggregates. We explain these

observations with the formation of misfolded INP aggregates as a result of the decreased charge repulsion. Upon lowering the pH towards pH 4.2, negatively charged amino acids are protonated, and the net charge is reduced. The large numbers of glutamic acid residues contained in INPs of *P. syringae* are likely candidates for protonation⁸. As a result of the protonation, the charge repulsion between INP monomers is absent, and hydrophobic interactions cause the INPs to misfold. These newly formed INP aggregates differ substantially from the precisely aligned functional class A INP aggregates and lack the ice nucleation activity at high subzero temperatures. Consequently, only single INPs or smaller aggregates, i.e. class C nucleators remain active at acidic pH, which is in line with our experimental observations. The finding that the ice nucleation activity can be recovered by going back to neutral pH, further provides evidence that no irreversible denaturation of INPs occurs. In nature, the aggregation of INPs occurs in the cell membrane of ice-nucleating bacteria under conditions of stress that require them to nucleate ice⁵. The alteration of the pH in the system would provide a means to trigger INP aggregation in the cell membrane.

Associated Content

Experimental setups, experimental methods, effect of alkaline pH on ice nucleation activity, effect of reversibility of acidic pH change, Supporting Figures S1-S4

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Notes

The authors declare no competing financial interests.

Acknowledgment

We are grateful to the MaxWater initiative from the Max Planck Society and the Max Planck Graduate Center with the Johannes Gutenberg-Universität Mainz for financial support.

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Supplementary Material

Electrostatic Interactions Control the Functionality of Bacterial Ice Nucleators

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submitted to the Journal of the American Chemical Society (2019)

Supporting Information:

Electrostatic Interactions Control the Functionality of Bacterial Ice Nucleators

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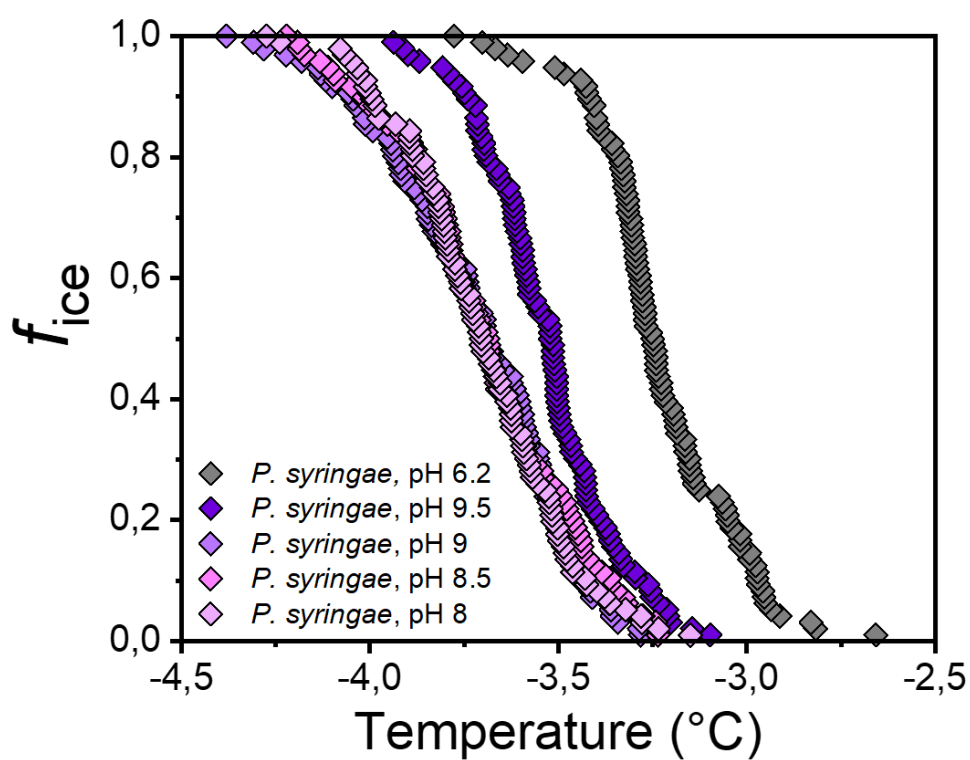


Figure S1: Freezing experiments of aqueous solutions of Snomax[®] containing bacterial ice nucleators from *P. syringae* at alkaline pH values. Fraction of frozen droplets (f_{ice}) for highly concentrated (0.1 mg/mL samples) Snomax[®] (*P. syringae*) solutions.

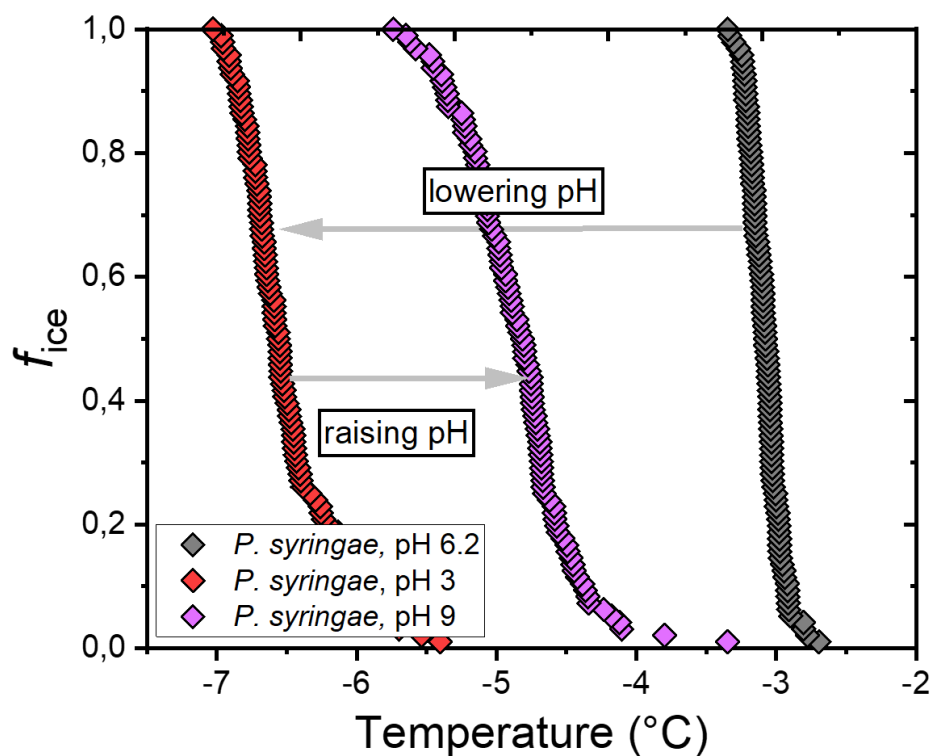


Figure S2: Freezing experiments of aqueous solutions of Snomax[®] containing bacterial ice nucleators from *P. syringae* as a function of pH values. Fraction of frozen droplets (f_{ice}) for highly concentrated (0.1 mg/mL) Snomax[®] (*P. syringae*) solutions, in which the pH was first decreased using HCl and then increased using NaOH.

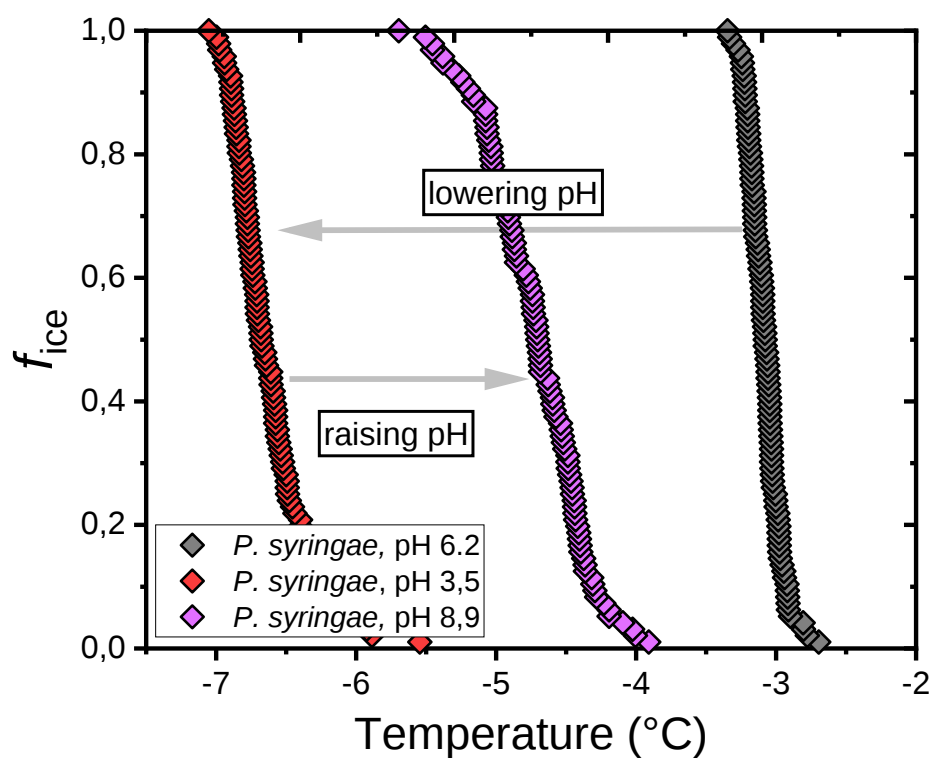


Figure S3: Freezing experiments of aqueous solutions of Snomax[®] containing bacterial ice nucleators from *P. syringae* as a function of pH values. Fraction of frozen droplets (f_{ice}) for highly concentrated (0.1 mg/mL) Snomax[®] (*P. syringae*) solutions, in which the pH was first decreased using HCl and then increased using NaOH.

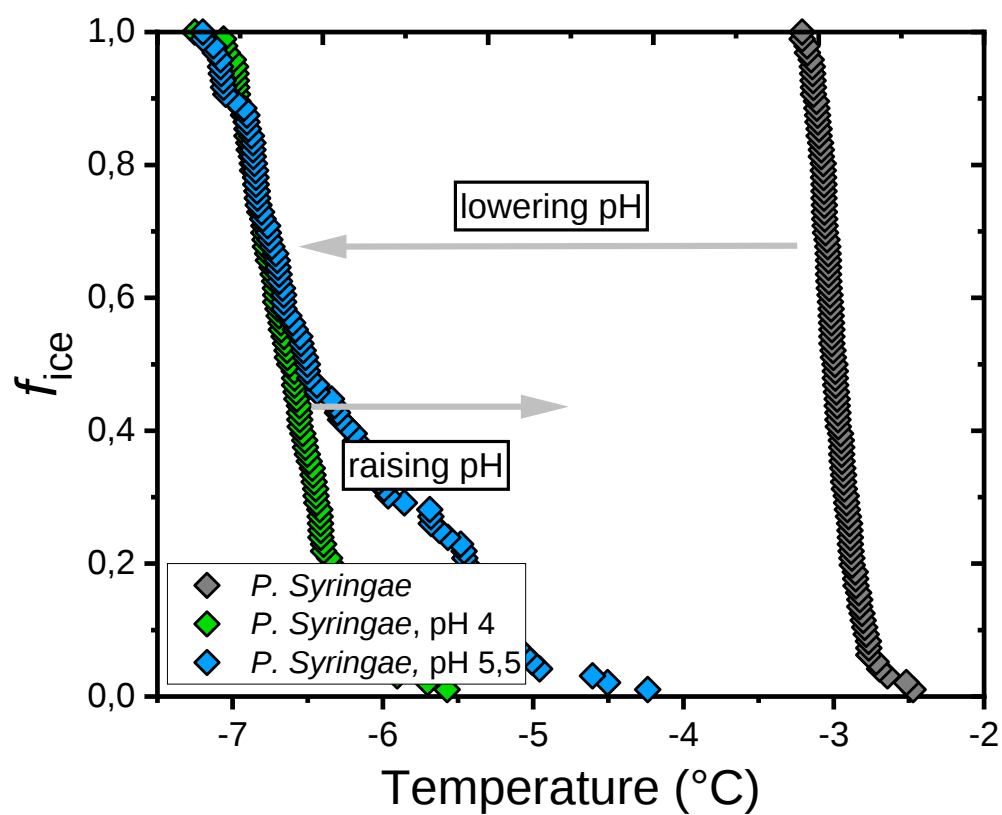


Figure S4: Freezing experiments of aqueous solutions of Snomax[®] containing bacterial ice nucleators from *P. syringae* as a function of pH values. Fraction of frozen droplets (f_{ice}) for highly concentrated (0.1 mg/mL) Snomax[®] (*P. syringae*) solutions, in which the pH was first decreased and then increased using acetate buffer.

Materials and Methods

Samples: Pure water was prepared as described elsewhere¹. Snomax[®] was obtained from SMI Snow Makers AG (Thun, Switzerland) and contains a preparation of freeze-dried, irradiated bacteria cells of *Pseudomonas syringae*. Buffer materials (Tris, PBS, acetate) as well as NaOH and HCl were obtained from Sigma. The concentration of Snomax[®] was 0.1 mg/mL, and the pH value in pure water was 6.2 +/- 0.2. The SFG experiments and TINA experiments were performed in either pure water or in 0.1 M buffer solution (acetate, PBS, Tris,) of the respected pH. The ionic strength of the solutions was 0.1 M and adjusted by adding NaCl. The pH values of all samples were controlled before each measurement.

TINA Experiments. Ice nucleation experiments were performed using a high-throughput droplet freezing assay. The details of the instrument have been described recently¹. In a typical experiment, the investigated ice nucleator sample was serially diluted 10-fold by a liquid handling station (epMotion ep5073, Eppendorf, Hamburg, Germany). 96 droplets (3 μ L) per dilution were placed on two 384-well-plates and tested with a continuous cooling-rate of 1 $^{\circ}$ C/min from 0 $^{\circ}$ C to -20 $^{\circ}$ C. The temperature of droplet-freezing was determined by two infrared cameras (Seek Therman Compact XR, Seek Thermal Inc., Santa Barbara, CA, USA). The uncertainty in the temperature of the setup was $\pm 0.2^{\circ}$ C. The obtained fraction of frozen droplets were used to calculate the cumulative number of ice nucleators using the Vali formula. Experiments were performed 3 - 6 times on independent samples.

Sum-Frequency Generation Spectroscopy Experiments:

The details of the experimental setup have been described previously². For the SFG intensity spectra, we used a conventional SFG setup in reflection geometry. A broadband IR (~ 5 μ J, full width half maximum (FWHM) of ~ 450 cm^{-1}) and a narrowband VIS (~ 13 μ J, centered at ~ 800 nm, FWHM of ~ 15 cm^{-1}) beam were focused and spatially and temporally overlapped on the sample surface, with incident angles of 36° (VIS) and 41° (IR) with respect to the surface normal. The spectra presented in this study were simultaneously recorded in the C-H and O-H-stretch region. The spectral resolution was limited by the bandwidth of the VIS. The generated SFG signal was collimated by a lens, directed and focused onto a spectrograph (Acton SP 300i, Princeton Instruments) and detected by a camera (Newton 970, Andor Instruments). All spectra were obtained in the ssp-polarization combination (s-polarized SFG, s-polarized VIS, p-polarized IR). Background spectra were taken with a blocked IR beam and

all spectra were normalized to reference spectra from z-cut quartz. The SFG measurements were performed in a custom-made Teflon trough at room temperature. The Snomax[®] (*P. syringae*) solutions were measured at a concentration of 0.1 g/L. Samples were measured in allowed to equilibrate for two hours before measurements. The equilibration of the samples at the interface was complete, when the surface pressure was constant and the SFG spectra did not change within one hour.

Colligative effects

We exclude significant contributions of colligative effects on our results, since we performed measurements in buffer systems or by adjusting the pH value using NaOH and HCl. The ionic strength was kept constant at 0.1 M for all solutions and adjusted by adding NaCl. The maximum theoretical shift due to colligative effects in the TINA measurements would be ~ 0.4 °C which is less than the observed shifts.

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C.5 Kunert *et al.*, in preparation, 2020

Reaction kinetics and products of nitration and oligomerization of the major grass pollen allergen Phl p 5 by peroxyxynitrite and nitrogen dioxide with ozone

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Contribution

KL, UP, and JFN designed the study. ATK, KRS, KZ, CSK, and CW performed the experiments. ATK, KRS, KZ, KL, UP, JF analyzed, interpreted and discussed the results. CSK, CW, and MGW contributed to the discussion. ATK and JFN wrote the manuscript.

Reaction kinetics and products of nitration and oligomerization of the major grass pollen allergen Phl p 5 by peroxyne nitrite and nitrogen dioxide with ozone

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Abstract. Air pollution is associated with the increasing burden of allergic diseases. The allergenic potential of proteins may be enhanced via posttranslational modifications induced by anthropogenic air pollutants like ozone (O₃) and nitrogen dioxide (NO₂). Furthermore, air pollution causes severe lung diseases involving inflammations, and inhaled allergens can be chemically modified by peroxyne nitrite (ONOO⁻), which is released during inflammations in the human body. The molecular mechanisms and kinetics of the chemical modifications that enhance the allergenicity of proteins, however, are still not fully understood. We studied protein tyrosine nitration and oligomerization of the major timothy grass pollen allergen Phl p 5 upon simultaneous exposure of O₃ and NO₂ and reaction with ONOO⁻. The nitration degree and oligomer formation were analyzed by reversed-phase and size-exclusion liquid chromatography. Both, exposure to O₃/NO₂ and reaction with ONOO⁻, led to nitration and oligomerization of the allergen. The nitration degree and oligomer formation were dependent on the nitrating agent and the reaction conditions (concentration, exposure time), and they were higher for ONOO⁻ than for O₃/NO₂. For the reaction with ONOO⁻, we found a maximum nitration degree of 25 %, which corresponds to maximal ~6 nitrated tyrosine residues per Phl p 5 monomer, whereas exposure to O₃/NO₂ resulted in a maximum nitration degree of 8 % corresponding to ~1 nitrotyrosine per Phl p 5 monomer. The obtained results were similar to earlier studies investigating other allergenic proteins and reference substances (Bet v 1, BSA). The data will be further interpreted with regard to protein structure, and kinetic modeling will be included in the future.

1 Introduction

The prevalence of allergic diseases and asthma are increasing worldwide, and especially the complexity and severity of allergic diseases in children and young adults continue to rise (Asher et al., 2006; Pawankar et al., 2008, 2013; Pawankar, 2014). The rapid increase in the prevalence, however, especially in industrialized countries, cannot be explained by genetic factors alone, and environmental factors need to be included (Pawankar et al., 2008, 2013; Reinmuth-Selzle et al., 2017; Traidl-Hoffmann et al., 2009). Indeed, several studies have suggested that allergic diseases are enhanced by traffic-related air pollution with high concentrations of the reactive trace gases ozone (O₃) and nitrogen dioxide (NO₂) (D'Amato et al., 2017; Shiraiwa et al.,

2012a), and it has been shown that birch and ragweed pollen from urban areas had a higher allergenic potential than pollen from rural areas (Bryce *et al.*, 2010; Ghiani *et al.*, 2012). Pollen can rupture upon exposure to O₃ or NO₂ or under humid conditions and release cytoplasmic material including allergenic proteins and biogenic adjuvants into the atmosphere (Motta *et al.*, 2006; Ouyang *et al.*, 2016; Taylor *et al.*, 2002, 2004). The released proteins can interact with gaseous and particulate air pollutants leading to the formation of modified proteins. Moreover, allergenic proteins can be chemically modified in the human body during oxidative stress and inflammation induced upon exposure to air pollutants (Reinmuth-Selzle *et al.*, 2017). These chemical posttranslational modifications can change the protein structure and stability, affect hydrophobicity and acidity of binding sites, and thereby alter the immunogenicity of the proteins (Abello *et al.*, 2009; Ackaert *et al.*, 2014; Greenacre and Ischiropoulos, 2001; Ischiropoulos, 2009; Gruijthuisen *et al.*, 2006; Karle *et al.*, 2012).

In the atmosphere, allergenic proteins can react with O₃ and NO₂ resulting in oxidized and nitrated proteins as well as protein oligomerization and degradation (Franze *et al.*, 2005; Shiraiwa *et al.*, 2012b). The reaction requires a two-step mechanism, in which the protein reacts first with ozone, forming long-lived reactive oxygen intermediates (ROIs) such as tyrosyl radicals. In a second step, the ROI reacts either with nitrogen dioxide resulting in the formation of 3-nitrotyrosine (NTyr) residues, combines with another ROI forming dityrosine crosslinks, or undergo further oxidation reactions (Kampf *et al.*, 2015; Liu *et al.*, 2017a; Shiraiwa *et al.*, 2011).

A similar two-step mechanism is known for protein modification inside the human body. During oxidative stress and inflammation, nitrogen oxide radicals (NO) and superoxide anions (O₂⁻) are formed, which in turn rapidly react to the strong oxidizing and nitrating peroxyne (ONOO⁻) (Beckman *et al.*, 1990). This intermediate can react with carbon dioxide (CO₂) forming nitrosoperoxycarbonate (ONOOOCO₂⁻), which in turn decomposes to a nitrogen dioxide radical (NO₂) and a carbonate radical (CO₃⁻) (Gunaydin and Houk, 2009). These one-electron oxidants attack the aromatic ring of a tyrosine residue (Tyr), leading to the formation of a tyrosyl radical. In a second step, the ROI either combines with NO₂ to yield 3-nitrotyrosine or reacts with another ROI forming dityrosine crosslinks, similar to the reaction with O₃ and NO₂ (Gunaydin and Houk, 2009; Pfeiffer *et al.*, 2000).

In this study, we investigated the nitration and oligomerization of the major timothy grass pollen allergen Phl p 5.0101, simply termed Phl p 5 hereafter, by O₃ and NO₂ mimicking air pollution effects and by ONOO⁻ mimicking endogenous protein modification during oxidative stress. The recombinant Phl p 5 has a molecular mass of 28.6 kDa and consists of 278 amino acids including 12 tyrosine residues. The exposure experiments were performed in aqueous phase at low (50 ppb/50 ppb) or high (200 ppb/200 ppb) O₃ and NO₂ concentrations for 0.5-10 h. The endogenous protein modification was performed with 0.07-0.4 mM ONOO⁻ concentrations. Reversed-phase high-performance liquid chromatography coupled to diode array detection analysis was used to determine the total tyrosine nitration degree for the native and modified proteins. Additionally, we used size exclusion chromatography to investigate protein oligomer formation and to determine the individual tyrosine nitration degrees for the different oligomer fractions.

55 **2 Materials and methods****2.1 Chemicals**

Pure water was taken from a Barnstead™ GenPure™ xCAD plus water purification system (Thermo Scientific, Braunschweig, Germany). The water was autoclaved at 121 °C for 20 min and filtered three times through a sterile 0.1 µm pore diameter sterile polyethersulfone (PES) vacuum filter unit (VWR International, Radnor, PA, USA). Timothy grass (*Phleum pratense*) pollen allergen 5 (Phl p 5.0101) was obtained from Biomay AG (Vienna, Austria), and a stock solution (1 mg mL⁻¹) was prepared in pure water prior to each experiment. The solution was incubated at room temperature for at least 30 min to ensure complete reconstitution of the protein. Sodium peroxyxynitrite (160-200 mM) was purchased from Merck (Darmstadt, Germany), and ammonium bicarbonate (NH₄HCO₃) was from Carl Roth GmbH & Co. KG (Karlsruhe, Germany). Water with 0.1 % trifluoroacetic acid (TFA) was purchased from VWR International GmbH (Darmstadt, Germany), and acetonitril (ACN) and sodium phosphate monobasic monohydrate (NaH₂PO₄ · H₂O) were purchased from Carl Roth GmbH & Co. KG. A 150 mM NaH₂PO₄ · H₂O buffer was prepared, and the pH was adjusted to 7 by the addition of sodium hydroxide (NaOH), which was obtained from VWR.

2.2 Protein modification by O₃/NO₂

The experimental setup was described previously (Kampf et al., 2015; Kunert et al., 2018; Liu et al., 2017a; Reinmuth-Selzle et al., 2014), and it was extended by incorporating an oven (Heratherm IGS60, Thermo Scientific) to maintain a constant temperature during the exposure (Fig. S1). Briefly, O₃ was generated by synthetic air passing through a UV lamp (L.O.T.-Oriol GmbH & Co. KG, Darmstadt, Germany) at ~1.9 L min⁻¹, and the O₃ concentration was adjusted by tuning the amount of UV light. The gas flow was humidified by passing through a Nafion® gas humidifier (MH-110-12F-4, PermaPure, Lakewood, NJ, USA) operated with pure water. The gas flow was mixed with N₂ containing ~5 ppmV NO₂ (Air Liquide Deutschland GmbH, Düsseldorf, Germany), and the NO₂ concentration was regulated by the amount of the added ~5 ppmV NO₂ gas. After mixing, the gas flow was split, one flow was equipped with a humidity sensor (FHA 646-E1C with Alemo 2590-3 instrument, Ahlborn, Mess- und Regelungstechnik, Germany) and used to react with the samples, and the other one was used to determine the O₃ and NO₂ concentrations. The O₃ and NO₂ concentrations were monitored using commercial monitoring instruments (ozone analyzer: 49i, Thermo Scientific, Braunschweig, Germany; NOx analyzer: 42i-TL, Thermo Scientific). The O₃/NO₂ gas mixtures were directly bubbled through 1.5 mL of 0.13 mg mL⁻¹ Phl p 5 aqueous solutions at a flow rate of 60 mL min⁻¹ using a Teflon tube (ID: 1.59 mm). The experiments were performed in duplicates.

2.3 Protein modification by ONOO⁻

For each reaction, 300 µL protein stock solution (1 mg mL⁻¹) was transferred into a brown reaction tube (Eppendorf, Hamburg, Germany), and 7.7 µL 2 M NH₄HCO₃ buffer was added to yield a final buffer concentration of 0.05 M. After being thawed on ice, an aliquot of peroxyxynitrite was diluted with 0.3 M NaOH in the ratio 1/1 to reduce the pipetting error for small volumes.

The diluted peroxynitrite solution was added to the protein sample to yield ONOO⁻/Tyr molar ratios of 0.1/1 as well as 0.5/1, and the original peroxynitrite solution was added to yield ONOO⁻/Tyr molar ratios of 1/1, 3/1, and 5/1. The reaction was performed on ice for 110 min. Afterwards, the sample was pipetted into a 10 kDa centrifugal filter (Amicon® Ultra; Merck) and centrifuged at 14 000 g for 2 min (5427 R, Eppendorf). Then, 200 µL pure water was added, and the procedure was repeated
 90 four times. For sample recovery, the filter was turned upside down into a clean micro centrifuge tube and centrifuged at 1 000 g for 2 min. To remove possible sample residues, the filter was washed with 200 µL pure water and centrifuged again upside down into the concentrated protein sample. The experiments were performed in duplicates.

2.4 HPLC-DAD analysis

The total nitration degrees were determined as described previously (Selzle *et al.*, 2013) using an high-performance liquid chromatography coupled to diode array detection (HPLC-DAD; Agilent Technologies 1260 Infinity series, Waldbronn, Germany).
 95 The HPLC system consisted of a quaternary pump (G1311B), an autosampler (G7129A), a column thermostat (G1316C), and a photodiode array detector (G1315C). ChemStation software (Rev. C.01.07, Agilent) was used for system control and data analysis. A monomerically bound C₁₈ column (Vydac 238TP, 250 mm x 2.1 mm i.d., 5 µm, Hichrom, Berkshire, UK) was used for chromatographic separation. Gradient elution was performed at a flow rate of 0.2 mL min⁻¹ with 0.1 % (*v/v*) TFA in water
 100 and ACN. The sample injection volume was 10 µL, and each chromatographic run was performed in duplicate.

The protein oligomer mass fractions and the individual nitration degrees were determined simultaneously as described in Liu *et al.* (2017b) using an HPLC-DAD system (Agilent Technologies 1260 Infinity II series), which consisted of a binary pump (G7112B), a multisampler (G7167A), a column thermostat (G7116A), and a photodiode array detector (DAD, G7115A). ChemStation software (Rev. C.01.08, Agilent) was used for system control and data analysis. Molecular weight separation by size exclusion chromatography (SEC) was carried out using a SEC column (PSS Proteoma BioInert Micro 300 Å, 250 mm x 4.6 mm inner diameter, 3 µm particle size; PSS Polymer Standards Service GmbH, Mainz, Germany). The sample
 105 injection volume was 30 µL, and each chromatographic run was performed in duplicate.

3 Results and discussion

The nitration and oligomerization of recombinant Phl p 5 modified in aqueous solution by O₃/NO₂ and ONOO⁻, respectively,
 110 was studied in a total of 34 experiments. We determined the total nitration degree averaged over all tyrosine residues in the protein and the oligomerization combined with individual nitration degrees for the different protein oligomer mass fractions using reversed-phase and size-exclusion chromatography.

3.1 Protein modification by O₃/NO₂

3.1.1 Total nitration degree

115 Figure 1 shows the effect of low (50 ppb/50 ppb) and high (200 ppb/200 ppb) concentrations of O₃ and NO₂ on the nitration of recombinant Phl p 5. Exposure experiments were performed in aqueous phase and at varying exposure times between 0.5 h and 10 h. The total nitration degree (ND_{tot}) increased over the course of reaction time and with O₃/NO₂ concentrations. The maximum total nitration degree was up to 8 ± 1 % after 10 h of exposure at high O₃/NO₂ concentrations, reaching a plateau after 4 h. This plateau corresponds to ~ 1.1 nitrated Tyr residue per Phl p 5 monomer. No plateau is reached for the exposure at
120 low concentrations, and a linear increase is found with the highest value of 4 ± 1 % after 10 h. We assume that a similar plateau will be reached for the exposure at low concentrations for longer exposure times. Reinmuth-Selzle et al. (2014) investigated the chemical modification of the major birch pollen allergen Bet v 1.0101 by O₃/NO₂ in aqueous phase and found a nitration degree of $\sim 22 \pm 7$ % after 17 h of exposure at 120 ppb O₃ and 120 ppb NO₂. As these reaction conditions are comparable to the ones used in this study (longer exposure times but lower O₃/NO₂ concentrations for Reinmuth-Selzle et al. (2014)), these
125 results suggest that the allergen Bet v 1 is more easily nitrated than Phl p 5. This would mean that the Tyr residues of Bet v 1 are easier accessible for the air pollutants than the ones in the grass pollen allergen. Considering that the Bet v 1 molecule contains in total 7 Tyr residues, the nitration degree would correspond to ~ 1.5 nitrated Tyr residues per Bet v 1 monomer, which is similar to the results obtained in this study.

3.1.2 Oligomerization and individual nitration degree

130 To investigate the temporal evolution of protein dimers and higher oligomers and to determine their individual nitration degrees, the Phl p 5 samples exposed to O₃ and NO₂ were further analyzed by size-exclusion chromatography. Figure 2 shows the effect of low (50 ppb/50 ppb) and high (200 ppb/200 ppb) concentrations of O₃ and NO₂ on the protein oligomerization of recombinant Phl p 5. Generally, the mass fractions of Phl p 5 dimers and higher oligomers increased over the course of reaction time while the mass fraction of Phl p 5 monomers decreased. We found up to 17 ± 2 % dimers at low O₃/NO₂ concentrations
135 and up to 21 ± 3 % at high concentrations after 10 h of exposure (Fig. 2b). This is not a big difference, but the temporal evolution of the dimers differed strongly for the two conditions. At low O₃/NO₂ concentrations, the dimer fraction increased only slightly in the beginning and became stronger after 4 h. In contrast, at high concentrations, the dimer fraction increased rapidly within the first 4 h of exposure reaching a plateau for longer exposure times. A stronger effect of the different O₃/NO₂ concentrations on the oligomerization was found for higher oligomers. Only 7 ± 1 % of higher oligomers ($n \geq 3$) were formed upon exposure
140 to low O₃/NO₂ concentrations, whereas up to 30 ± 5 % were formed at high concentrations after 10 h of exposure (Fig. 2c). The minimum mass fraction of Phl p 5 monomers correspondingly was 76 ± 2 % for low O₃/NO₂ concentrations and 49 ± 2 % for high concentrations (Fig. 2a). Our results are in good agreement with Liu et al. (2017a), who investigated the nitration and oligomerization of the model protein bovine serum albumin (BSA) upon exposure to different O₃/NO₂ concentrations in aqueous phase. They found up to ~ 20 % dimers and up to ~ 5 % of higher oligomers after 10 h of exposure at 50 ppb O₃ and
145 50 ppb NO₂.

Additionally, we determined the individual nitration degrees for the different Phl p 5 oligomer mass fractions (Fig. 3). The individual nitration degrees for the Phl p 5 monomers and dimers increased over the course of reaction time. The maximum nitration degree for the Phl p 5 monomer mass fraction was $3 \pm 1 \%$ for low O_3/NO_2 concentrations and $8 \pm 2 \%$ for high concentrations after 10 h of exposure, reaching a plateau corresponding to ~ 1 nitrated Tyr residue per Phl p 5 monomer (NTyr/Monomer) after 4 h of exposure at high O_3/NO_2 concentrations (Fig. 3a), which is in good agreement with the results obtained by reversed-phase chromatography. The maximum nitration degree for the Phl p 5 dimer mass fraction was $8 \pm 1 \%$ for low O_3/NO_2 concentrations and $12 \pm 2 \%$ for high concentrations after 10 h of exposure (Fig. 3b). A plateau was reached already after 2 h of exposure for both concentrations, corresponding to ~ 2.2 and ~ 3.4 nitrated Tyr residues per Phl p 5 dimer (NTyr/Dimer), respectively. The results for the higher oligomer mass fraction ($n \geq 3$) show a different trend independently of the exposure time and O_3/NO_2 concentrations 3c). The individual nitration degree for this fraction is on average $14 \pm 4 \%$. As we found higher nitration degrees for the dimer than for the monomer mass fraction, our results are in contrast to the results of Liu *et al.* (2017a), who found nitration degrees of $\sim 6 \%$ for the monomer and $\sim 4 \%$ for the dimer mass fraction after 10 h of exposure at 50 ppb O_3 and 50 ppb NO_2 .

The observed time and concentration dependence for the nitration and oligomerization of the major grass pollen allergen Phl p 5 confirms earlier studies with the model protein BSA (Liu *et al.*, 2017a; Shiraiwa *et al.*, 2012b), who discovered that the first and rate limiting step is the reaction with O_3 forming tyrosyl radicals as long-lived reactive oxygen intermediates (ROIs). The second step is then the fast reaction of the ROIs with NO_2 leading to the formation of 3-nitrotyrosine or the reaction with another ROI forming dityrosine cross-links and thereby protein dimers and higher oligomers.

3.2 Protein modification by $ONOO^-$

3.2.1 Total nitration degree

For the investigation of endogenous protein modifications, e.g., during oxidative stress, recombinant Phl p 5 was treated with different concentrations of $ONOO^-$. Experiments were performed in aqueous phase with $ONOO^-$ /Tyr molar ratios of 0.1/1, 0.5/1, 1/1, 3/1, and 5/1. Figure 4 shows the effect of different $ONOO^-$ concentrations on the nitration of recombinant Phl p 5. The total nitration degree (ND_{tot}) increased with increasing $ONOO^-$ concentration. The maximum total nitration degree was up to $25 \pm 3 \%$ for a $ONOO^-$ /Tyr molar ratio of 5/1. For a molar ratio larger than 1/1, the total nitration degree leveled to a plateau between 20 and 25 %, which corresponds to $\sim 3-3.5$ nitrated Tyr residues per Phl p 5 monomer. These results are in good agreement with Reinmuth-Selzle *et al.* (2014), who investigated the nitration of the major birch pollen allergen Bet v 1.0101 by $ONOO^-$. At a $ONOO^-$ /Tyr molar ratio of 5/1, they found a total nitration degree of $\sim 24 \pm 2 \%$.

3.2.2 Oligomerization and individual nitration degree

The temporal evolution of protein dimers and higher oligomers and their individual nitration degrees were analyzed by size-exclusion chromatography. Figure 5 shows the effect of different concentrations of $ONOO^-$ on the oligomerization of recombinant Phl p 5. The highest oligomer formation was found for the $ONOO^-$ /Tyr molar ratios of 0.5/1 and 1/1, reaching up to

26 ± 1 % for dimers and 18 ± 3 % for higher oligomers ($n \geq 3$). The minimum mass fraction of monomers correspondingly was 56 ± 5 %. For ONOO⁻/Tyr molar ratios of 3/1 and 5/1, however, lower oligomer formation was found.

180 Additionally, we determined the individual nitration degrees for the different oligomer mass fractions. Figure 6 shows that the individual nitration degrees for the different Phl p 5 mass fractions increased with increasing concentration of ONOO⁻. The maximum nitration degree for the Phl p 5 monomer mass fraction was 29 ± 1 %, which corresponds to ~ 4 nitrated Tyr residues per Phl p 5 monomer. The maximum nitration degree for the Phl p 5 dimer mass fraction was 37 ± 3 %, which corresponds to ~ 10 nitrated Tyr residues per Phl p 5 dimer. The maximum nitration degree for the Phl p 5 mass fraction of higher oligomers
185 ($n \geq 3$) was 37 ± 4 %.

Our results show that the formation of oligomers predominates at low ONOO⁻ concentrations, whereas the nitration of Tyr residues mainly takes place at high concentrations. This is in good agreement with Pfeiffer et al. (2000), who found that the dimerization of Tyr radicals outcompeted the nitration reaction at low ONOO⁻ concentrations.

4 Conclusions

190 In this study, we analyzed the reaction products of the timothy grass pollen allergen Phl p 5 exposed to O₃ and NO₂ mimicking air pollution and ONOO⁻ mimicking oxidative stress and inflammation. The results show that the tyrosine nitration degree and the formation of protein dimers and higher oligomers depend on the nitrating agent and the reaction conditions (concentration, exposure time). Both, the nitration degree and the oligomer formation, were higher for the reaction with ONOO⁻. The maximum nitration degree of 25% was achieved by reaction with ONOO⁻, while exposure to O₃/NO₂ yielded ND values of up to 8%.
195 Nitrated oligomers were formed upon exposure to O₃/NO₂ and reaction with ONOO⁻. We observed that the tyrosine cross-linking predominates at low ONOO⁻ concentrations, whereas nitrotyrosines are mainly formed at higher concentrations. Our results are in accordance with earlier studies investigating the chemical modification of other proteins by O₃/NO₂ and ONOO⁻. The data will be further interpreted with regard to the protein structure, and kinetic modeling will be included in the future. The various reaction products including nitrated and cross-linked proteins might have different allergenic potential. Thus, related
200 further studies are required to investigate the immunogenicity of the different nitrated and cross-linked variants separately. The efficient nitration and oligomerization of allergenic proteins by atmospheric air pollutants and the enhanced allergenic and immunostimulatory potential of chemically modified proteins call for action to improve air quality and public health in the Anthropocene.

Data availability. All data are available from the corresponding authors upon request.

205 *Author contributions.* KL, UP, and JFN designed the study. ATK, KRS, KZ, CSK, and CW performed the experiments. ATK, KRS, KZ, KL, UP, JF analyzed, interpreted and discussed the results. CSK, CW, and MGW contributed to the discussion. ATK and JFN wrote the manuscript.

Competing interests. The authors declare that they have no conflict of interest.

Acknowledgements. We thank N. Bothen, N.-M. Kropf, and U. Kuhn for technical assistance, B. Sánchez-Parra for helpful discussions, and
210 the Max Planck Society (MPG) for financial support.

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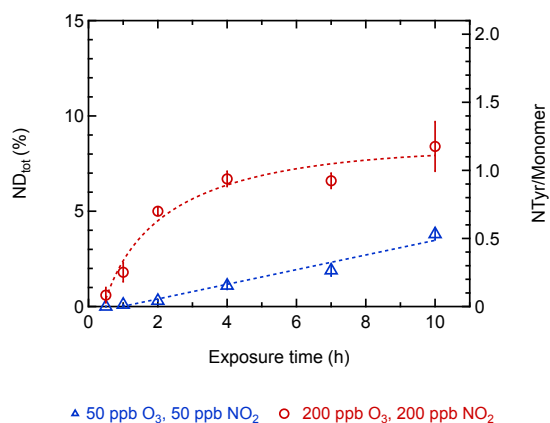


Figure 1. Total nitration degree (ND_{tot}) of PhIp5 exposed to low and high concentrations of O₃ and NO₂ (primary y-axis) and number of nitrotyrosines (NTyr) per PhIp5 monomer (secondary y-axis) for different exposure times. The data points and error bars represent the arithmetic mean values and standard deviations of duplicate experiments. Dashed lines represent preliminary polynomial fits for better visualization.

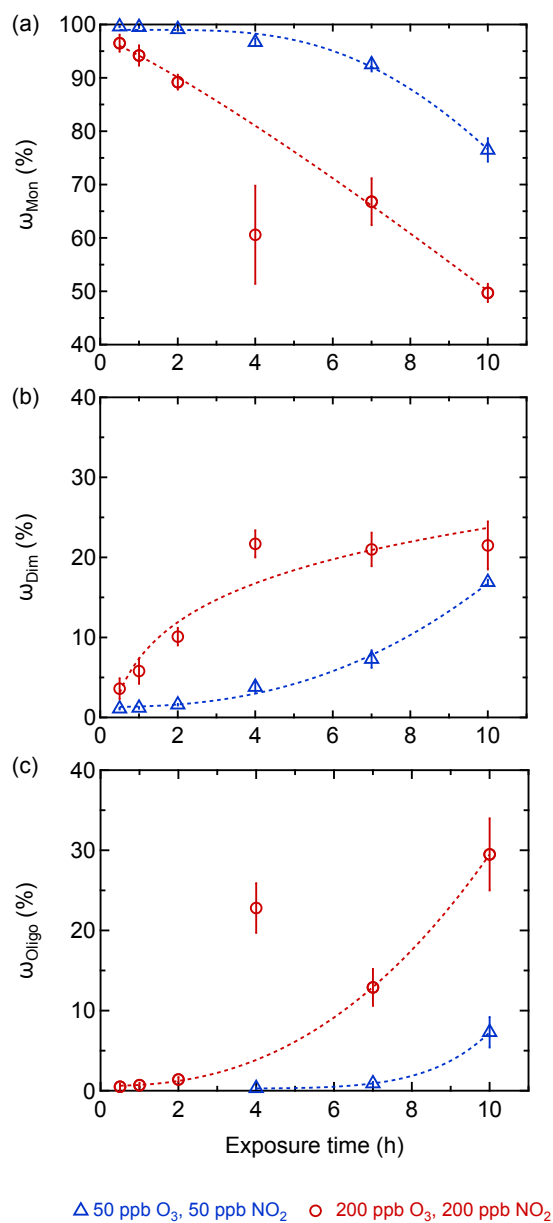


Figure 2. Temporal evolution of protein oligomer mass fractions of PhI p5 exposed to low and high concentrations of O₃ and NO₂ for different exposure times: (a) monomers (ω_{Mon}), (b) dimers (ω_{Dim}), and (c) higher oligomers ($n \geq 3$) (ω_{Oligo}). The data points and error bars represent the arithmetic mean values and standard deviations of duplicate experiments. Dashed lines represent preliminary polynomial fits for better visualization.

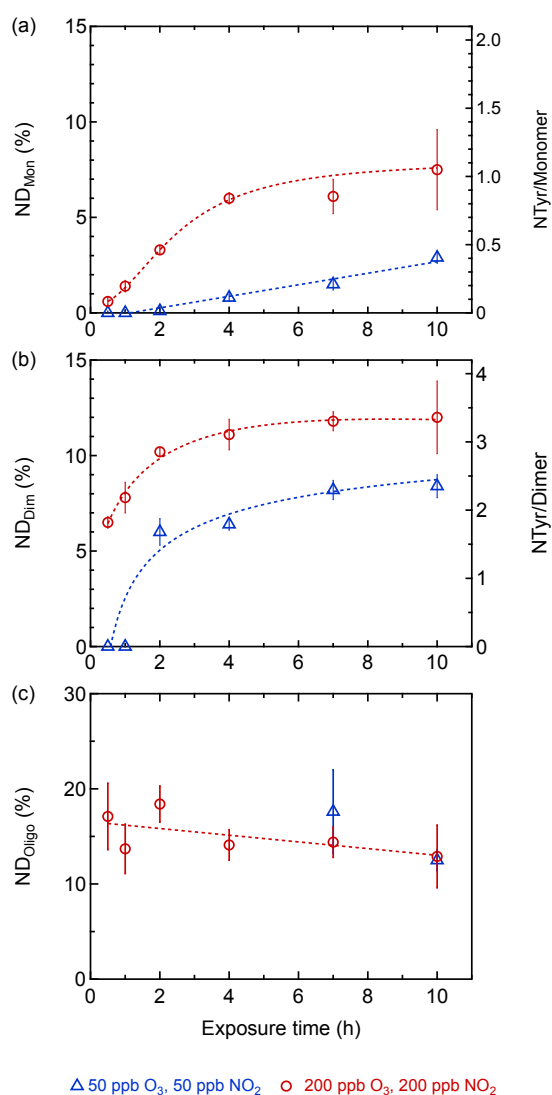


Figure 3. Individual nitration degrees of Phl p 5 oligomer mass fractions (primary y-axis) and number of nitrotyrosines (NTyr) per Phl p 5 monomer and dimer, respectively, (secondary y-axis) for different exposure times: (a) monomers (ND_{Mon}), (b) dimers (ND_{Dim}), and (c) higher oligomers ($n \geq 3$) (ND_{Oligo}). The data points and error bars represent the arithmetic mean values and standard deviations of duplicate experiments. Dashed lines represent preliminary polynomial fits for better visualization.

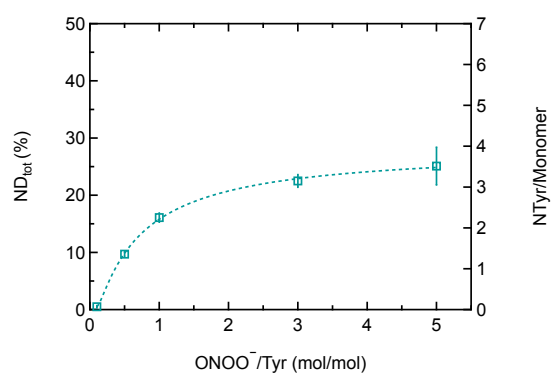


Figure 4. Total nitration degree (ND_{tot}) of Phlp5 (primary y-axis) and number of nitrotyrosines (NTyr) per Phlp5 monomer (secondary y-axis) for different $ONOO^-$ concentrations. The data points and error bars represent the arithmetic mean values and standard deviations of duplicate experiments. Dashed lines represent preliminary polynomial fits for better visualization.

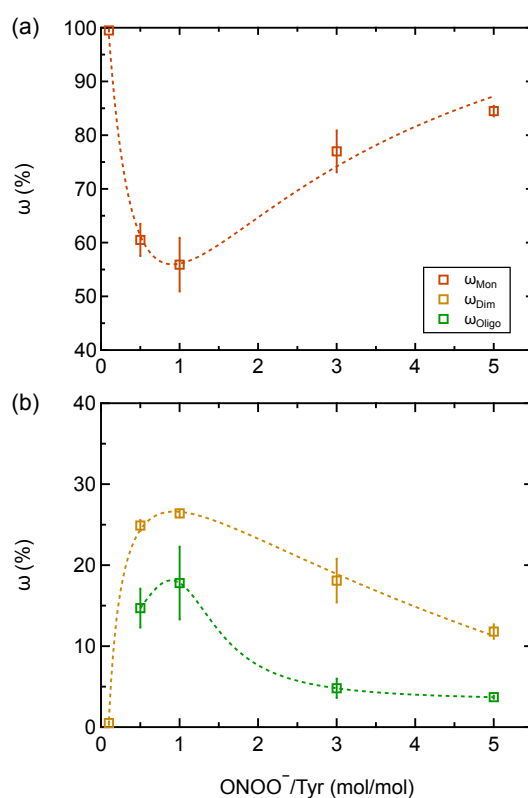


Figure 5. Formation of protein oligomer mass fractions of Phlp5 treated with different ONOO⁻ concentrations: (a) monomers (ω_{Mon}), (b) dimers (ω_{Dim}) and higher oligomers ($n \geq 3$) (ω_{Oligo}). The data points and error bars represent the arithmetic mean values and standard deviations of duplicate experiments. Dashed lines represent preliminary polynomial fits for better visualization.

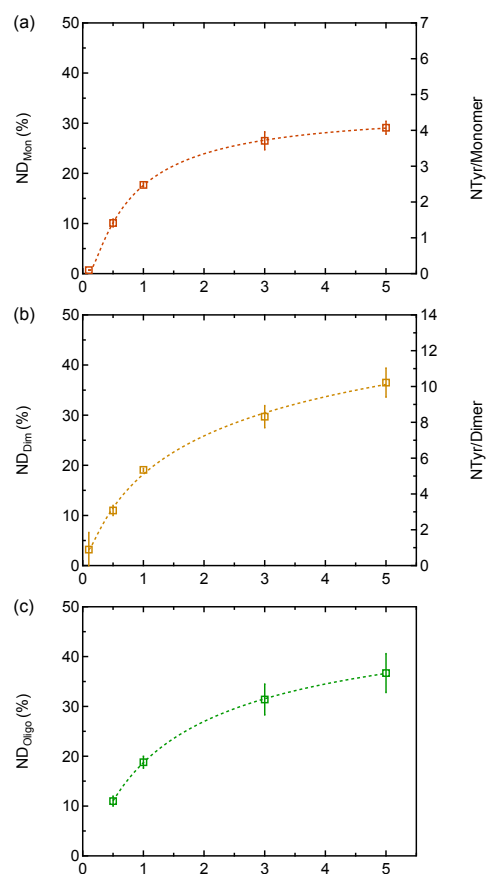


Figure 6. Individual nitration degrees of Phlp5 oligomer mass fractions (primary y-axis) and number of nitrotyrosines (NTyr) per Phlp5 monomer and dimer, respectively, (secondary y-axis) for different ONOO⁻ concentrations: (a) monomers (ND_{Mon}), (b) dimers (ND_{Dim}), and (c) higher oligomers ($n \geq 3$) (ND_{Oligo}). The data points and error bars represent the arithmetic mean values and standard deviations of duplicate experiments. Dashed lines represent preliminary polynomial fits for better visualization.

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Chemical modification of pro-inflammatory proteins by peroxynitrite increases activation of TLR4 and NF- κ B: Implications for the health effects of air pollution and oxidative stress

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submitted to Redox Biology (2020)

Author contributions

KL and UP conceived and directed the study. KZ, ATK, KRS, ALL, and KL designed and performed the experiments. KZ, ATK, KRS, MGW, JFN, KL, and UP analyzed, interpreted and discussed the data. DW and DS contributed to the design, interpretation, and discussion. All authors contributed to the preparation and editing of the manuscript. KZ and ATK contributed equally.

Chemical modification of pro-inflammatory proteins by peroxynitrite increases activation of TLR4 and NF- κ B: Implications for the health effects of air pollution and oxidative stress

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Keywords

Protein modification

TLR4

Peroxynitrite

α -Synuclein

HSP60

HMGB1

Abbreviations

α -Syn - α -Synuclein

HSP60 - heat shock protein 60

HMGB1 - high-mobility-group box 1 protein

OVA - ovalbumin

TLR4 - Toll-like receptor 4

NF- κ B - nuclear factor 'kappa-light-chain-enhancer' of activated B-cells

DAMP - damage-associated molecular patterns

IL - interleukin

Abstract

Environmental pollutants like fine particulate matter can cause adverse health effects through oxidative stress and inflammation. Reactive oxygen and nitrogen species (ROS/RNS) such as peroxynitrite can chemically modify proteins, but the effects of such modifications on the immune system and human health are not well understood. In the course of inflammatory processes, the Toll-like receptor 4 (TLR4) can sense damage-associated molecular patterns (DAMPs). Here, we investigate how the TLR4 response and pro-inflammatory potential of the proteinous DAMPs α -Synuclein (α -Syn), heat shock protein 60 (HSP60), and high-mobility-group box 1 protein (HMGB1), which are relevant in neurodegenerative and cardiovascular diseases, changes upon chemical modification with peroxynitrite.

For the peroxynitrite-modified proteins, we found a strongly enhanced activation of TLR4 and of the pro-inflammatory transcription factor NF- κ B in stable reporter cell lines as well as increased mRNA expression and secretion of the pro-inflammatory cytokines TNF- α , IL-1 β , and IL-8 in human monocytes (THP-1). This enhanced activation of innate immunity via TLR4 is mediated by covalent chemical modifications of the studied DAMPs.

Our results show that proteinous DAMPs modified by peroxynitrite more potently amplify inflammation via TLR4 activation than the native DAMPs, and provide first evidence that such modifications can directly enhance innate immune responses via a defined receptor. These findings suggest that environmental pollutants and related ROS/RNS may play a role in promoting acute and chronic inflammatory disorders by structurally modifying the body's own DAMPs. This may have important consequences for chronic neurodegenerative, cardiovascular or gastrointestinal diseases that are prevalent in modern societies, and calls for action, to improve air quality and climate in the Anthropocene.

1. Introduction

Reactive oxygen and nitrogen species (ROS/RNS) like peroxynitrite play important roles in oxidative stress and adverse health effects induced upon exposure to environmental pollutants and in the course of inflammatory diseases [1–8]. Peroxynitrite (ONO_2^-) is generated upon reaction of superoxide (O_2^-) and nitric oxide (NO) [9]. It can react with amino acids like tyrosine, leading to the formation of nitrotyrosine, dityrosine, and protein oligomers [10–15]. Nitrotyrosine and dityrosine are known as markers of inflammation and oxidative stress, which can influence the chemical and physiological properties of proteins [16–19]. For example, nitration can change the binding of proteins to certain receptors and thus modulate downstream signaling cascades [12,14,20–22]. Notably, preferential recognition of nitrotyrosine epitopes by immunoglobulins was reported for several inflammatory diseases [23–25]. Dityrosine crosslinks can be intra- or intermolecular, altering protein structures, causing protein aggregation/oligomerization, and affecting protein function [13,26,27].

During inflammation, damage-associated molecular patterns (DAMPs) can be actively secreted by epithelial cells, monocytes, macrophages, and other cells, or passively released by damaged and dying cells [28]. Proteinous DAMPs can have various intracellular functions, e.g., as chaperones, and when released as extracellular proteins, they can stimulate pattern recognition receptors (PRR) such as the Toll-like receptor 4 (TLR4) [29]. TLR4 signaling leads to the activation of transcription factors like the nuclear factor 'kappa-light-chain-enhancer' of activated B-cells (NF- κ B) and the interferon regulatory transcription factor 3 (IRF3), which are key activators of inflammatory cascades [30,31]. NF- κ B induces the expression and modulates the secretion of pro-inflammatory cytokines such as TNF- α , IL-1 β , and IL-8, which can stimulate the corresponding cytokine receptors leading to further activation of NF- κ B and other signaling pathways [30,32–34]. This positive feedback can amplify and propagate inflammatory processes in autocrine or paracrine fashion [35,36].

In this study, we investigate if and how peroxynitrite can enhance the activation of TLR4 and NF- κ B by chemical modification of disease-related proteins acting as DAMPs (Table S1): α -Synuclein, heat shock protein 60, and high-mobility-group box 1 protein (HMGB1).

α -Synuclein (α -Syn) is a neuronal protein that regulates exocytosis and endocytosis of synaptic vesicles as well as ATP synthase [37–40]. Oxidative and nitrating conditions can lead to the formation of nitrated, misfolded or aggregated α -Syn, which is linked to neurodegenerative disorders such as Parkinson's disease [39,41–46]. Oligomeric and misfolded α -Syn can be released from neuronal cells by exocytosis and stimulate TLR4 and other receptors [43,47–49].

Heat shock protein 60 (HSP60) is a mitochondrial chaperone that assists correct folding of imported mitochondrial proteins [50,51]. HSP60 can be released by stressed or necrotic cells, and modulate immune response by stimulating TLR4 and other receptors [52–58]. It contributes to the pathogenesis of chronic inflammatory diseases such as Crohn's disease, diabetes, and atherosclerosis [59–63].

High-mobility-group box 1 protein (HMGB1) is a ubiquitously expressed nuclear protein involved in DNA replication, recombination, transcription, and repair [29,64–66]. HMGB1 can be released by activated or damaged cells and secreted by immune cells in response to microbial or pro-inflammatory stimuli [66–69]. HMGB1 can stimulate TLR4, the advanced glycosylation end product-specific receptor (RAGE), and other receptors [28,65,70–72]. It is involved in severe acute and chronic diseases, such as sepsis, cancer, cardiovascular and neurodegenerative diseases [65,66,69,73,74].

The investigated proteins were exposed to peroxynitrite, and covalent chemical modifications were analyzed by liquid chromatography (HPLC), gel electrophoresis (SDS-PAGE), and western blots (anti-nitrotyrosine, anti-dityrosine). TLR4 and NF- κ B activation were determined in stable reporter cell lines using bioluminescence detection (HeLa TLR4, THP-1 NF- κ B). Moreover, mRNA expression and secretion of various cytokines were measured in human monocytes (THP-1) as detailed below.

2. Materials and Methods

2.1 Protein modification and analysis

2.1.1 Peroxynitrite modification

Protein stock solutions (1 mg mL⁻¹) of α -Syn (Merck Millipore, Darmstadt, Germany), HSP60 (Abcam, Cambridge, UK), HMGB1 (Sigma Aldrich, Taufkirchen, Germany), and Ovalbumin (InvivoGen, Toulouse, France) were prepared in PBS (Thermo Fisher Scientific, Darmstadt, Germany). Ammonium bicarbonate ($\geq 98\%$, Ph. Eur., BP, Carl Roth, Karlsruhe, Germany) was dissolved in water to yield a final buffer concentration of 2 M, and the pH was adjusted to 7.8 by the addition of 1 M hydrogen chloride (37 % stock solution, Merck Millipore). For each reaction, 300-500 μ L protein solution was transferred into a brown reaction tube (Eppendorf, Hamburg, Germany), and 7.5-12.5 μ L ammonium bicarbonate buffer (2 M) was added to yield a final buffer concentration of 50 mM. After being thawed on ice, 1-8 μ L sodium peroxynitrite solution (160-200 mM, Merck Millipore) was added to yield a molar ratio of peroxynitrite to tyrosine residues of 5:1, and the reaction was performed on ice for 110 min. Thereafter, the

sample was pipetted into a 10 kDa centrifugal filter (Amicon®, Merck Millipore) and centrifuged at 14,000 x g for 2 min (5427 R, Eppendorf). The sample was washed five times with 200 µL PBS and centrifugation at 14,000 x g for 2 min. For sample recovery, the filter was turned upside down, transferred into a clean microcentrifuge tube, and centrifuged at 1,000 x g for 2 min. To recover possible sample residues, the filter was washed with 200 µL fresh PBS and centrifuged upside down at 1,000 x g for 2 min into the concentrated protein sample. Ovalbumin (OVA), which is not a TLR4 agonist, was treated the same way and served as negative control in the cell culture experiments described below. For mock samples, 500 µL pure PBS was mixed with 12.5 µL ammonium bicarbonate buffer (2 M) and 8 µL peroxyntirite and treated as described above.

Protein concentrations were determined using a Synergy Neo plate reader (BioTek, Bad Friedrichshall, Germany) measuring the absorbance at 260 nm / 280 nm. For each protein, a dilution series of the native protein (25-1000 µg mL⁻¹) was used for calibration. Three microliters of modified protein solution were transferred into a micro-volume plate in triplicates (Take3 trio, BioTek), and the absorbance was measured. The protein concentrations were confirmed by SDS-PAGE and silver stain (see section 2.1.3).

For each of the investigated proteins, multiple samples of protein solution were chemically modified as outlined above, and selected samples were characterized by the analytical methods described below. An overview of the analytical results is given in Table S2 (tyrosine nitration degree, dimer/oligomer fraction).

2.1.2 HPLC-DAD analysis

For the investigated proteins, tyrosine nitration degrees were determined as described in Selzle *et al.* [75]. Briefly, an HPLC-DAD system (Agilent Technologies 1260 Infinity series, Waldbronn, Germany) equipped with a monomerically bound C₁₈ column (Vydac 238TP, 250 mm x 2.1 mm i.d., 5 µm, Hichrom, Berkshire, UK) was used for chromatographic separation. Gradient elution was performed with 0.1 % (v/v) trifluoroacetic acid in water (VWR International GmbH, Darmstadt, Germany) and acetonitrile (Carl Roth), and absorbance was measured at wavelengths of 280 nm and 357 nm. Injection volume was 10 µL, and each chromatographic run was performed in duplicates. For system control and data analysis, ChemStation Software was used (Rev. C.01.07, Agilent). Nitration was observed in the modified samples of all proteins, and tyrosine nitration degrees were quantified in two independent experiments as specified in Table S2. The tyrosine nitration degree is defined as the concentration of nitrotyrosine divided by the sum of the concentrations of nitrotyrosine and tyrosine [75].

2.1.3 SDS-PAGE and silver stain

Protein oligomerization was visualized and quantified by silver stained SDS-PAGE (Thermo Fisher Scientific). Protein samples dissolved in PBS were mixed with an equivalent volume of 2x Laemmli buffer, containing 65.8 mM Tris-HCl (pH 6.8, Carl Roth), 26.3 % glycerol (v/v, Carl Roth), 2.1 % SDS (Carl Roth), 0.02 % bromophenol blue (Sigma Aldrich) and 5.0 % 2-mercaptoethanol (Sigma Aldrich), and heated at 95 °C for 5 min. The samples (50 ng in 10 µL) were loaded onto PROTEAN Precast gels (4–20 %, Bio-Rad, Munich, Germany) together with 5 µL Color Prestained Protein Standard, Broad Range (11–245 kDa, New-England Biolabs, Frankfurt, Germany) and stained with a silver stain kit (Thermo Fisher Scientific) following manufacturer's instructions. For image acquisition and quantification of protein monomers, dimers, and higher oligomers, a ChemiDoc system (Bio-Rad) with Image Lab software 5.2.1 (Bio-Rad) was used. Protein dimers occurred in the modified samples of all proteins (Fig. S1, Table S2), and signals of higher oligomers were observed but exceeded the analytical detection limit (3 %) only for modified HSP60 (23 ± 13 %).

2.1.4 Western blot analysis

The presence of nitrotyrosine and dityrosine residues was investigated by SDS-PAGE and subsequent western blot analysis. The native and modified protein samples were prepared in Laemmli-buffer as described in section 2.1.3. As sensitivities for the antibodies varied, different amounts of protein per lane were applied for nitrotyrosine staining (2 µg α-Syn, 5 µg HSP60, and 10 µg HMGB1) and dityrosine staining (5 µg α-Syn, 5 µg HSP60, and 5 µg HMGB1). For nitrotyrosine staining, α-Syn, HSP60, and HMGB1 were separated by a 12 %, 8 %, and 10 % SDS polyacrylamide gel, respectively, and by a Mini-PROTEAN Precast gel (4-20 %, Bio-Rad) for dityrosine staining. Thereafter, the gels were transferred onto 0.45 µm PVDF membranes (Merck Millipore) using a semi-dry transfer unit (Hoefer, Holliston, MA, USA). The membranes were blocked with 5 % fat-free milk powder (Cell Signaling Technology, Leiden, Netherlands) in Tris-buffered saline with 0.1 % Tween-20 (TBS-T) at RT for 2 h, followed by an overnight incubation at 4 °C with nitrotyrosine antibody (mouse monoclonal [clone HM11], 1:1000, Cat # 321900, Thermo Fisher Scientific) and dityrosine antibody (mouse monoclonal [clone 7D4], 1:10,000, Cat # NBP2-59360, Novus Biologicals, Centennial, CO, USA), respectively, diluted in 5 % fat-free milk powder in TBS-T. All membranes were washed four times with TBS-T for 5 min each and incubated with the horseradish peroxidase coupled secondary antibody (goat anti-mouse, polyclonal, 1:10,000, Cat # 115-035-062, Jackson Immuno Research Europe Ltd, Ely, UK) at RT for 1 h, diluted in TBS-T. For detection, the membranes were washed four times with TBS-T for 5 min, and protein bands were developed using the chemiluminescence system ECL-femto (Thermo Fisher Scientific) for nitrotyrosine and ECL-plus (Thermo Fisher Scientific) for dityrosine according to the manufacturer's protocol. The bands were detected in a ChemiDoc system, and images were processed using Image Lab software 5.2.1.

Nitrotyrosine and dityrosine were detected in peroxynitrite-modified samples of all investigated proteins. For the native proteins, neither nitrotyrosine nor dityrosine were detectable in the western blots (Fig. S1).

2.1.5 Endotoxin quantification

The amount of endotoxin in native and modified protein samples was quantified by a Pierce™ LAL Chromogenic Endotoxin Quantitation Kit (Thermo Fisher Scientific) according to the manufacturer's protocol. For this purpose, protein stock solutions were diluted to 1 µg mL⁻¹ with endotoxin-free water. The endotoxin level in all protein samples was less than 1 EU per µg of protein.

2.2 Cell culture experiments

All cell experiments were performed with the same final concentrations of native and modified protein samples (α-Syn: 50.0 µg mL⁻¹, HSP60: 14.3 µg mL⁻¹, HMGB1: 3.6 µg mL⁻¹, OVA: 50.0 µg mL⁻¹) except for mRNA expression of α-Syn (35.7 µg mL⁻¹). LPS from *E. coli* O111:B4, 25 ng mL⁻¹, InvivoGen) was used as positive control (LPS-EB for HeLa TLR4 experiments; LPS-EB ultrapure for THP-1-Lucia™ and THP-1 experiments), and medium and mock served as negative controls. Protein solutions were diluted in the respective cell culture medium.

2.2.1 HeLa TLR4 dual reporter cells

Using a HeLa TLR4 dual luciferase reporter cell line (Novus Biologicals), we simultaneously determined TLR4 activation and cell viability [15]. Cells were grown in Dulbecco's Modified Eagle's Medium (DMEM, Thermo Fisher Scientific) containing 25 mM D-glucose and 1 mM sodium pyruvate supplemented with 10 % heat-inactivated fetal bovine serum (FBS superior, Cat # S0615, Lot #0973F, Biochrom, Berlin, Germany), 1 % penicillin/streptomycin (Thermo Fisher Scientific) and 140 µg mL⁻¹ hygromycin B (InvivoGen) in a humidified atmosphere of 5 % CO₂ at 37 °C. For each sample and replicate, 20,000 HeLa TLR4 dual reporter cells in 100 µL medium per well were seeded in a flat-bottom 96-well plate (Greiner, Frickenhausen, Germany). After 24 h, the cells were incubated in triplicates with 50 µL of the respective protein or control sample for 7 h. Thereafter, cells were washed with 200 µL PBS and lysed using 15 µL passive lysis buffer (Promega, Mannheim, Germany) at -80 °C overnight. The read out was performed using the Dual-Luciferase® Reporter Assay System (Promega) according to the manufacturer's protocol. Both luminescence signals were measured in a Synergy Neo plate reader. To calculate the normalized TLR4 activity, TLR4-driven *Renilla* luciferase signal was divided by the constitutive, CMV-driven firefly luciferase signal that served as a surrogate marker for cell viability. For each experiment, LPS-treated cells were used as a positive

control, and the arithmetic mean was set to 100 %. This value was used to normalize the measurement results of all protein and control samples. Arithmetic mean values and standard deviations were calculated from the normalized values of three (HSP60) or four (α -Syn, HMGB1, OVA) independent experiments performed in triplicates.

2.2.2 THP-1 NF- κ B reporter cells and receptor antagonists

To assess the NF- κ B activity of cells treated with different native and modified proteins, THP-1-Lucia™ NF- κ B cells (InvivoGen) were grown in Roswell Park Memorial Institute (RPMI) 1640 medium (Thermo Fisher Scientific) containing 25 mM D-glucose and 1 mM sodium pyruvate supplemented with 10 % heat-inactivated FBS, 100 μ g mL⁻¹ Zeocin™ (InvivoGen), and 1 % penicillin/streptomycin in a humidified atmosphere of 5 % CO₂ at 37 °C. For each sample and replicate, 100,000 cells in 50 μ L medium per well were seeded in a flat-bottom 96-well plate. To inhibit TLR4 or RAGE signaling, cells were pre-incubated with 50 μ L of the TLR4 antagonist TAK242 or the RAGE antagonist FPS-ZM1 for 4 h. TAK242 (25 mM, Merck Millipore) was provided in dimethyl sulfoxide (DMSO) and further diluted with medium to a final concentration of 0.36 μ g mL⁻¹. FPS-ZM1 (25 mg, Merck Millipore) was dissolved in 250 μ L DMSO and further diluted with medium to a final concentration of 3.3 μ g mL⁻¹. If no antagonist was applied, 50 μ L medium was added to the cells. Subsequently, cells were incubated in triplicates with 50 μ L of the respective protein or control samples for 24 h. Cells treated with DMSO in medium (4.4 μ g mL⁻¹) showed no NF- κ B activation or toxic effects. Activity of NF- κ B was measured by QUANTI-Luc™ reagent (InvivoGen) according to manufacturer's instructions. Briefly, 10 μ L of cell culture supernatant was transferred into a white plate (LUMITRAC™, Greiner), and mixed with 50 μ L of QUANTI-Luc™ reagent. The luminescence was detected in a Synergy Neo plate reader. For each experiment, LPS-treated cells were used as positive control, and the arithmetic mean was set to 100 %. This value was used to normalize the measurement results of all protein and control samples. Arithmetic mean values and standard deviation were calculated from the normalized values of three (α -Syn) or four (HSP60, HMGB1, OVA) independent experiments performed in triplicates. Assessment of cell viability was performed in triplicates using the alamarBlue™ cell viability reagent (Thermo Fisher Scientific) according to the manufacturer's protocol. Excitation was performed at 560 nm, and emission was measured at 590 nm in a Synergy Neo plate reader.

2.2.3 Cytokine immunoassay

Human THP-1 monocyte cells (ATCC, Manassas, VA, USA) were grown in RPMI 1640 medium supplemented with 10 % heat-inactivated FBS, 0.05 mM 2-mercaptoethanol (Sigma Aldrich), and 1 % penicillin/streptomycin. Quantification of cytokine secretion was performed using a multiplex immunoassay. For each sample and replicate, 100,000 human THP-1

monocytes in 100 μ L medium per well were seeded in a V-shaped 96-well plate (Greiner) and incubated with 50 μ L of the respective protein or control samples over 24 h. Thereafter, cells were centrifuged at 200 x g for 5 min (5427 R, Eppendorf), and the supernatant was analyzed in duplicates for TNF- α , IL-1 β , and IL-8 by a multiplex assay kit (R&D systems Inc., Minneapolis, MN, USA) according to the manufacturer's protocol. The read-out was performed on a MAGPIX device (Luminex, Austin, TX, USA), and arithmetic mean values and standard deviation of the duplicates were calculated. Three independent experiments were performed showing similar trends. Assessment of cell viability was performed as described in section 2.2.2.

2.2.4 mRNA extraction and qPCR analysis

Quantification of mRNA expression of TNF- α , IL-1 β , IL-8, and CXCL-10 was performed using real-time quantitative PCR (qPCR). For each sample and replicate, 400,000 human THP-1 monocyte cells in 1 mL medium per well were seeded in a 6-well cell culture plate (Greiner). On the next day, cells were incubated with 1 mL of the respective protein or control samples for 4 h. For qPCR analysis, cells were harvested by centrifugation at 500 x g for 5 min. Total RNA was extracted from cells using RNeasy Mini Kit (Qiagen, Hilden, Germany) following the spin technology protocol. Total RNA yield was determined using a Take3 trio micro-volume plate in a Synergy Neo plate reader. Using the High Capacity cDNA Reverse Transcription Kit (Thermo Fisher Scientific), 500 ng of total RNA per sample were transcribed into cDNA. Afterwards, qPCR was performed in duplicates using 10 ng of cDNA mixed with SsoAdvanced Universal SYBR Green Supermix (Bio-Rad) to a final concentration of 333 nM for each primer, which were designed using PrimerBlast (NCBI) and the respective template genes (Table S3). As reference genes served Peptidylprolyl isomerase A (PPIA) and TATA-binding protein (TBP). Reactions were performed at 98 °C for 30 s, followed by 37 cycles of 98 °C for 10 s, and 60 °C for 25 s. Gene expression was calculated according to the $2^{-\Delta\Delta CT}$ method using the Bio-Rad CFX Manager Software 3.1. Arithmetic mean values and standard deviation of the duplicates were calculated. Three independent experiments were performed showing similar trends.

2.2.5 Statistical analyses

GraphPad Prism version 6.07 (GraphPad, San Diego, CA, USA) was used for statistical analysis. Unpaired t-tests were performed to observe differences between the native and peroxynitrite-modified proteins. The results were considered as significant when * $p < 0.05$, ** $p < 0.01$, *** $p < 0.005$.

3. Results

For the investigated proteins in native and peroxynitrite-modified form, TLR4 and NF- κ B activation were determined in stable reporter cell lines (HeLa TLR4, THP-1 NF- κ B). Additionally, mRNA expression and secretion of various cytokines induced by NF- κ B were measured in human monocytes (THP-1).

As shown in Figures 1 and 2, chemical modification by peroxynitrite significantly increased activation of TLR4 and NF- κ B for all three investigated proteins. Relative to the native protein, the TLR4 activity increased by a factor of ~1.4 for α -Syn and HSP60 and by a factor of ~2.2 for HMGB1 (Fig. 1, Table S4), while the NF- κ B activity increased by a factor of ~1.6 for α -Syn and HSP60 and by a factor of ~4.2 for HMGB1 (Fig. 2, Table S4). Inhibition of the TLR4 receptor by the antagonist TAK242 reduced the NF- κ B response to both, the modified and native protein by more than 80 % for α -Syn, by more than 90 % for HSP60, and by 30-40 % for HMGB1, while the RAGE inhibitor FPS-ZM1 had no substantial effect (Figs. 2). These inhibition experiments indicate that the enhancement of NF- κ B activation by peroxynitrite-modification is predominantly mediated by TLR4 for α -Syn and HSP60, while for HMGB1 also other receptors such as RAGE and TLR2 may be involved [70,71].

Figure 3 shows that the chemically modified proteins also enhanced the secretion of the pro-inflammatory cytokines TNF- α , IL-1 β , and IL-8. Relative to the native protein, the secretion increased 1.2-2.0 fold for α -Syn, 1.2-8.8 fold for HSP60, and 2.1-16.7 fold for HMGB1 (Table S5). The mRNA expression of the investigated cytokines showed the same trend for modified HSP60 and HMGB1, but not for modified α -Syn (Fig. S2, Table S6). The mRNA results obtained for CXCL-10 indicate that the chemically modified proteins increased not only the activation of NF- κ B (MyD88 pathway), but also the activation of the IRF3 (TRIF/TRAFF) pathway of TLR4 signaling [76,77]. The negative controls of native and peroxynitrite-modified OVA exhibited no substantial TLR4 and NF- κ B activity, cytokine secretion, or mRNA expression (Figs. 1-3, Fig. S2), and the applied protein concentrations did not affect the viability of the investigated cells (Figs. S3-S5), suggesting that the induced protein modifications were only relevant in the context of the TLR4-activating proteins.

Overall, our experimental results demonstrate that peroxynitrite treatment enhances the inflammatory potential of α -Syn, HSP60, and HMGB1 *in vitro*. This can be explained by changes in protein-receptor interactions related to chemical modifications like nitrotyrosine, intramolecular dityrosine crosslinks, protein dimers and higher oligomers as detected in the peroxynitrite-modified samples of the investigated proteins.

4. Discussion

Nitrotyrosine was detected in all chemically modified protein samples (Fig. S1, Table S2), and the introduction of a nitro group can strongly affect the chemical and physiological properties of a protein [78]. Notably, nitrotyrosine is more acidic than tyrosine, which can affect the isoelectric point of a protein, change the binding to receptors and ligands, and modulate downstream signaling cascades [12,14,20–22]. Nitration tends to enhance the immunogenicity of proteins, and nitrated proteins are thus used as model antigens or allergens [2,24,25,79–83]. The reason for this striking immunological property of nitrated proteins, which has also been linked to autoimmunity, is not completely clear [16,84]. The generation of neo-epitopes seems to be an important factor in adaptive immune responses [22,24], and nitrotyrosine was found to be associated with increased TLR4 signaling and innate immune responses related to chronic inflammatory diseases [15,18,78,85–88]. However, we here show for the first time that there is a direct enhancement of TLR4 activation by the peroxynitrite-induced modification of natural (human) DAMPs that play a role as innate immune activators in acute and chronic inflammation.

Besides nitrotyrosine, also dityrosine was detected in all chemically modified protein samples (Fig. S1). The detection of dityrosine in the monomeric form of the modified proteins implies that intramolecular crosslinks were formed and may have induced conformational changes affecting protein-receptor interactions. For HMGB1, it is well known that TLR4 binding depends on the oxidation state of the protein, and that an intramolecular disulfide bridge is crucial for TLR4 dimerization and activation [89,90]. By analogy, intramolecular dityrosine crosslinks among the tyrosine residues in HMGB1 might also induce or stabilize conformational arrangements that are relevant for protein-receptor interactions and TLR4 activation. Further investigations will be required to clarify if and how intramolecular dityrosine crosslinks may contribute to the enhancement of TLR4 activity observed for peroxynitrite-modified HMGB1 and other proteins.

Protein dimers and higher oligomers were detected in all chemically modified protein samples (Fig. S1, Table S2). They comprise more than one receptor interaction domain, can act as multivalent ligands, and may thus more efficiently promote TLR4 dimerization and signaling [15,91–93]. Such multivalency effects are common in biological systems, where higher functional affinities are observed for dimeric and oligomeric agonists, leading to enhanced receptor clustering and signal transduction [94–97]. Thus, the formation of dimers and oligomers by intermolecular dityrosine crosslinking may well explain the enhanced TLR4 and NF- κ B activity observed upon stimulation with peroxynitrite-modified samples compared to the native form of the investigated proteins. For α -Syn, this is consistent with earlier studies

reporting that oligomeric α -Syn can enhance TLR4 signaling and pro-inflammatory cytokine production in microglia and astrocytes, which may contribute to development and progression of Parkinson's disease [43,98]. Similar effects have been reported for the interaction of amyloid beta in Alzheimer's disease, where aggregates trigger TLR2 and TLR4 [99,100], and nitration critically enhances amyloid beta aggregation and plaque formation [101]. For all receptors of the TLR family [102,103] as well as RAGE [104–106] and other receptors [107,108], activation and signal transduction requires dimerization of the receptor molecules, which may be promoted by dimerized or oligomeric ligands. Thus, we suggest that the peroxynitrite-induced dimerization or oligomerization of HSP60 and HMGB1 may also play a critical role for PRR signaling in inflammatory diseases.

As outlined above, the proteins α -Syn, HSP60, and HMGB1 are involved in many severe diseases. If *in vivo* studies confirm that nitrotyrosine residues or dityrosine crosslinks in proteinous DAMPs play an important role in amplifying inflammatory processes through enhanced TLR4 signaling, the chemically modified proteins may serve as useful clinical markers providing mechanistic insight. Such markers and insight may also help to advance medical treatment options that involve scavenging of peroxynitrite [109–111] or the inhibition of peroxynitrite formation [112,113].

5. Conclusions

In this study, we have shown that peroxynitrite can induce chemical modifications that enhance the TLR4 activity and inflammatory potential of proteinous DAMPs like α -Syn, HSP60, and HMGB1. Besides peroxynitrite, a wide range of other endogenous or exogenous ROS/RNS can also modify the chemical structure, properties, and effects of proteins [13]. In particular, air pollutants such as fine particulate matter, nitrogen oxides, and ozone can trigger or enhance oxidative stress, nitration and oligomerization of proteinous DAMPs and allergens, immune reactions, and feedback cycles of inflammation [1,2,13,15,36,79,114–116].

Figure 4 illustrates how chemically modified DAMPs may amplify oxidative stress and innate immune responses through a positive feedback loop of pro-inflammatory signaling via TLR4 or other PRR. Such feedback and self-amplification provide a potential mechanistic rationale for the development of inflammatory disorders related to environmental pollution. In particular, it may help to explain the massive burden of disease attributable to air pollution, where the underlying chemical and physiological mechanisms are not yet well understood [4–6]. Environmental pollutants may generate exogenous ROS/RNS and oxidative stress, triggering inflammatory processes that lead to the formation of endogenous ROS/RNS and release of DAMPs. The DAMPs can activate PRR (TLR, RAGE, etc.) that induce further pro-

inflammatory signaling and responses via transcription factors (NF- κ B, IRF3, etc.), cytokines and cytokine receptors (IL-1, IL-8, etc.). This positive feedback can be additionally enhanced if the DAMPs undergo chemical modification by ROS/RNS, and if the modified DAMPs lead to stronger activation of PRR than the native DAMPs, as observed in this study. Such effects may have important consequences for chronic neurodegenerative, cardiovascular or gastrointestinal diseases and allergies that are prevalent in modern societies. We suggest to consider and further elucidate these processes, the role of environmental pollutants, and potential needs and perspectives for societal action with regard to the steeply increasing and globally pervasive human influence on air quality, climate, and public health in the Anthropocene [2,117].

Conflict of interest

The authors declare no conflict of interest.

Author contributions

KL and UP conceived and directed the study. KZ, ATK, KRS, ALL, and KL designed and performed the experiments. KZ, ATK, KRS, MGW, JFN, KL, and UP analyzed, interpreted and discussed the data. DW and DS contributed to the design, interpretation, and discussion. All authors contributed to the preparation and editing of the manuscript. KZ and ATK contributed equally.

Acknowledgement

This work was funded by the Max Planck Society and the Max Planck Graduate Center with the Johannes Gutenberg University Mainz. D. Schuppan receives project related support by research grants from the German Research Foundation DFG Schu 646/17-1 (ATI), DFG Pic/Sch SPP 1656 (Intestinal microbiota), DFG Schu 646/20-1 (Allergy), Collaborative Research Center 128 A08 (Multiple sclerosis), and from the Leibniz Foundation (WheatScan, SAW-2016-DFA-2). The authors acknowledge technical support by N. Bothen, C. S. Krevert, and N.-M. Kropf as well as helpful discussions with the members of the Mainz Program for Chemical Allergology (MPCA).

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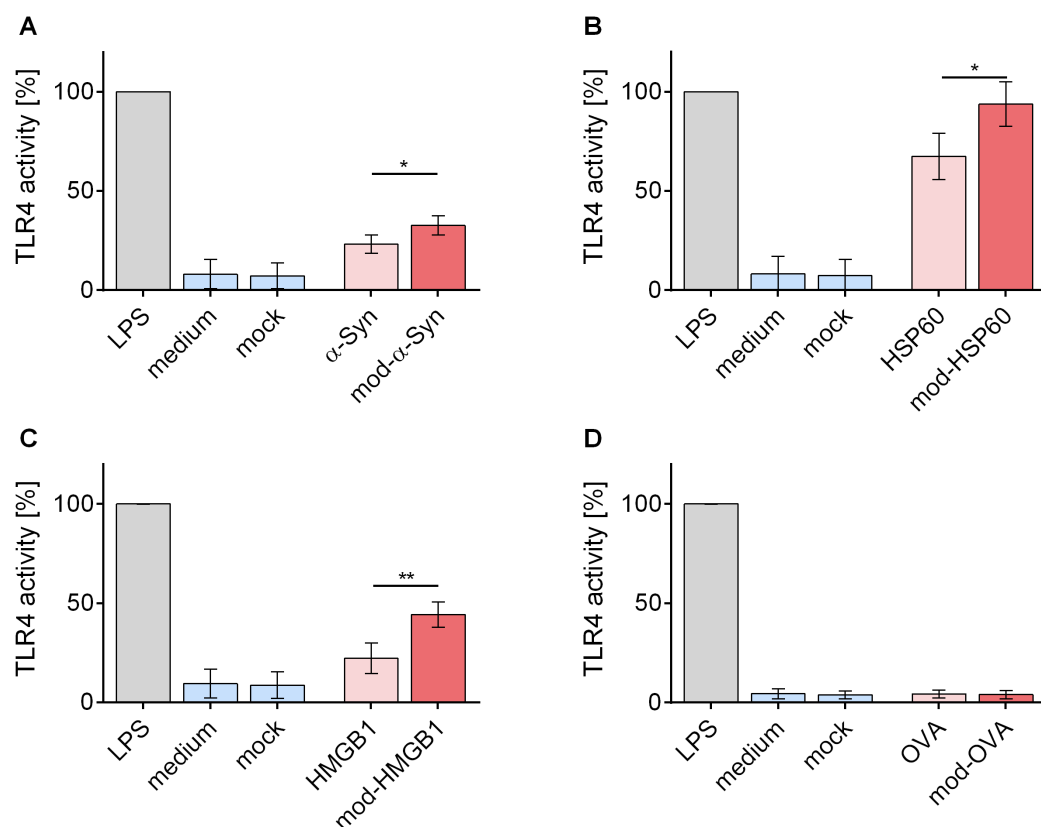


Figure 1: TLR4 activation by native and peroxynitrite-modified proteins. TLR4 activity in HeLa TLR4 dual luciferase reporter cells determined for α -Syn (A), HSP60 (B), HMGB1 (C), and OVA (D) normalized to LPS. Arithmetic mean values and standard deviations of three to four independent experiments performed in triplicates (*p < 0.05, **p < 0.01).

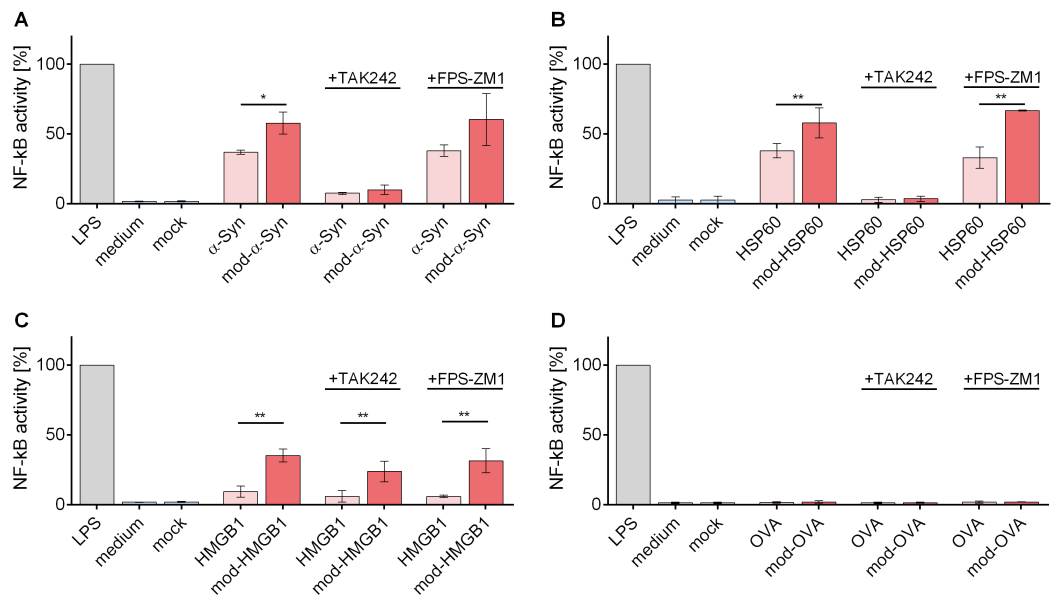


Figure 2: NF-κB activation by native and peroxynitrite-modified proteins. NF-κB activity in THP-1-Lucia™ NF-κB cells determined for α-Syn (A), HSP60 (B), HMGB1 (C), and OVA (D) normalized to LPS. Inhibition experiments with TLR4 antagonist TAK242 and RAGE antagonist FPS-ZM1. Arithmetic mean values and standard deviations of three to four independent experiments performed in triplicates (*p < 0.05, **p < 0.01).

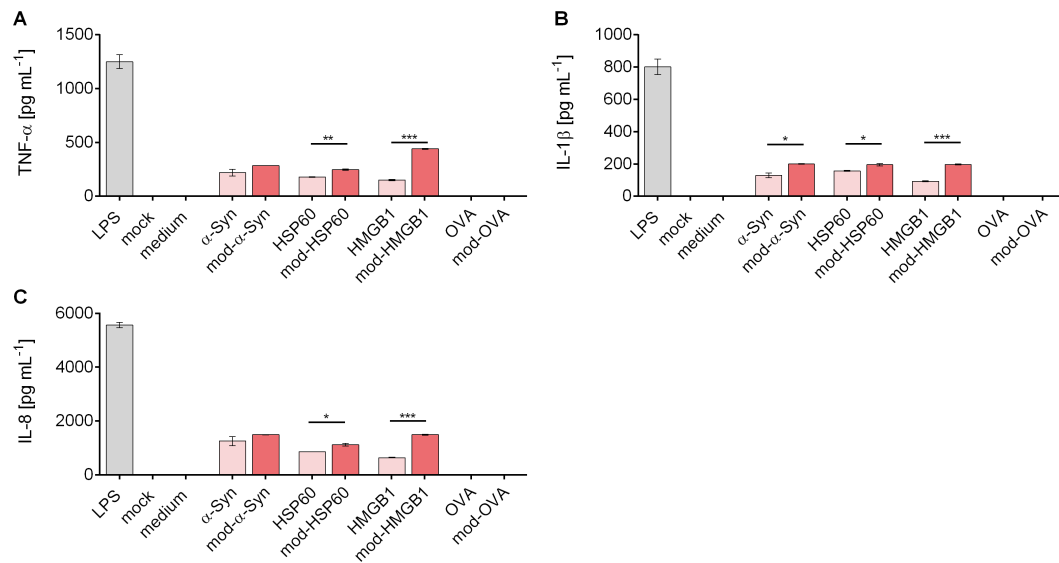


Figure 3: Pro-inflammatory cytokine secretion in response to native and peroxynitrite-modified proteins. Release of TNF-α (A), IL-1β (B), and IL-8 (C) determined for THP-1 monocytes after incubation over 24 h. Data for mock, medium, and OVA/mod-OVA are near or below the relevant limits of detection ($\sim 2\text{--}20\text{ pg mL}^{-1}$). Arithmetic mean values and standard deviations of a representative experiment performed in duplicates (* $p < 0.05$, ** $p < 0.01$, *** $p < 0.005$).

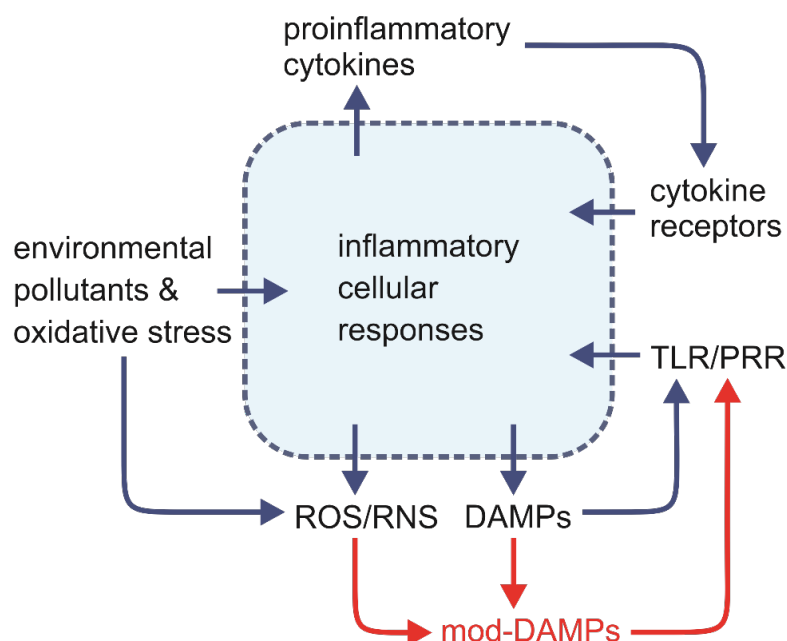


Figure 4: Amplification of inflammatory processes and innate immune responses through chemically modified DAMPs. Environmental pollutants and oxidative stress can induce an increase of reactive oxygen and nitrogen species (ROS/RNS), the formation of chemically modified damage-associated molecular patterns (mod-DAMPs), an increase of pro-inflammatory signaling via Toll-like receptors and other pattern recognition receptors (TLR/PRR), an increase of proinflammatory cytokines, and further inflammatory cellular responses.

Supplementary Material

Chemical modification of pro-inflammatory proteins by peroxynitrite increases activation of TLR4 and NF- κ B: Implications for the health effects of air pollution and oxidative stress

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submitted to Redox Biology (2020)

Supplement to

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Table S1: Overview of investigated proteins: molecular mass, chain length (number of amino acids), number of tyrosine residues, occurrence, and characteristic features.

Protein	Mass [kDa]	Length	Tyrosine residues	Occurrence	Characteristic features
α -Synuclein (α -Syn)	19.5	140	4	Neurons	TLR4-activating DAMP associated with Parkinson's disease
Heat shock protein 60 (HSP60)	60.0	573	5	Most human cell types	Mitochondrial chaperone and TLR4-activating DAMP, many diseases
High-mobility-group box 1 (HMGB1)	25.0	215	7	Most human cell types	Nuclear protein and TLR4-activating DAMP, many diseases
Ovalbumin (OVA)	42.8	386	10	Egg white	Negative control without TLR4 activation

Table S2: Tyrosine nitration degree and relative fractions of monomers and dimers in native and peroxynitrite-modified proteins (arithmetic mean values and standard deviations). Nitration degrees were determined by HPLC for two independently prepared samples measured in duplicates; monomer/dimer fractions were determined by SDS-PAGE for three to four independently prepared samples.

Protein	Nitration degree [%]	Monomer fraction [%]	Dimer fraction [%]
α -Syn	< 1	97 $\pm$ 4	< 3
mod- α -Syn	39 $\pm$ 6	81 $\pm$ 11	16 $\pm$ 7
HSP60	4 $\pm$ 3	98 $\pm$ 1	< 3
mod-HSP60	68 $\pm$ 16	55 $\pm$ 7	23 $\pm$ 8
HMGB1	2 $\pm$ 0.2	92 $\pm$ 9	12 $\pm$ 7
mod-HMGB1	13 $\pm$ 8	87 $\pm$ 11	17 $\pm$ 9
OVA	< 1	84 $\pm$ 4	16 $\pm$ 4
mod-OVA	13 $\pm$ 2	83 $\pm$ 8	15 $\pm$ 5

Table S3: Target genes and sequences of real-time quantitative PCR (qPCR) primers for mRNA expression of cytokines and reference genes (PPIA, TBP).

Gene	Accession number	Sequence 5'→3'
TNF- α	NM_000594.4	fw GCCCAGGCAGTCAGATCATCTT
		rev CCTCAGCTTGAGGGTTTGCTACA
IL-1 β	NM_000576.2	fw GCCCTAAACAGATGAAGTGCTC
		rev GAACCAGCATCTTCCTCAG
IL-8	NM_000584.3	fw AGTCCTTGTTCCACTGTGCCTTGG
		rev TGCTTCCACATGTCCTCACAACATC
CXCL-10	NM_001565.4	fw CTGTACGCTGTACCTGCATCAGCA
		rev ACACGTGGACAAAATTGGCTTGC
PPIA	NM_021130.4	fw TCTGCACTGCCAAGACTGAG
		rev TGGTCTTGCCATTCCTGGAC
TBP	NM_001172085.1	fw TGAGCCAGAGTTATTTCTGGT
		rev AATTTCTGCTCTGACTTTAGCACC

Table S4: Enhancement factors of TLR4 and NF- κ B activity for peroxynitrite-modified protein relative to the native protein. Arithmetic mean values and standard errors for three to four independent experiments performed in triplicates.

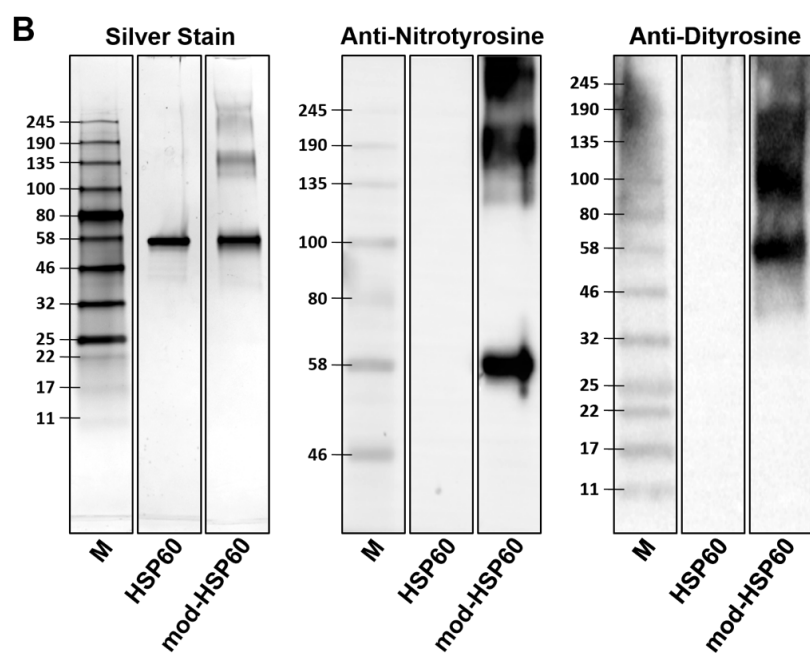
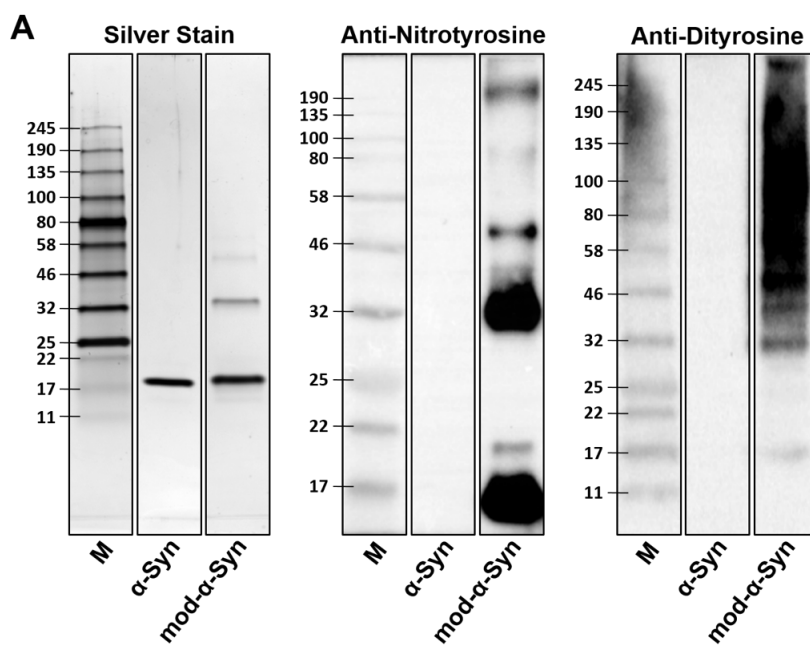
	α-Syn	HSP60	HMGB1
TLR4	1.4 $\pm$ 0.2	1.4 $\pm$ 0.1	2.2 $\pm$ 0.4
NF- κ B	1.6 $\pm$ 0.1	1.6 $\pm$ 0.2	4.2 $\pm$ 0.8

Table S5: Enhancement factors of pro-inflammatory cytokine secretion for peroxynitrite-modified protein relative to the native protein. Arithmetic mean values and standard errors for independent experiments performed in duplicates.

	α-Syn	HSP60	HMGB1
TNF- α	1.3 $\pm$ 0.1	1.4 $\pm$ 0.1	3.0 $\pm$ 0.2
	1.3 $\pm$ 0.2	1.4 $\pm$ 0.04	2.9 $\pm$ 0.1
	2.0 $\pm$ 0.2	8.8 $\pm$ 0.5	16.7 $\pm$ 2.5
IL-1 β	1.5 $\pm$ 0.1	1.3 $\pm$ 0.1	2.2 $\pm$ 0.2
	1.5 $\pm$ 0.2	1.2 $\pm$ 0.1	2.1 $\pm$ 0.1
	1.7 $\pm$ 0.1	4.2 $\pm$ 0.1	6.8 $\pm$ 0.9
IL-8	1.2 $\pm$ 0.1	1.3 $\pm$ 0.1	2.3 $\pm$ 0.2
	1.2 $\pm$ 0.2	1.3 $\pm$ 0.1	2.3 $\pm$ 0.1
	1.6 $\pm$ 0.2	8.8 $\pm$ 0.5	8.6 $\pm$ 1.0

Table S6: Enhancement factors of pro-inflammatory cytokine mRNA expression for peroxynitrite-modified protein relative to the native protein. Arithmetic mean values and standard errors for three to four experiments performed in duplicates.

	α-Syn	HSP60	HMGB1
TNF- α	2.2 $\pm$ 0.1	-	4.4 $\pm$ 0.1
	1.0 $\pm$ 0.1	14.7 $\pm$ 0.4	1.9 $\pm$ 0.3
	1.0 $\pm$ 0.03	1.3 $\pm$ 0.1	2.4 $\pm$ 0.1
	1.0 $\pm$ 0.1	1.2 $\pm$ 0.1	-
IL-1 β	2.1 $\pm$ 0.1	-	4.3 $\pm$ 0.1
	0.9 $\pm$ 0.1	56.3 $\pm$ 3.1	3.3 $\pm$ 0.5
	0.9 $\pm$ 0.03	1.4 $\pm$ 0.04	3.0 $\pm$ 0.2
	1.1 $\pm$ 0.1	1.4 $\pm$ 0.1	-
IL-8	1.2 $\pm$ 0.1	-	3.0 $\pm$ 0.1
	0.8 $\pm$ 0.1	12.9 $\pm$ 0.9	1.7 $\pm$ 0.2
	0.8 $\pm$ 0.03	1.3 $\pm$ 0.1	2.3 $\pm$ 0.2
	1.0 $\pm$ 0.1	1.3 $\pm$ 0.02	-
CXCL-10	1.1 $\pm$ 0.03	-	8.4 $\pm$ 0.5
	1.1 $\pm$ 0.1	8.7 $\pm$ 0.6	1.9 $\pm$ 0.2
	0.9 $\pm$ 0.1	1.3 $\pm$ 0.1	2.4 $\pm$ 0.1
	1.1 $\pm$ 0.1	1.0 $\pm$ 0.03	-



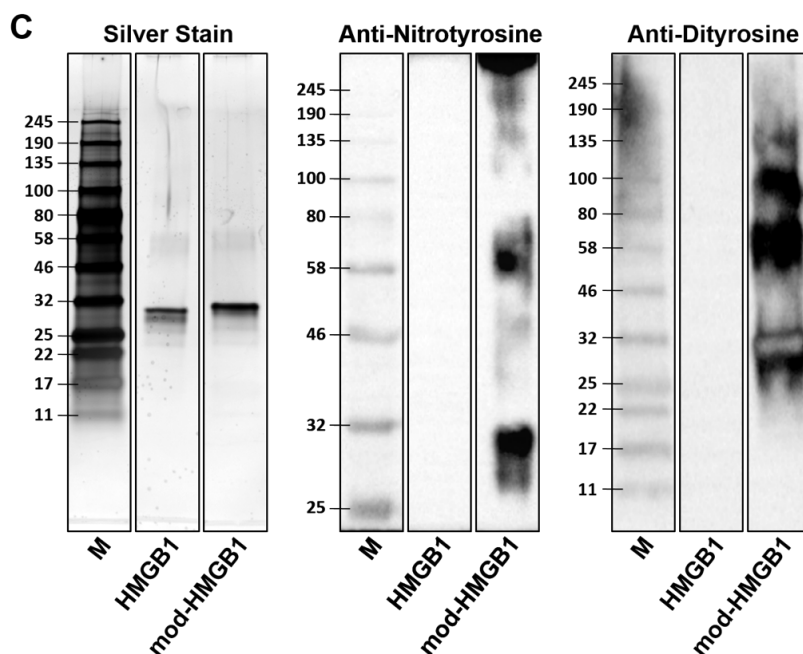


Figure S1: SDS-PAGE with silver stain and western blots against nitro- and dityrosine for native and peroxynitrite-modified proteins. Expected molecular masses of monomers: α -Syn, 19.5 kDa (A), HSP60, 60.0 kDa (B), and HMGB1, 25.0 kDa (C). Silver stain and marker (M) of western blots were taken in bright field and merged with the chemiluminescence western blot pictures against nitro- and dityrosine of native and peroxynitrite-modified proteins.

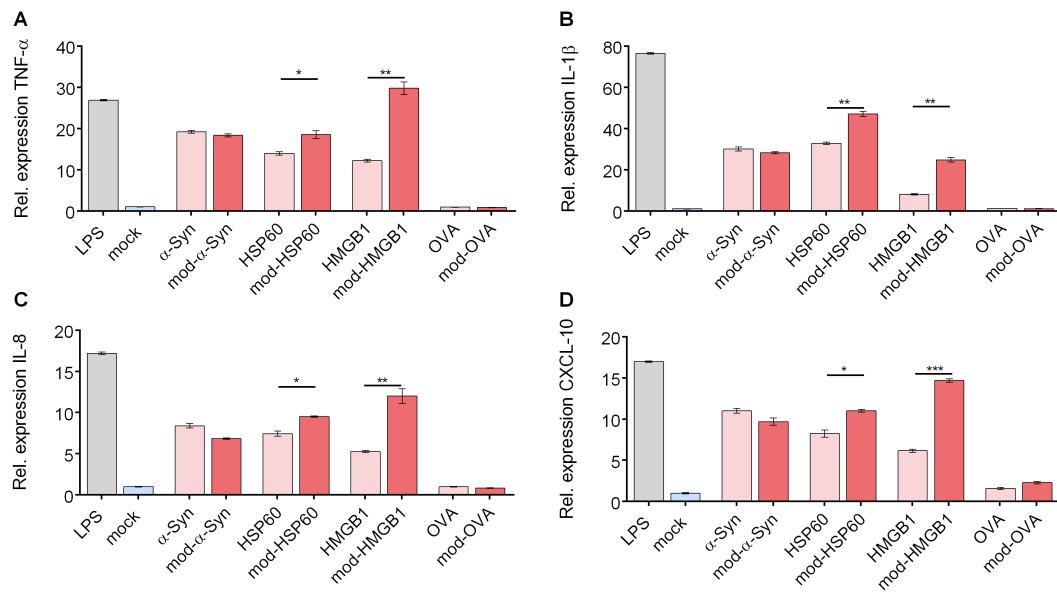


Figure S2: Expression of pro-inflammatory cytokines in response to native and peroxynitrite-modified proteins. mRNA quantification of gene expression of TNF- α (A), IL-1 β (B), IL-8 (C), and CXCL-10 (D) determined for THP-1 monocytes after incubation over 4 h, corrected against two reference genes (PPIA, TBP) and normalized to mock. Arithmetic mean values and standard deviation of a representative experiment performed in duplicates (* $p < 0.05$, ** $p < 0.01$, *** $p < 0.005$).

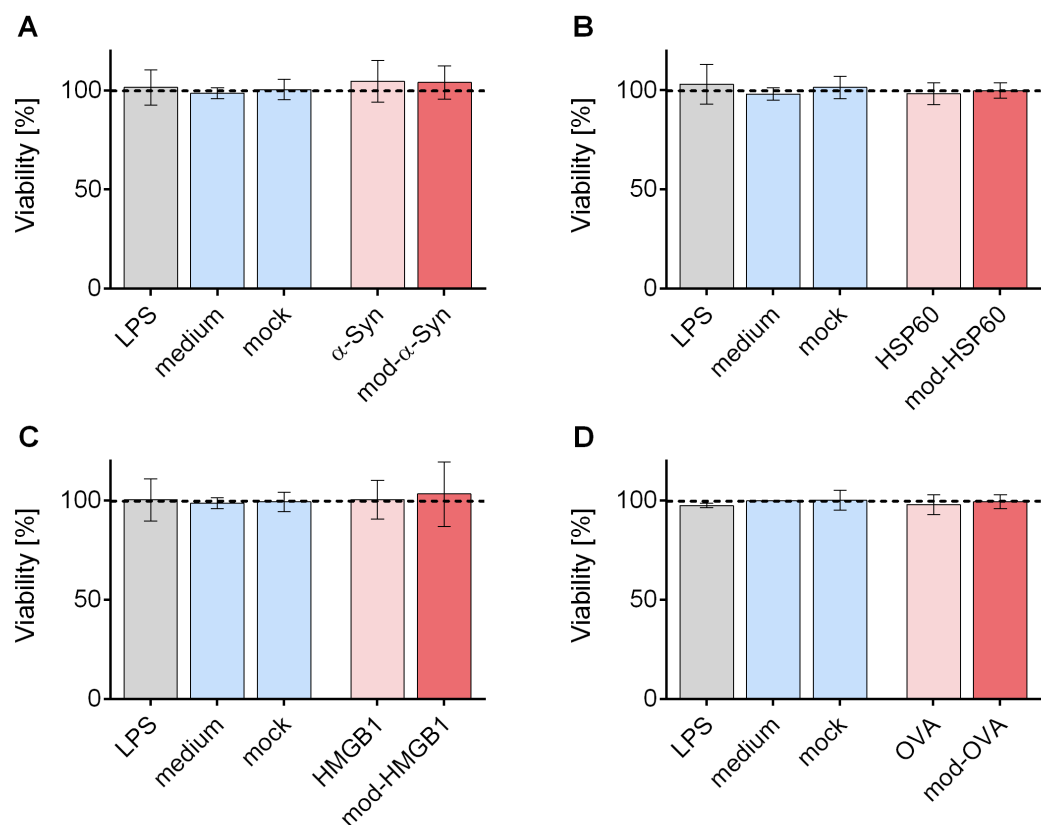


Figure S3: Viability of HeLa TLR4 dual reporter cells treated with native and peroxynitrite-modified proteins. Cell viability assessed for α -Syn (A), HSP60 (B), HMGB1 (C), and OVA (D) using the firefly signal normalized to medium-treated cells. Arithmetic mean values and standard deviations of three to four independent experiments performed in triplicates.

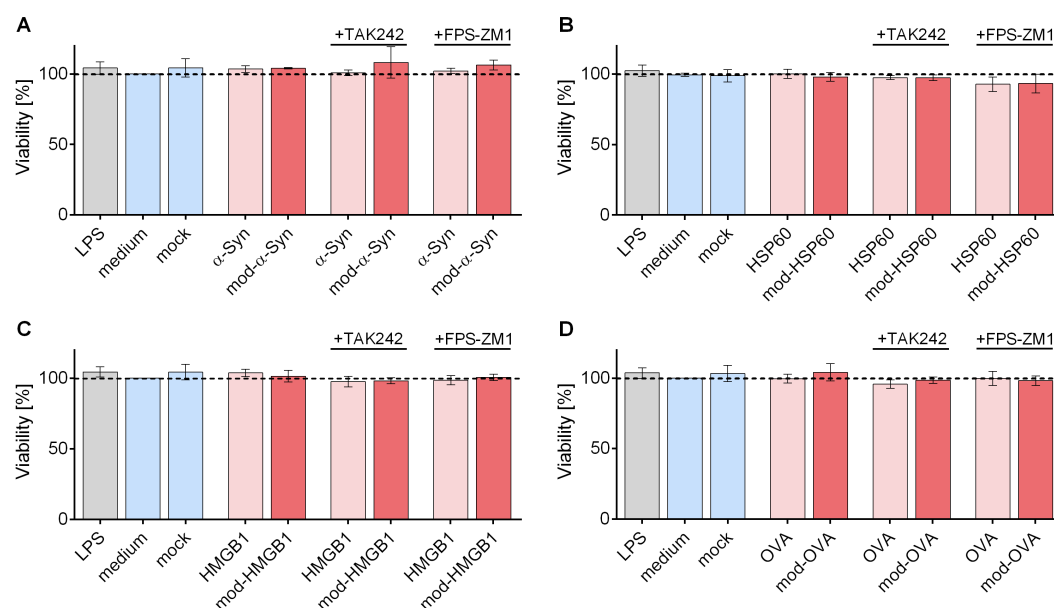


Figure S4: Viability of THP-1-Lucia™ NF-κB reporter cells treated with native and peroxynitrite-modified proteins. Cell viability assessed for α-Syn (A), HSP60 (B), HMGB1 (C), and OVA (D) with and without TLR4 or RAGE antagonist using the alamarBlue™ cell viability reagent normalized to medium-treated cells. Arithmetic mean values and standard deviations of three to four independent experiments performed in triplicates.

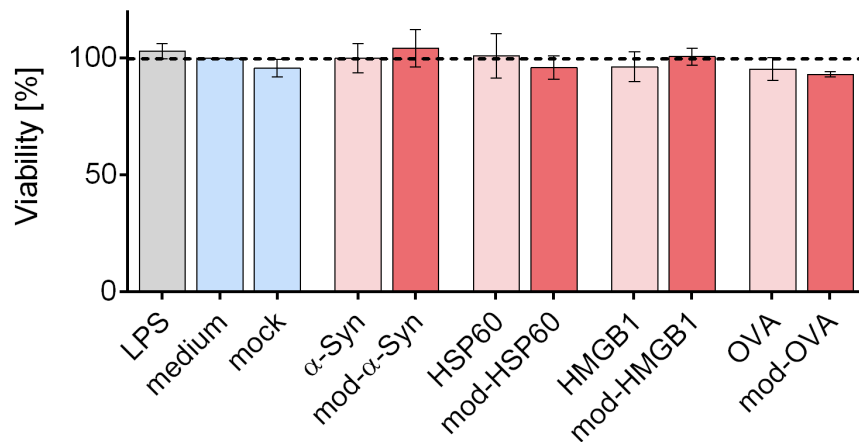


Figure S5: Viability of human THP-1 monocytes treated with native and peroxynitrite-modified proteins. Cell viability assessed for α -Syn, HSP60, HMGB1, and OVA using the alamarBlue™ cell viability reagent normalized to medium-treated cells. Arithmetic mean values and standard deviations of three independent experiments performed in triplicates.

Air Pollution and Climate Change Effects on Allergies in the Anthropocene: Abundance, Interaction, and Modification of Allergens and Adjuvants

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Environmental Science and Technology **51**, 4119-4141 (2017)

Online supplement available at:

https://pubs.acs.org/doi/suppl/10.1021/acs.est.6b04908/suppl_file/es6b04908_si_001

Contribution

Aided in writing the manuscript.



Air Pollution and Climate Change Effects on Allergies in the Anthropocene: Abundance, Interaction, and Modification of Allergens and Adjuvants

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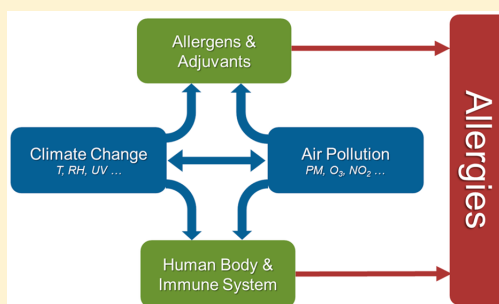
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Supporting Information

ABSTRACT: Air pollution and climate change are potential drivers for the increasing burden of allergic diseases. The molecular mechanisms by which air pollutants and climate parameters may influence allergic diseases, however, are complex and elusive. This article provides an overview of physical, chemical and biological interactions between air pollution, climate change, allergens, adjuvants and the immune system, addressing how these interactions may promote the development of allergies. We reviewed and synthesized key findings from atmospheric, climate, and biomedical research. The current state of knowledge, open questions, and future research perspectives are outlined and discussed. The Anthropocene, as the present era of globally pervasive anthropogenic influence on planet Earth and, thus, on the human environment, is characterized by a strong increase of carbon dioxide, ozone, nitrogen oxides, and combustion- or traffic-related particulate matter in the atmosphere. These environmental factors can enhance the abundance and induce chemical modifications of allergens, increase oxidative stress in the human body, and skew the immune system toward allergic reactions. In particular, air pollutants can act as adjuvants and alter the immunogenicity of allergenic proteins, while climate change affects the atmospheric abundance and human exposure to bioaerosols and aeroallergens. To fully understand and effectively mitigate the adverse effects of air pollution and climate change on allergic diseases, several challenges remain to be resolved. Among these are the identification and quantification of immunochemical reaction pathways involving allergens and adjuvants under relevant environmental and physiological conditions.



1. INTRODUCTION AND MOTIVATION

Allergies are hypersensitivities initiated by specific immunologic mechanisms (abnormal adaptive immune responses).^{1–3} They constitute a major health issue in most modern societies, and related diseases, such as allergic rhinitis, atopic asthma, eczema (atopic dermatitis), and food allergies, have strongly increased during the past decades.^{4–12} While some of the perceived rise

in allergies may be due to improved diagnosis, the prevalence of allergic diseases has genuinely increased with industrialization

Received: October 17, 2016

Revised: March 7, 2017

Accepted: March 22, 2017

Published: March 22, 2017



and with the adoption of a “Western” lifestyle.¹³ The development of allergies is a complex multifactorial process that involves various factors influencing the body’s predisposition and immune response, and the manifestation of allergic diseases depends on exposure to allergens, adjuvants and other environmental and lifestyle factors (Figure S1 and section S1).^{3,4,14–16} Among the risk factors for allergic diseases are the genetic predisposition of the individual (referred to as atopy), reduced childhood exposure to pathogens and parasites (“hygiene hypothesis”), diet/nutrition, psychological/social stress, and environmental pollution, including outdoor and indoor air pollutants (ozone, nitrogen oxides, diesel exhaust particles, tobacco smoke, etc.).^{4,12,17–35} As outlined in Figure 1,

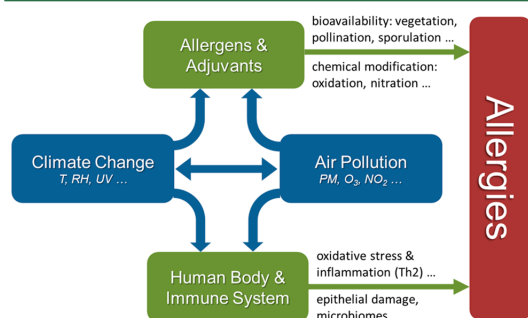


Figure 1. Interplay of air pollution and climate change can promote allergies by influencing the human body and immune system, as well as the abundance and potency of environmental allergens and adjuvants.

climate change and air pollution can influence the bioavailability and potency of allergens and adjuvants in multiple ways, including changes in vegetation cover, pollination and sporulation periods, and chemical modifications. Moreover, climatic conditions and air pollutants may skew physiological processes and the immune system toward the development of allergies, for example, by oxidative stress and inflammation, disruption of protective epithelial barriers, and disturbance of related microbial communities (microbiomes).^{4,8,35–38}

The term Anthropocene describes the present era of globally pervasive and steeply increasing anthropogenic/human influence on planet Earth, including the land surface, biosphere and atmosphere.^{38–44} Human activities have become a driving force that changes many characteristics of our environment such as biodiversity and air quality on local, regional, and global scales, for example, through land use change, agriculture, fossil fuel burning, traffic emissions, and the release of industrial products.^{38,39,41,43,45–49} While the basic concept of the Anthropocene, as introduced by Nobel laureate Paul J. Crutzen and colleagues,^{39,44,50} is widely accepted and increasingly used across the sciences and humanities, the actual beginning of the Anthropocene as a new geological epoch is still under investigation and discussion.^{38,45–47,51–64} The proposed dates range from early human history via the 19th century (industrialization) to the 1960s (nuclear weapon testing and “Great Acceleration”).^{45–47,58–64} Since the industrialization of the 19th century and especially during the “Great Acceleration” of the 20th century, the primary emission, secondary formation, and concentration of air pollutants like ozone, nitrogen, and sulfur oxides, soot, and a wide range of other reactive trace gases and aerosols have greatly increased relative to

preindustrial times, especially in densely populated and industrialized areas but also in agricultural environments and around the globe.^{38,47,65–69} For example, the average mixing ratios of ozone in continental background air have increased by factors of 2–4 from around 10–20 ppb from the beginning of the 19th century to 30–40 ppb in the 21st century, and the number and mass concentrations of aerosol particles in polluted urban air are typically by 1–2 orders of magnitude higher than in pristine air of remote continental regions ($\sim 10^2$ – 10^3 cm^{-3} and ~ 1 – 10 $\mu\text{g m}^{-3}$ vs $\sim 10^3$ – 10^5 cm^{-3} and ~ 10 – 100 $\mu\text{g m}^{-3}$).^{38,70}

Numerous studies indicate that ozone and air particulate matter have strong effects on human health and mortality as well as on agricultural crop yields.^{71–80} In view of these findings, it appears unlikely that the strong environmental changes of the Anthropocene would have no effect on the interaction of the human immune system with environmental stimuli, including allergens and adjuvants. Indeed, it seems necessary to address the question whether human activities are creating a hazardous atmosphere that may severely affect public health.^{35,37,38,81,82} Figure 2 illustrates how climate parameters and air pollutants can exert proinflammatory and immunomodulatory effects.⁸ As detailed in the following sections, both air pollutants and climate parameters can influence the environmental abundance of allergenic bioparticles and the release of allergenic proteins and biogenic adjuvants. Moreover, air pollutants can chemically modify and agglomerate allergenic proteins, and they can act as adjuvants inducing epithelial damage and inflammation.

Several reviews have addressed the general determinants of allergenicity^{3–8,83–85} and various environmental risk factors for allergic diseases.^{4,9,12,34,36,86–101} In this Critical Review, we attempt to summarize, update, and synthesize the different perspectives and most relevant findings reported in earlier reviews and recent research articles addressing the effects of air pollutants and climate parameters on allergies. A central aim of this article is to review and outline both proven and potential effects of the globally pervasive environmental changes that are characteristic for the Anthropocene; a holistic view of environmentally caused changes in the abundance, interaction, and modification of allergens and related substances is provided. Our target audience comprises physical, chemical, and biomedical scientists interested in environmental effects on public health. Sections 2–4 deal with specific environmental processes and air pollutants that are likely to affect the development of allergies in the Anthropocene, that is, in an environment strongly influenced by human activity. Section 5 provides an outlook identifying key questions and promising directions of future research. For orientation of readers not familiar with the basics of allergic sensitization and response, section S1 outlines key features of the immunochemical interactions involved in IgE-mediated allergies (type I hypersensitivities)^{3–5,14–16,84,102–136} on which this article is mainly focused and which usually involve Th2 cell-mediated inflammation^{137,138} (Figure S2).

2. ABUNDANCE AND RELEASE OF ALLERGENS AND ADJUVANTS

Environmental allergens are mostly proteins derived from plants, animals, and fungi that can trigger chemical and biological reaction cascades in the immune system leading to allergic sensitization and formation of IgE antibodies (section S1).^{8,84,103,105–109} Prominent examples are major allergens of

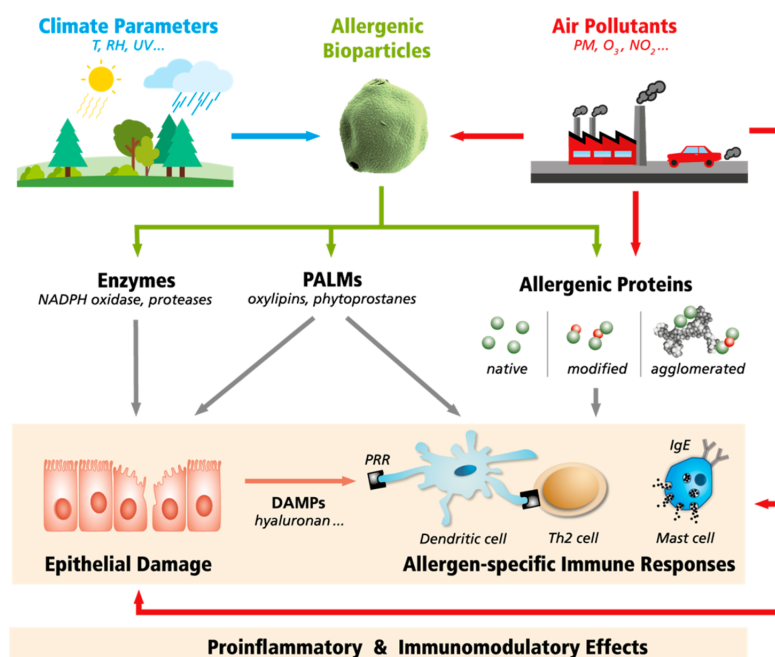


Figure 2. Pathways through which climate parameters and air pollutants can influence the release, potency, and effects of allergens and adjuvants: temperature (T), relative humidity (RH), ultraviolet (UV) radiation, particulate matter (PM), ozone and nitrogen oxides (O_3 , NO_x), reduced nicotinamide adenine dinucleotide phosphate (NADPH) oxidase, pollen-associated lipid mediators (PALMs), damage-associated molecular patterns (DAMPs), pattern recognition receptors (PRR), type 2 T helper (Th2) cells, immunoglobulin E (IgE), allergenic proteins (green dots), and chemical modifications (red dots).

birch pollen (Bet v 1), timothy grass pollen (Phl p 1), ragweed (*Ambrosia*, Amb a 1), molds (*Alternaria alternata*, Alt a 1, *Cladosporium herbarum*, Cla h 1, *Aspergillus fumigatus*, Asp f 1), and dust mites (Der p 1).^{4,139,140} Besides allergens, also adjuvants and their interaction with the immune system play a critical role in the development of allergies. Here, we use the term adjuvant generically for substances that are promoting pro-allergic immune responses. Adjuvants can trigger the immune system by inducing tissue damage and subsequent enhanced uptake of allergens, by inducing oxidative stress and activation of immune cells, by coexposure with the allergen favoring Th2 responses, or by modification of allergens enhancing their allergic potential. An overview of biogenic and anthropogenic adjuvants, including particulate matter as well as trace gases, and their effects on the immune system is given in Table 1.

Climate change is influencing vegetation patterns and plant physiology through spatial and temporal changes in temperature and humidity (Figure 1),^{141–143} and increasing atmospheric carbon dioxide (CO_2) affects plant biology by supplying more carbon for photosynthesis, biomass production, and growth (CO_2 fertilization).^{144,145} These factors can influence the spread of invasive plants, the beginning, duration, and intensity of pollination, the fruiting patterns and sporulation of fungi, as well as the allergen content and allergenicity of pollen grains, fungal spores, and other biological aerosol particles (Figure 2).^{12,90,93,96–98,145–162} Specific examples of climate change effects on allergenic plants and fungi are outlined in Table 2. Climate and land use change are also expected to influence the composition and spread of microbial

surface communities (cryptogamic covers), from which allergenic cyanobacteria and other microbial allergens or adjuvants can be emitted to the atmosphere.^{163–174} Moreover, the frequency and intensity of dust storms are expected to increase,^{141,175–179} and dust particles are known to carry biological and organic components with pathogenic, allergenic, and adjuvant activity.^{152,154,180–187} Dust storms have been shown to cause and aggravate respiratory disorders including atopic asthma and allergic rhinitis.^{181,188–191} So-called “thunderstorm asthma” is characterized by acute asthma exacerbations possibly caused by the dispersion of inhalable allergenic particles derived from plant pollen and fungal spores by osmotic rupture.^{145,192} On the other hand, climate change-related regional enhancements of outdoor humidity and indoor home dampness may also lead to an increase of respiratory symptoms and atopic asthma induced by allergenic and adjuvant substances from fungi, other microbes, and mite.^{12,193–196}

Pollen grains generally belong to the coarse fraction of air particulate matter (particle diameters $>10 \mu m$), but fungal spores and pollen fragments are also found in fine particulate matter ($<2.5 \mu m$; $PM_{2.5}$), which can penetrate deep into the human respiratory tract and alveolar regions of the lung.^{152,153,197–203} Allergenic proteins can be released from pollen and spores after cell damage or under humid conditions.²⁰⁴ In particular, pollen rupture due to an osmotic shock during rain can lead to outbreaks of thunderstorm asthma.^{145,192,205,206} Furthermore, peaks of high concentrations of pollen, fungal spores, and other primary biological aerosol (PBA) particles have been observed at the onset of heavy rain

Table 1. Biogenic and Anthropogenic Adjuvants with Reported Pro-allergic Effects: (I) Pollen-Associated and Microbial Compounds, Such as Pollen-Associated Lipid Mediators (PALMs), Bacterial Lipopolysaccharides (LPS), and Fungal β -Glucans and (II) Anthropogenic Pollutants and Chemicals Including Air Particulate Matter, Gaseous Oxidants, and Organic Compounds

substances	effects
(I) Pollen-Associated and Microbial Compounds	
proteases	disrupt intracellular adhesion; stimulate protease activated receptors (PAR) inducing inflammation and enhanced IgE production ^{204,447,448} fungal proteases activate TLR4 ⁴⁴⁹
leukotrien-like PALMs	attract and activate innate cells like neutrophils and eosinophils ⁴⁵⁰
phytoprostane PALMs	inhibit IL12 production and enhance IgE production ¹⁰⁷
NADPH oxidase	ROS production and inflammation ⁴⁵¹
adenosine	Th2 cytokine profile and inflammation ⁴⁵²
flavonoids	modulate immune responses as ligands of allergenic proteins, e.g., a natural ligand of Bet v 1 is a quercetin and binds to the C-terminal helix ^{372,453} the pollen-derived flavonoid isorhamnetin modulates the immunological barrier function of the epithelium ⁴⁵⁴
bacterial LPS	trigger TLR4 in dose dependent manner, induce a Th2 bias and allergic inflammation ⁴⁵⁵
gram-positive bacteria	induce DC maturation by upregulation of CD80, CD83, and CD86 ⁴⁴⁵
fungal β -glucans	activate C-type lectin receptor ¹⁰⁵
fungal VOC	stimulate inflammatory response ⁴⁵⁶
(II) Anthropogenic Pollutants and Chemicals	
air particulate matter (PM)	diesel exhaust particles (DEP) increase Th2 sensitization to coinhaled allergens (IgE isotype switching and production, mast cell and basophil degranulation, cytokine production (e.g., IL-4); exacerbates allergic airway responses ^{86,457–462} PM and DEP induce ROS production and inflammation ^{86,463–465} DEP suppress alveolar macrophage function ^{466,467} DEP and cigarette smoke can increase thymic stromal lymphopoietin (TSLP) expression in epithelial cells ^{468,469} DEP induce permeability of epithelial cells; disrupt tight junctions by a ROS-mediated pathway ^{470,471} PM increase the expression of costimulatory molecules on DCs (MHC class II, CD40, CD80, CD86) ^{86,469} ultrafine particles (UFP < 100 nm) and DEP alter soluble protein levels (e.g., surfactant protein D, complement protein C3), increase levels of e.g., glycerin-aldehyde-3-phosphate-dehydrogenase (GADPH), manganese superoxide dismutase (MnSOD), or mitochondrial heat shock protein (Hsp 90) ^{472,473} PM2.5 and DEP activate complement proteins (C3) ^{474,475} black carbon (BC) and DEP induce epigenetic effects: DNA methylation in genes associated with Th2 polarization ^{476–478} DEP and cigarette smoke induce epithelial damage, oxidative stress, and inflammation ⁴⁶⁰ prenatal and postnatal exposure to environmental tobacco smoke (EST) is associated with asthma and wheezing ^{34,479,480} transition metals and other redox-active compounds (organic peroxides, quinones) induce ROS production and inflammation via Fenton-like reactions ^{38,129,309,311,481–483} colocalization of allergens on gold nanoparticles can facilitate IgE-receptor cross-linking ²⁴⁴ cause oxidative stress, airway inflammation, increased airway permeability ^{329,362,368,484} formation of protein ROI (reactive oxygen intermediates) and protein dimers ^{329,362} elevated levels of complement protein C3a ⁴⁸⁵ degradation of high molecular weight to low molecular weight hyaluronan, which is a DAMP that activates the TLR4 pathway ^{407,486}
ozone (O ₃)	nitration of allergens ^{328,329} skew towards Th2 response, ⁴⁸⁷ increase eosinophilic inflammation, ⁴⁸⁸ and enhance airway permeability ⁴⁸⁴
nitrogen oxides (NO _x = NO + NO ₂)	significant positive association between formaldehyde exposure and childhood asthma ²⁷²
volatile, semivolatile and low-volatile organic compounds (VOC, SVOC, LVOC)	antimicrobial endocrine disrupting compounds such as parabens and triclosan are associated with allergic sensitization ^{489,490} Bisphenol A can increase IL-4 and IgE levels ⁴⁹¹ dermal and pulmonary exposure to indoor VOC elicit irritant and allergic responses ^{270,271}

and moist weather conditions,^{200,207,208} and increased concentrations of free allergen molecules in fine air particulate matter have been observed after rainfall.²⁰⁹ Prominent airborne fungi, such as *Cladosporium herbarum*, *Alternaria alternata*, and *Aspergillus fumigatus*, have been found to release higher amounts of allergens after germination under humid conditions,²¹⁰ and certain allergens are expressed only following germination.^{210,211} Air pollutants, such as ozone, nitrogen oxides, and acids, can also interact with PBA particles, damage their envelope, and facilitate the release of allergenic substances, such as cytoplasmic granules from pollen (Figure S3).^{205,212,213}

Besides allergenic proteins, pollen and fungal spores also release other compounds that can act as adjuvants (Table 1). In particular, the release of nonallergenic, bioactive, pollen-associated lipid mediators (PALMs) with pro-inflammatory

and immunomodulatory effects can trigger and enhance allergies (Figure 2).^{8,109,214–217} For example, skin prick tests of pollen allergens elicited larger wheals when tested together with low molecular weight compounds extracted from pollen.²¹⁸ The release of these substances can be influenced by climatic conditions and air pollution, and significantly higher levels were found for pollen collected near roads with heavy traffic.²⁰⁵ Leukotriene-like PALMs (oxylipins) have the potential to attract and activate innate immune cells like neutrophils and eosinophils.^{214,217} Other PALMs such as phytoprostanes (lipophilic counterparts of prostaglandins) are water-soluble and can inhibit the production of interleukin 12 (IL-12) by dendritic cells in the lower respiratory tract, thus favoring an allergenic Th2 T cell response.^{8,215} A recent study showed that the low-molecular-weight fraction of phytopros-

Table 2. Climate Change Effects on the Abundance and Properties as Reported for Selected Plants and Fungi Emitting Aeroallergens

allergenic species	effect of increasing temperature and CO ₂ concentration
<i>Ambrosia artemisiifolia</i> (ragweed)	increased pollen and allergen production, plant migration and spreading ^{157,492–495} changes in pollen transcriptome, changes in allergenic potential, increase in flavonoid metabolites ¹⁵⁸
<i>Betula</i> spp. (birch)	earlier pollination start, increased pollen production ^{161,267,496}
<i>Phleum pratense</i> L. (timothy grass)	increased pollen production ¹⁵⁹
<i>Alternaria</i> spp. (mold)	increased spore numbers, decreased allergen per spore ^{146,156,160}
<i>Aspergillus fumigatus</i> (mold)	modified allergenicity and Asp f 1 content, increased spore numbers ^{146,155,497}
<i>Cladosporium</i> spp. (mold)	increased spore numbers ¹⁴⁶
<i>Penicillium</i> spp. (mold)	increased spore numbers ¹⁴⁶

tane E1 (PPE1) in ragweed pollen extract specifically enhanced IgE production in Th2 primed B cells. It was suggested that pollen-derived nonallergenic substances might be responsible for aggravation of IgE-mediated allergies.²¹⁹

Fine aerosol particles and a wide range of inorganic, organic and biological substances from both natural and anthropogenic sources (e.g., secondary organic material, sulfuric and nitric acid, microbial compounds) can agglomerate and accumulate on the surface of pollen, fungal spores, and other PBA particles as illustrated in Figure S3.^{152,205,220–223} An overview of reported air pollutant effects on the allergenic potential of plant pollen and fungal spores is given in Table S1.^{38,205,221,224–240} Moreover, free allergens and adjuvants can bind to particulate pollutants, such as dust, soot, black/elemental carbon (BC/EC), and diesel exhaust particles (DEP) carrying the allergens and adjuvants into peripheral and deep

airways.^{241–243} The colocalization of allergens and adjuvants on particle surfaces (sorption layers, protein coronas) might also promote allergic sensitization and response by providing multiple/multivalent epitopes that facilitate receptor cross-linking (similar to parasitic organisms, against which IgE is naturally deployed).^{244,245}

During recent years, great progress has been made in the development and application of efficient sampling and measurement methods for bioaerosol particles and components, including microscopic, spectroscopic, mass spectrometric, genomic, and proteomic analyses.^{152,246–253} These and related advances in measurement and modeling techniques for health and climate relevant air contaminants (aerosols and trace gases) are expected to enable comprehensive characterization and forecasting of allergenic and adjuvant substances, as well as their mixing state in outdoor and indoor air.^{38,70,254–268} Note that indoor air quality is usually influenced by both outdoor air pollutants (O₃, NO_x, PM_{2.5}, etc.) and additional pollutants from indoor sources (e.g., formaldehyde and other organic compounds).^{35,37,265,269–274} The data from ambient and individual monitoring and modeling of aeroallergen and adjuvant exposure can then be applied in epidemiological studies to better understand the risk factors of allergic sensitization and disease.^{74–76,275–280}

Several epidemiological studies and meta-analyses reported that respiratory allergies and atopic dermatitis are associated with exposure to traffic-related air pollution (TRAP), but different results were obtained for different diseases and locations/studies.^{281–293} TRAP is a complex mixture comprising variable proportions of particulate matter and gaseous pollutants from traffic-related primary emissions, as well as secondary pollutants formed by chemical reactions in the atmosphere.²⁸³ Among the pollutants from primary emissions (combustion and noncombustion sources) are road dust, tire and brake wear, soot/DEP, BC/EC, metals, polycyclic aromatic

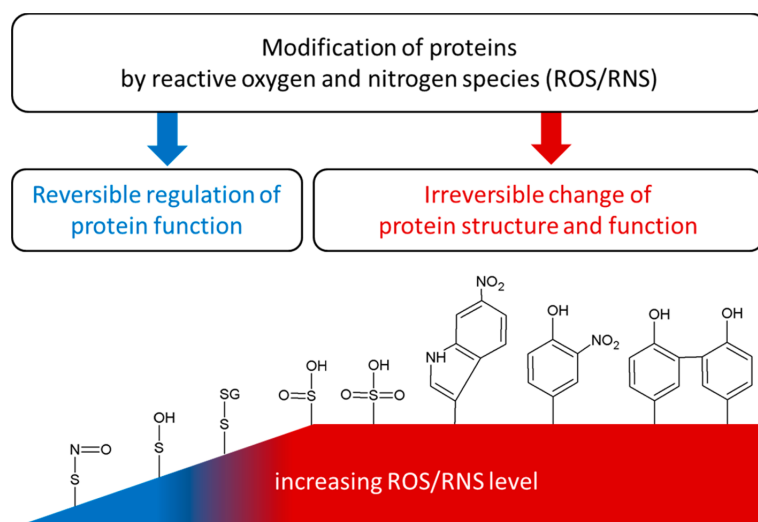


Figure 3. Upon interaction with reactive oxygen and nitrogen species (ROS/RNS), proteins can undergo a wide range of reversible and irreversible chemical modifications. Among the most commonly formed functional groups and products are S-nitrosothiol (SNO), sulfinic acid (SO₂H), disulfides with protein thiols or low molecular mass thiols (e.g., with glutathione, SSG), sulfonic acid (SO₃H), nitrotyrosine, and dityrosine. Adapted from ref 317. Copyright 2013 American Chemical Society.

hydrocarbons (PAH), and nitrogen oxides (NO_x); among the secondary pollutants are ozone, nitrates, and secondary organic aerosols (SOA).^{38,70,273,283} A recent review concluded that epidemiological studies were restricted by imprecise methods of assessing both TRAP exposure and related health effects.²⁸³ Accordingly, several studies called for more comprehensive investigations of TRAP markers, personal exposure, and lifetime outcomes.^{281,294,295} The application of improved measurement and modeling techniques as outlined above should enable refined epidemiological studies and more targeted testing of hypotheses by resolving different types of TRAP (e.g., freshly emitted DEP vs resuspended road dust; soot and polycyclic hydrocarbons vs trace metals; ozone vs nitrogen oxides; etc.).

3. CHEMICAL MODIFICATION OF PROTEINS AND AMINO ACIDS

Chemical modification by air pollutants can lead to changes in the structure of protein macromolecules (amino acid oxidation, peptide backbone cleavage, conformational changes, cross-linking, and oligomerization), and affect protein stability and other properties, such as hydrophobicity and acidity of binding sites.^{296–303} These and other posttranslational protein modifications may induce multiple effects in the molecular interaction of allergens with the immune system: (1) stability effects influencing the accumulation and degradation of allergenic proteins, the duration of exposure times to cellular receptors, and the process of antigen presentation via major histocompatibility complex (MHC) class II;^{304,305} (2) epitope effects, that is, generation of new epitopes or modification of existing epitopes, changing the binding properties of antibodies and receptors, by direct chemical modification or as a result of conformational changes; (3) adjuvant effects, that is, generation of new adjuvant functions or modification of existing adjuvant functions such as lipid-binding capacities due to modified ligand binding sites; and (4) agglomeration effects, that is, multiplication or shielding of epitopes or adjuvant functions by cross-linking (oligomerization) of allergenic proteins, which may enhance the cross-linking of effector cell receptors (FcεRI) or sterically hinder molecular and cellular interactions.^{307,308,306,229}

In the atmosphere, reactive oxygen and nitrogen species (ROS/RNS) are generated via photochemistry and gas-phase, heterogeneous, and multiphase reactions involving atmospheric oxidants and aerosol particles. In the human body, ROS/RNS can be formed upon exposure to air pollutants^{38,309–312} or radiation (UV, X-rays, γ -rays),³¹³ and by regular physiological reactions.³¹⁴ For example, ROS/RNS are generated during oxidative metabolism as well as in cellular responses to foreign or danger signals (cytokines, xenobiotics, bacterial invasion).³¹⁵ Low amounts of ROS/RNS are involved in intra- and intercellular redox signaling processes, for example, oxidizing low molecular mass thiols and protein thiols (Figure 3).^{316,317} An imbalance between oxidants and antioxidants in favor of oxidants (e.g., induced by air pollutants) can lead to irreversible damage of cellular lipids, proteins, nucleic acids, and carbohydrates, eventually resulting in cell death.^{38,317,318} Rising levels of atmospheric oxidants and air particulate matter may lead to protein modifications in the atmosphere, as well as in the human body because of elevated oxidative stress levels, especially in the epithelial lining fluid (section 4).³⁸ Moreover, air pollutants and climatic stress factors, such as UV radiation, drought, salinity, and temperature extremes, can also induce

higher ROS/RNS levels inside plants, which may lead to chemical modification of plant proteins, including allergens.^{38,142,143} In the course of the Anthropocene, the ambient concentrations of many ROS/RNS have strongly increased because of emissions from traffic and combustion sources, as well as other industrial and agricultural activities like nitrogen fertilization of soils.^{37,38,82,319,320}

In the following, we focus on irreversible modifications of allergenic proteins, such as oxidation, nitration, and cross-linking (Figure 3) by endogenous and exogenous ROS and RNS, like ozone (O_3), hydroxyl radicals (OH), hydrogen peroxide (H_2O_2), superoxide anion (O_2^-), nitric oxide (NO), nitrogen dioxide (NO_2), nitrous acid (HONO), nitric acid (HNO_3), peroxyacetylnitrate (PAN), peroxyxynitrite (ONOO^-), and nitrate radicals (NO_3). ROS and RNS can react with oxidation-sensitive amino acids, such as cysteine (Cys), methionine (Met), tryptophan (Trp), tyrosine (Tyr), phenylalanine (Phe), and histidine (His), as well as with aliphatic side chains and the peptide backbone.^{317,321–324} For example, OH radicals can cause backbone cleavage by abstracting hydrogen atoms from the α -carbon of any amino acid in the polypeptide backbone. Subsequent reactions lead to oxidative degradation of the protein and the formation of amide and carbonyl groups.^{321,325,326} Oxidation reactions can result in aggregation, fragmentation, and denaturation of proteins.^{327–329} While oxidative degradation appears likely to reduce the recognition of allergenic proteins, other chemical modifications, such as nitration or cross-linking may enhance the potency of allergens.^{8,229,306–308,328,330–332}

The reaction of proteins with nitrating agents leads mainly to the nitration of the aromatic amino acid tyrosine forming 3-nitrotyrosine (NTyr).³³³ The addition of the rather bulky NO_2 group at the ortho position of the aromatic ring induces a significant shift in the pK_a value of the tyrosine residue (Tyr) from ~ 10 to ~ 7 , thus increasing the acidity of the hydroxyl group. These structural and chemical changes of the amino acid can affect the conformation and function of proteins.^{334,335} For example, the modification of tyrosine residues can influence cell signaling through the important role of receptor tyrosine kinases, which are key regulators of cellular processes.³³⁶ Moreover, nitrotyrosine has been reported as a biomarker for oxidative stress, inflammation, and a wide range of diseases.^{296,301,337,338}

Early immunological studies already suggested that dinitrophenyl derivatives of proteins and peptides evade immune tolerance and boost immune responses.^{339,340} As early as 1934, the allergic reaction to dinitrophenol was described,³⁴¹ and dinitrophenyl haptens became very popular reagents for the experimental induction of allergies.^{342–344} Thus, nitrated aromatics and especially nitrophenols can be considered corner stones in the field of allergy research, suggesting that protein nitration by air pollutants might play a role in the development of allergies.³³⁰

Indeed, several studies showed enhanced allergenic potentials for nitrated pollen allergens,^{229,305,306} nitrated fungal allergens,²³⁷ and nitrated food allergens.^{304,345} For example, the most efficiently nitrated tyrosine residue in the food allergen ovalbumin (OVA) is part of human and murine IgE epitopes and also belongs to a human T cell epitope.³⁰⁴ Recent studies suggest that nitration may also affect the allergenic potential and adjuvant activity of α -amylase/trypsin inhibitors (ATIs) from wheat and other gluten-containing grains, which act as aeroallergens in baker's asthma and are involved in hyper-

sensitivities and chronic inflammation of the gastrointestinal tract.^{346–351} Nitrated variants of the major birch pollen allergen Bet v 1 induced enhanced levels of specific IgE in murine models, possibly because of the formation of neo-epitopes.²²⁹ Nitration of Bet v 1 also increased the presentation of allergen-derived peptides by antigen presenting cells (APC).³⁰⁵ Moreover, increased proteolytic stability, up-regulation of CCL17 (Th2-associated chemokine secreted by dendritic cells, DC), and alterations of T cell proliferation and stimulatory capacities have been observed for nitrated Bet v 1.³⁰⁶ Nitrated proteins also have been observed to modulate the antioxidant levels in murine pneumocytes.³⁵² In a recent study, in vivo fumigation of ragweed pollen with NO₂ resulted in an altered proteomic pattern including nitrosylation products and the treated pollen showed higher IgE recognition in immunoblots.²³⁹ Enhanced allergenic potential was also observed for *Betula pendula*, *Ostrya carpinifolia*, and *Carpinus betulus* pollen after NO₂ exposure (Table S1).²³⁶

Reaction product studies and kinetic experiments have shown that environmentally relevant O₃ and NO₂ concentrations can induce protein nitration on tyrosine residues.^{237,328–330,333,353–355} This is in line with earlier observations that atmospheric oxidation and nitration processes leads to the formation of nitrophenols and dinitrophenols,³⁵⁶ and that nitration is an important reaction pathway particularly in the atmospheric aqueous phase.^{357,358} Especially, aromatic amino acids like tyrosine and tryptophan can react with atmospheric nitrating agents, such as ozone/NO₂ mixtures or peroxyacetyl nitrate (PAN).^{330,359} Under photochemical smog conditions in polluted urban environments (high O₃ and NO₂ concentrations), proteins on the surface of aerosol particles can be efficiently nitrated within minutes to hours.^{328,330} The reaction kinetics also depends strongly on ambient relative humidity: At high relative humidity and especially during aqueous phase processing (when aerosol particles are activated as cloud or fog droplets), nitration may proceed efficiently also within the particle bulk.^{328,360,361}

Mechanistically, the reaction between O₃/NO₂ and tyrosine involves the formation of long-lived reactive oxygen intermediates (ROI), likely via hydrogen abstraction from the phenolic OH group, yielding tyrosyl radicals (phenoxy radical derivatives of tyrosine) that can further react with NO₂ to form nitrotyrosine residues as shown in Figure 4.^{329,362,363} The two-step protein nitration by air pollutants is similar to the endogenous nitration of proteins by peroxynitrite (ONOO[−])^{298,328,364} formed from nitrous oxide (NO) and superoxide anions (O₂[−]).^{301,365,366} For endogeneous protein nitration by ONOO[−], both radical and electron transfer reaction pathways have been proposed.³⁶⁷ Besides nitration, tyrosyl radicals can also undergo hydroxylation or self-reaction (cross-linking) to form dityrosine derivatives (Figure 4).³⁶⁸

The site selectivity of protein nitration is influenced by the molecular structure of the protein, the nitrating agent, and the reaction conditions. For example, different preferred reaction sites were observed for the birch pollen allergen Bet v 1, the egg allergen ovalbumin, and bovine serum albumin.^{304,328,333,354} Upon exposure of Bet v 1 to atmospherically relevant concentrations of O₃/NO₂ and physiologically relevant concentrations of ONOO[−], the preferred sites of nitration were tyrosine residues with high solvent accessibility and within a hydrophobic environment. Accordingly, nitrated tyrosine residues occurred mainly in the C-terminal helix and in the hydrophobic cavity (Figure S4).³²⁸ Both are key positions for

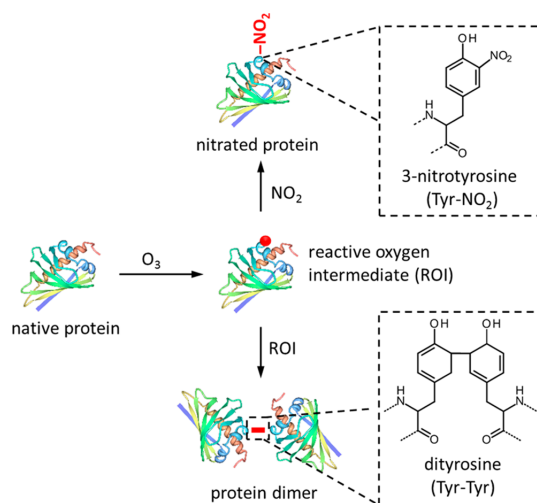


Figure 4. Posttranslational modification of proteins exposed to ozone (O₃) and nitrogen dioxide (NO₂). The initial reaction with O₃ leads to the formation of reactive oxygen intermediates (ROI, tyrosyl radicals), which can further react with each other to form cross-linked proteins (dityrosine) or with NO₂ to form nitrated proteins (nitrotyrosine). The shown protein is Bet v 1.0101 (PDB accession code 4A88,³⁷⁰ created with the PDB protein workshop 3.9⁴⁹⁸), for which nitration and cross-linking were found to influence the immunogenicity and allergenic potential.^{229,305,306,328} Red dot indicates a tyrosyl radical; red bar indicates dityrosine cross-link.

the binding of specific IgE,³⁶⁹ as well as ligands like fatty acids, cytokines, and flavonoids.^{370–372} The binding of such ligands may be involved in allergic and inflammatory immune responses by stabilizing Bet v 1 against endo/lysosomal degradation.³⁷³ Moreover, nitration-related changes in ligand-binding capacity might influence the interaction of allergenic proteins like Bet v 1 with adjuvant substances like lipopolysaccharide (LPS) and induce a shift from Th1 to Th2 responses, thus resulting in increased allergenicity.³⁰⁶

Dimerization and oligomerization are supposed to have a strong influence on the immunogenicity of allergenic proteins and are common features of major allergens like Bet v 1.^{307,308} The cross-linking of IgE receptors (FcεRI) on effector cells is a key element of allergic reactions and requires IgE antibody clustering on the cell surface,^{374,375} which may be facilitated by multivalent allergens, such as oligomers of allergenic proteins providing multiple epitopes of the same kind.^{122,376} Moreover, cross-linking can make proteins less susceptible to enzymatic proteolysis and influence immune responses.^{313,373,377} Indeed, immune responses to oligomers and aggregates of certain allergenic proteins were found to be enhanced compared to the monomeric form of the allergenic protein.^{307,308,378–380} The clustering of allergenic proteins on nanoparticle surfaces (protein coronas) can also modulate allergic responses depending on protein and particle properties.²⁴⁴ Accordingly, the investigation and effects of allergen colocalization on the surface of inhalable ambient particles, such as pollen fragments or soot (DEP), are potentially important research perspectives.

Oxidative protein cross-linking can occur upon (a) tyrosyl radical coupling through dityrosine cross-links, (b) Schiff-base coupling of oxidation-derived protein carbonyl groups with the ε-amino groups of lysine residues, and (c) intermolecular

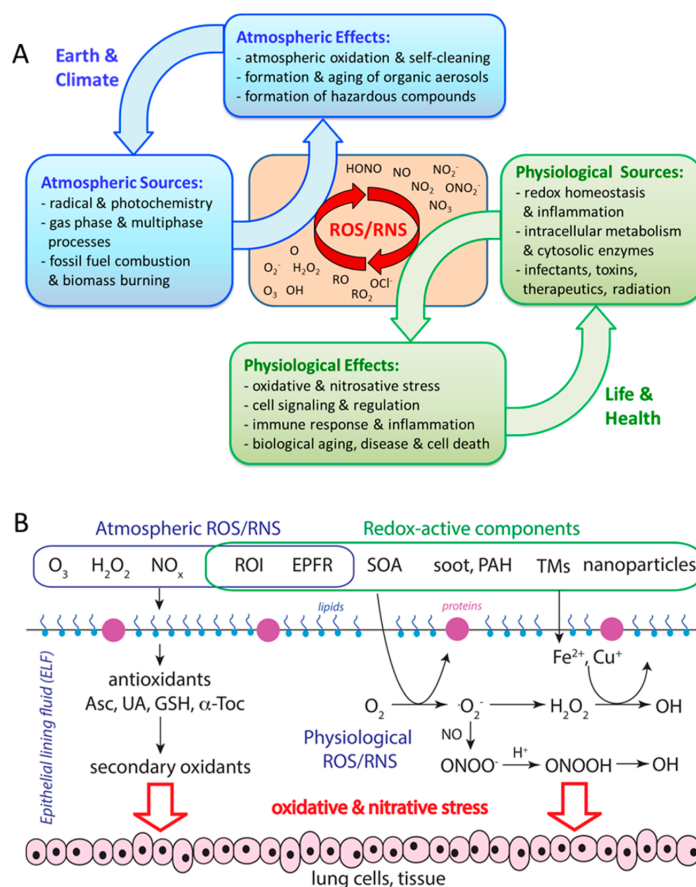


Figure 5. (A) Sources, effects, and interactions at the interface of atmospheric and physiological chemistry with feedback loops involving Earth System, climate, life, and health. (B) Interactions of atmospheric and physiological ROS/RNS with antioxidants (ascorbate, uric acid, reduced glutathione, α -tocopherol) in the epithelial lining fluid (ELF) of the human respiratory tract. Redox-active components, including reactive oxygen intermediates (ROI), soot, quinones and transition metals can induce ROS formation in vivo, leading to oxidative stress and biological aging. Adapted from ref 38. Copyright 2015 American Chemical Society.

disulfide coupling.³⁸¹ Recently, protein cross-linking and oligomerization upon exposure to atmospherically relevant concentrations of O_3 have been shown to proceed via the formation dityrosine cross-links as outlined in Figure 4, yielding up to ~10% of dimers, trimers, and higher oligomers of a model protein within minutes to hours of exposure under summer smog conditions.³⁶⁸ Similar reaction mechanisms involving reactive oxygen intermediates may also be responsible for the protein cross-linking observed upon reaction with physiological and synthetic nitrating agents like $ONOO^-$ and tetranitromethane, respectively.^{306,313,382,383} Cross-linking upon reaction with tetranitromethane was suggested to alter the immunogenicity and enhance the allergenicity of Bet v 1 through decreased endolysosomal degradation leading to extended MHC class II antigen presentation.³⁰⁶ On the other hand, oligomerization of allergens induced by modification with glutaraldehyde, that is, formation of glutaraldehyde bridges between nucleophilic amino acid residues (in particular lysine), was suggested to reduce immunogenicity and allergenicity due to delayed allergen uptake and presentation by dendritic cells.^{384,385}

As illustrated in Figure S2, the processes of allergic sensitization and response involve a wide range of interactions between protein molecules dissolved in liquids (blood, lymph, etc.) and embedded in semisolid structures (membranes, cells, tissues), which can be regarded as protein multiphase chemistry.³⁸ Protein reactions with ROS/RNS are generally pH-dependent and yield a mixture of hydroxylated, nitrated, cross-linked, aggregated or degraded products.^{386–391} To assess immune responses to specific posttranslational modifications of proteins, it is necessary to carefully characterize the investigated samples and avoid artifacts or misinterpretations that might arise from interferences between different reaction products and pathways, for example, nitration vs dimerization or oligomerization of proteins exposed to oxidizing and nitrating agents (Figure 4).

4. EPITHELIAL SURFACE INTERACTIONS

The deposition of particles in the respiratory tract is size-dependent and deposited particles are removed by a number of physical, chemical, and biological clearance processes, including

mucociliary movement, endo- and phagocytosis, dissolution, leaching, and protein binding.²⁰¹ Thus, the first step of an inhaled allergen-carrying particle is evading the mechanical defenses of the respiratory tract and passing, for example, alveolar macrophages, which prevent inappropriate immune activation by removing inhaled allergens via phagocytosis.^{392–394} The epithelial surface is a protective barrier, which protects the underlying tissue from many inhaled substances. The epithelial cells are covered by a viscous mucosal lining rich in immune cells and soluble components, such as antioxidants, complement proteins, and surfactant proteins.^{201,395,396} As the epithelium is more than a passive protective barrier, it recruits and activates more specialized immune cells and promotes inflammatory responses.³⁹⁷ Allergy is also discussed to be an epithelial barrier disease.^{15,131,398–400} For example, nasal epithelium is clearly different between healthy and allergic subjects and only in allergic subjects the transport of Bet v 1 is caveolar-mediated.⁴⁰¹

Air pollutants interacting with epithelial surfaces can act as adjuvants promoting pro-allergic innate and adaptive immune reactions as outlined in Table 2 and section S1. For example, they can induce inflammation and disrupt epithelial barriers, facilitating the access of allergens to immunogenic effector cells.^{8,86} In particular, air particulate matter can trigger ROS production through Fenton-like reactions and the activation of macrophages, mitochondria and enzymes related to the oxidant/antioxidant balance (e.g., NADPH oxidase, glutathione peroxidase).^{309,310,402–405} Additionally, pollution-derived ROS can induce proinflammatory responses by the production of damage associated molecular patterns (DAMPs oxidized phospholipids, hyaluronic acid, etc.) and trigger immune reactions leading to acute or chronic inflammation,^{29,406} for example, through feedback cycles involving Toll-like receptors (TLR) and other pattern recognition receptors (PRR) (Figure S5).⁴⁰⁷ Ozone and particulate matter can prime the airways for pro-allergic responses, and TLR signaling plays an important role in pollutant-induced inflammation.^{408,409} During inflammation, inducible nitric oxide synthase (iNOS) that is mainly expressed in innate immune cells (monocytes, macrophages, dendritic cells) provides high amounts of nitrogen oxide (NO), which can react with superoxide radicals to form peroxynitrite (ONOO[−]), a central endogenous nitrating agent for proteins.³⁰¹ In addition, particulate and gaseous pollutants may also drive pro-allergic inflammation through the generation of oxidative stress involving elevated levels of ONOO[−].⁴¹⁰

As illustrated in Figure 5A, epithelial surfaces are interfaces coupling the atmospheric and the physiological production, cycling, and effects of ROS/RNS.³⁸ Specific interactions of atmospheric ROS/RNS with antioxidants in the epithelial lining fluid are shown in Figure 5B. An increase of ozone from typical background concentration levels (~30 ppb) to summer smog conditions (>100 ppb) reduces the chemical half-life of antioxidants from days to hours,³⁰⁹ which may be comparable or shorter than the physiological replenishment rates.⁴¹¹ Furthermore, the adjuvant effect of ambient ultrafine particles was correlated with their oxidant potential.⁴¹² Major contributors to the redox properties of ambient particles are transition metals, polycyclic aromatic hydrocarbons, and derivatives (PAH, nitro/oxy-PAH), and semiquinones.^{38,312,412–415} In addition, the deposition of acidic particles may reduce the pH of the epithelial lining fluid (ELF). For healthy people the mean pH is ~7.4, while in people with diseases (e.g., asthma, acid reflux) it can be as low as ~4.^{416,417} Oxidant exposure and

changes of pH can alter reaction pathways of antioxidants⁴¹⁸ and also decrease the activities of antioxidant-related enzymes in the ELF, which are also reduced in smokers and people suffering from lung diseases.^{419–421}

Recent studies yielded chemical exposure-response relations between ambient concentrations of air pollutants and the production rates and concentrations of ROS in the ELF of the human respiratory tract.³⁰⁹ As illustrated in Figure 6, the total

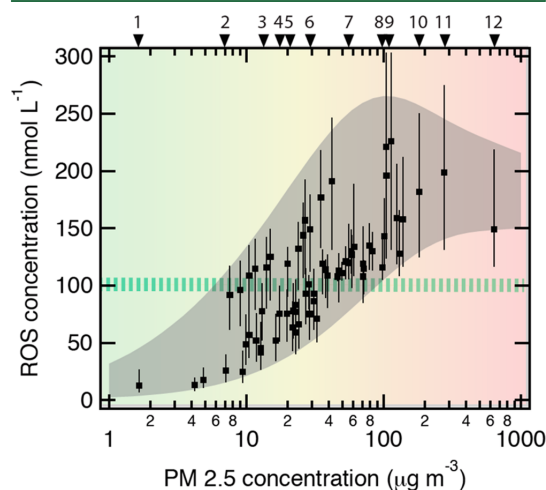


Figure 6. Chemical exposure-response relations between ambient concentrations of fine particulate matter (PM_{2.5}) and the concentration of reactive oxygen species (ROS) in the epithelial lining fluid (ELF) of the human respiratory tract. The green-striped horizontal bar indicates the ROS level characteristic for healthy humans (~100 nmol L^{−1}). The gray envelope represents the range of aerosol-induced ROS concentrations obtained with approximate upper and lower limit mass fractions of redox-active components observed in ambient PM_{2.5}. The data points represent various geographic locations for which measured or estimated mass fractions are available, including (1) Amazon, Brazil (pristine rainforest air); (2) Edinburgh, UK; (3) Toronto, Canada; (4) Tokyo, Japan; (5) Budapest, Hungary; (6) Hong Kong, China; (7) Milan, Italy; (8) Guangzhou, China; (9) Pune, India; (10) Beijing, China; (11) New Delhi, India; (12) Sumatra, Indonesia (biomass burning/peat fire smoke). Adapted from Lakey, S. J. P.; Berkemeier, T.; Tong, H.; Arangio, A. M.; Lucas, K.; Pöschl, U.; Shiraiwa, M. Chemical exposure-response relationship between air pollutants and reactive oxygen species in the human respiratory tract. *Sci. Rep.* **2016**, *6*, 32916. DOI: [10.1038/srep32916](https://doi.org/10.1038/srep32916).³⁰⁹ Copyright 2016 Lakey et al.

concentration of ROS generated by redox-active substances contained in fine particulate matter (PM_{2.5}) deposited in the ELF ranges from ~10 nmol L^{−1} under clean conditions up to almost ~250 nmol L^{−1} under highly polluted conditions. Thus, the inhalation of PM_{2.5} can increase ROS concentrations in the ELF to levels that exceed physiological background levels (50–200 nmol L^{−1}) and are characteristic for respiratory diseases.^{309,422} In addition to the effects of PM_{2.5}, ambient ozone readily saturates the ELF and can enhance oxidative stress by depleting antioxidants and surfactants.³⁰⁹ Ozone also reacts with skin lipids (e.g., squalene) and generates organic compounds (e.g., mono- and dicarbonyls) that can act as irritants.²⁶⁹ These and related organic compounds were found to act as adjuvants in the development of respiratory allergies as well as atopic dermatitis.^{270,271,423,424} Some air pollutants and

chemical reaction products formed at epithelial interfaces are sufficiently long-lived and mobile to diffuse through membranes and interact with the neural, cardiovascular, and immune system networks of the human body.^{314,425–429} Through these and related physiological interactions involving DAMPs, inflammatory mediators, cytokines, leukocytes etc., oxidative stress and inflammation caused by air pollutants may propagate from the respiratory tract and skin to other parts of the human organism and exert systemic influence on the development of allergies, reaching also the gastrointestinal tract.^{38,429}

A wide variety of commensal, symbiotic, and pathogenic microorganisms are found on the epithelial surfaces of the human body, such as the skin, lungs, and the gastrointestinal tract. Recent research suggests that the human microbiome is important to maintain physiological functions and to induce immune regulation by balancing the activities of Th1 and Th2 cells.^{430–433} Normal microbial colonization in early life can promote tolerance to aeroallergens via induced regulatory T cells.⁴³⁴ The development and composition of the human microbiome are influenced by many factors such as diet, infections, medical treatment, and also environmental factors.⁴³⁵ For example, air pollutants and climatic stress factors may disturb microbial communities through oxidative stress, inflammation, and changes in environmental biodiversity.⁴³⁶ Modifications in the composition of the gastrointestinal and lung microbiome can in turn affect the development of allergies in accordance with the “hygiene hypothesis”^{36,436–440} and may also promote pathogenic species that can contribute to these diseases.^{441–443} Recent studies revealed differences in the structure and composition of microbiota in the lower airways of healthy and asthmatic people: *Bacteroidetes*, *Firmicutes*, and *Proteobacteria* are the most common phyla found in airways of healthy subjects, whereas increased concentrations of pathogenic *Proteobacteria*, such as *Haemophilus*, *Moraxella*, and *Neisseria* spp., were found in asthma patients.^{442,443} Moreover, viral infections can exacerbate allergies.³¹ It is still unclear, however, if these changes are a cause or a consequence of the disease. Moreover, it has been suggested that air pollutants, especially air particulate matter, ingested together with food can trigger and accelerate the development of gastrointestinal inflammatory diseases by altering the gastrointestinal microbiome and immune functions.⁴⁴⁴ Besides the human microbiome, also microbes associated with allergenic pollen (pollen microbiome) and other aeroallergens may act as adjuvants when deposited on epithelial surfaces.^{235,445}

5. CONCLUSIONS AND OUTLOOK

As the globally pervasive anthropogenic influence continues to shape planet Earth and the human environment in the Anthropocene, it becomes increasingly important to understand and assess the potential effects of environmental change on human health. The widespread increase of allergies and their complex dependence on multiple influencing factors, including environmental pollution, indicate that allergic diseases are a major challenge with regard to maintaining and improving public health.

Anthropogenic emissions of atmospheric trace substances are affecting air quality and climate on local, regional, and global scales. Changes in atmospheric aerosol composition, oxidant concentrations, and climate parameters can induce chemical modifications of allergens, increase oxidative stress in the human body, and skew the immune system toward allergic reactions. In particular, air pollutants can act as adjuvants and

alter the immunogenicity of allergenic proteins, while climate change affects the abundance and properties of bioaerosols as carriers of aeroallergens. The production, release and properties of allergens and adjuvants are subject to various human interferences with the biosphere and climate system, including air pollutant interactions with natural and agricultural vegetation, fertilization and land-use change, as well as plant breeding and genetic engineering.

The following key questions remain to be resolved to understand and mitigate potential effects of air pollution and climate change on the observed increase and future development of allergies:

- (Q1) Which air pollutant and climate change effects have the largest potential to influence on the abundance and potency of allergens and adjuvants in the human environment (indoor and outdoor)?
- (Q2) Which elements and reaction pathways of the immune system are particularly susceptible to disturbance by air pollutants, and what are the most relevant chemical and physiological mechanisms (adjuvant activity vs allergen modification)?
- (Q3) Which environmental and physiological parameters are needed and best suited to account for and assess air pollutant and climate change effects in epidemiological studies of allergic diseases (attribution and prediction of environmental risk factors)?
- (Q4) How important are air pollutant and climate change effects relative to other environmental, lifestyle, genetic and epigenetic risk factors for allergic diseases?

Recommendations on how to address these key questions in future research are listed in Table S2, building on and extending suggestions given in related review and perspective articles (e.g., refs 8, 12, 93, and 280). Beyond addressing the above questions, it appears worthwhile to explore which components of the immune system could be modulated to prevent adverse effects of air pollution, for example, whether therapeutic monoclonal antibodies against relevant cytokines (e.g., IL-4, IL-5, IL-13) or IgE antibodies could make a difference. Further information about ongoing efforts and future perspectives of mitigating the health effects of climate change and air pollution is available from various national and international government agencies, medical institutions and related organizations (e.g., refs 4, 37, and 446). For efficient scientific progress, it will be important to combine and optimize state-of-the-art methods and results of environmental, immunological and epidemiological studies, tightly coupling physical, chemical, biological, and medical techniques and knowledge. One of the challenges consists in identifying and quantifying the mechanisms and feedback loops of immunochemical reactions in response to environmental influencing factors, including chemical modifications and interactions of allergens and adjuvants under realistic environmental and physiological conditions. For this purpose, the results of laboratory experiments and monitoring networks with improved detection methods for allergens, adjuvants and reactive intermediates should be used to design and inform epidemiological studies targeting the effects of different types and combinations of air pollutants and climate parameters.

■ ASSOCIATED CONTENT

■ Supporting Information

The Supporting Information is available free of charge on the ACS Publications website at DOI: 10.1021/acs.est.6b04908.

Allergic sensitization and response, air pollution effects on the allergenic potential of plant pollen and fungal spores, research activities proposed to address the key questions, essential steps and influencing factors in the development of IgE-mediated allergies, simplified scheme of major processes involved in allergic sensitization and response, scanning electron micrographs of oak and birch pollen, 3D-structure of the major birch pollen allergen Bet v 1.0101, and TLR4-radical cycle of inflammation (PDF)

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Notes

The authors declare no competing financial interest.

■ ACKNOWLEDGMENTS

K.R.-S. acknowledges financial support by the Max Planck Graduate Center with the Johannes Gutenberg University of Mainz (MPGC); C.J.K. acknowledges support by the MPGC and financial support by the German Research Foundation (DFG, grant number KA 4008-1/2); J.F.-N., U.P., and B.W. acknowledge support from the German Research Foundation (DFG FR3641/1-2, FOR 1525 INUIT). N.L.-Y. acknowledges support from the Max Planck Society and from the Weizmann Institute of Science (National Postdoctoral Award Program for Advancing Women in Science). F.L. and S.L. acknowledge financial support by the China Scholarship Council (CSC). A.D. acknowledges funding from the FWF (W-1213) and from the University of Salzburg via the ACNB program. Thanks to M. Trainic, M. Riekert, and S. Benner for support with graphical illustrations. The authors acknowledge stimulating exchange and discussions with Paul J. Crutzen, the members of the Mainz Program for Chemical Allergology (MPCA), the Allergie-Zentrum Rheinland-Pfalz (AZ-RP), and the Mainz Bioaerosol Laboratory (MBAL), as well as numerous colleagues in the scientific communities of the Earth, Environmental, and Life Sciences.

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Atmospheric protein chemistry influenced by anthropogenic air pollutants: nitration and oligomerization upon exposure to ozone and nitrogen dioxide

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Faraday Discussions **200**, 413-427 (2017)

Online supplement available at:

<http://www.rsc.org/suppdata/c7/fd/c7fd00005g/c7fd00005g1>

Contribution

Aided in writing the manuscript.



Atmospheric protein chemistry influenced by anthropogenic air pollutants: nitration and oligomerization upon exposure to ozone and nitrogen dioxide†

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Received 5th January 2017, Accepted 10th March 2017

DOI: 10.1039/c7fd00005g

The allergenic potential of airborne proteins may be enhanced *via* post-translational modification induced by air pollutants like ozone (O₃) and nitrogen dioxide (NO₂). The molecular mechanisms and kinetics of the chemical modifications that enhance the allergenicity of proteins, however, are still not fully understood. Here, protein tyrosine nitration and oligomerization upon simultaneous exposure of O₃ and NO₂ were studied in coated-wall flow-tube and bulk solution experiments under varying atmospherically relevant conditions (5–200 ppb O₃, 5–200 ppb NO₂, 45–96% RH), using bovine serum albumin as a model protein. Generally, more tyrosine residues were found to react *via* the nitration pathway than *via* the oligomerization pathway. Depending on reaction conditions, oligomer mass fractions and nitration degrees were in the ranges of 2.5–25% and 0.5–7%, respectively. The experimental results were well reproduced by the kinetic multilayer model of aerosol surface and bulk chemistry (KM-SUB). The extent of nitration and oligomerization strongly depends on relative humidity (RH) due to moisture-induced phase transition of proteins, highlighting the importance of cloud

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† Electronic supplementary information (ESI) available: Chemical mechanism and corresponding parameters used in kinetic modelling (Table S1), schematic experimental setup (Fig. S1), and SEC calibration curve (Fig. S2). See DOI: 10.1039/c7fd00005g



processing conditions for accelerated protein chemistry. Dimeric and nitrated species were major products in the liquid phase, while protein oligomerization was observed to a greater extent for the solid and semi-solid phase states of proteins. Our results show that the rate of both processes was sensitive towards ambient ozone concentration, but rather insensitive towards different NO₂ levels. An increase of tropospheric ozone concentrations in the Anthropocene may thus promote pro-allergic protein modifications and contribute to the observed increase of allergies over the past decades.

1. Introduction

Allergies represent an important issue for human health and the prevalence of allergic diseases has been increasing worldwide over the past decades.^{1,2} Among other explanations, air pollution has been proposed as a potential driver for this increase.^{3–6} It is well established that air pollutants, especially diesel exhaust particles (DEPs), can act as adjuvants and facilitate allergic sensitization in the human body.^{7,8} Air pollutants like nitrogen dioxide (NO₂), sulfur dioxide (SO₂), and ozone (O₃) have been shown to interact with and modify allergen carriers like pollen grains and fungal spores, increasing the release of allergenic proteins.^{8,9} Moreover, post-translational modifications (PTM) of allergenic proteins can be induced by reactive trace gases such as O₃ and NO₂ which modifies their structure and activity, thus altering the immunogenicity of the proteins.^{10–13}

Airborne allergenic proteins (aeroallergens) are contained not only in coarse biological particles such as pollen grains,¹⁴ but also in the fine fraction of air particulate matter (aerodynamic diameter < 2.5 μm).^{15–17} The occurrence of allergenic proteins in fine particles can be explained by several processes, including the release of pollen cytoplasmic granules (PCGs) from the rupture of pollen grains,⁹ fragmentation of airborne cellular material,¹⁸ and contact transfer of allergenic proteins onto fine particles.^{18,19} Therefore, aeroallergens can be directly exposed to ambient O₃ and NO₂, promoting chemical modifications like tyrosine (Tyr) nitration and oligomerization.

Laboratory and field investigations have shown that proteins can be oxidized, nitrated and/or oligomerized upon exposure to NO₂ and O₃ in synthetic gas mixtures or polluted urban air.^{11,12,15,20} The mechanisms of protein nitration by O₃ and NO₂, and protein cross-linking (oligomerization) by O₃ both involve the formation of long-lived reactive oxygen intermediates (ROIs), which are most likely tyrosyl radicals, as proposed earlier.^{12,20–24} The ROIs can subsequently react with each other forming dityrosine (DTyr) crosslinks with NO₂ to form nitro-tyrosine (NTyr) residues, or undergo further oxidation reactions. Using quantum chemical methods, Sandhiya *et al.*²² showed that six different intermediates can be formed through the initial oxidation of Tyr residues by O₃, out of which the tyrosyl radical is favorable due to a small energy barrier, particularly in the aqueous phase. In the absence of NO₂, tyrosyl radicals can undergo self-reaction to stabilize in the form of dimers. Under physiological conditions, Pfeiffer *et al.*²⁵ found that DTyr was a major product of Tyr modification caused by low steady-state concentrations of peroxyxynitrite, while high fluxes (>2 μM s^{−1}) of nitrogen oxide/superoxide anions (NO/O₂[−]) are required to render peroxyxynitrite an efficient trigger of Tyr nitration. Thus, kinetic competition between Tyr nitration and dimerization (or oligomerization) upon protein exposure to O₃ and NO₂ can be



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expected, which needs to be explored in detail to assess the relevant atmospheric conditions that favor potentially health relevant protein modifications.

In this study, we explored the oxidation, nitration, and oligomerization reactions of proteins induced by O_3 and NO_2 , and their kinetics under different atmospherically relevant conditions using bovine serum albumin (BSA) as a model protein. Coated-wall flow-tube and bulk solution experiments were performed to study the kinetics of protein nitration and oligomerization at O_3 and NO_2 concentrations of 5–200 ppb, and relative humidities (RH) of 45% and 96%, utilizing a size exclusion chromatography/spectrophotometry method. Additionally, we used the kinetic multilayer model of aerosol surface and bulk chemistry (KM-SUB)²⁶ to investigate which chemical reactions and transport processes control the concentration and time dependence of protein oligomerization and nitration.

2. Experimental

2.1. Materials

Bovine serum albumin (BSA, A5611) and sodium phosphate monobasic monohydrate ($NaH_2PO_4 \cdot H_2O$, 71504), were purchased from Sigma Aldrich (Germany). Sodium hydroxide (NaOH, 0583) was purchased from VWR (Germany). 10× Tris/glycine/SDS (161-0732) was purchased from Bio-Rad Laboratories (USA). High purity water (18.2 MΩ cm) for chromatography was taken from a Milli-Q Integral 3 water purification system (Merck Millipore, USA). The high purity water (18.2 MΩ cm) was autoclaved before use if not specified otherwise.

2.2. Protein O_3/NO_2 exposure setup

Reactions of BSA with O_3/NO_2 mixtures were performed both homogeneously in aqueous solutions and heterogeneously *via* the exposure of BSA-coated glass tubes to gaseous O_3/NO_2 at different levels of relative humidity (RH). Before the exposure experiments, BSA solutions (0.6 mL , 0.33 mg mL^{-1}) were loaded into the glass tube and dried by passing a nitrogen (N_2 , 99.999%) flow at $\sim 1 \text{ L min}^{-1}$ through a specific rotating device,²⁷ which is essential to ensure homogeneous coating and experiment reproducibility. The BSA-coated glass tube was then connected to the experimental setup. The experimental setup (Fig. S1, ESI†) was described previously,²⁰ and was extended by incorporating an additional flow of NO_2 after the humidifier.

Briefly, ozone was produced from synthetic air passed through a UV lamp (L.O.T.-Oriol GmbH & Co. KG, Germany) at $\sim 1.9 \text{ L min}^{-1}$. The gas flow was then split and one flow was passed through a Nafion® gas humidifier (MH-110-12F-4, PermaPure, USA) operated with autoclaved high purity water, while the other flow remained dry. RH could be varied in a wide range by adjusting the ratio between the dry and humidified air flow. During the experiments, the standard deviation from the set RH values was <2% RH. The gas flow with a set O_3 concentration and RH was then mixed with a N_2 flow containing $\sim 5 \text{ ppmV } NO_2$ (Air Liquide, Germany). The NO_2 concentrations were adjusted by varying the flow rate ($20\text{--}80 \text{ mL min}^{-1}$) of the $\sim 5 \text{ ppmV } NO_2$ flow. The combined gas flow was then directed through the BSA-coated glass tube. The concentrations of O_3 and NO_2 as well as RH were measured using commercial monitoring instruments (Ozone analyzer,



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49i, Thermo Scientific, Germany; NO_x analyzer, 42i-TL, Thermo Scientific, Germany; RH sensor FHA 646-E1C with ALMEMO 2590-3 instrument, Ahlborn, Mess- und Regelungstechnik, Germany). After exposure, the proteins were extracted from the glass tube with 1.5 mL of 1× Tris/glycine/SDS (pH 8.3) buffer to avoid precipitation of protein oligomers in the extract solution.

For homogeneous bulk solution reactions, the O₃/NO₂ gas mixtures were directly bubbled through 1.5 mL of 0.13 mg mL⁻¹ BSA aqueous solutions (pH 7.0 ± 0.2; measured by a pH meter, model WTW multi 350i, WTW, Germany) at a flow rate of 60 mL min⁻¹ using a Teflon tube (ID: 1.59 mm). All heterogeneous and homogeneous exposure experiments were performed in duplicate.

2.3. SEC-HPLC-DAD analysis

Product analysis was performed using high-performance liquid chromatography coupled to diode array detection (HPLC-DAD, Agilent Technologies 1200 series). The HPLC-DAD system consisted of a binary pump (G1379B), an autosampler with a thermostat (G1330B), a column thermostat (G1316B), and a photodiode array detector (DAD, G1315C). ChemStation software (Rev. B.03.01, Agilent) was used for system control and data analysis. Molecular weight (MW) separation by size exclusion chromatography (SEC) was carried out using an AdvanceBio SEC column (Agilent, 300 Å, 300 × 4.6 mm, 2.7 μm). Isocratic separation at a flow rate of 0.35 mL min⁻¹ was carried out using a mobile phase of 150 mM NaH₂PO₄ buffer (adjusted to pH 7 with 10 M NaOH (aq)) after injecting 40 μL of sample. The absorbance was monitored by the DAD at wavelengths of 220, 280 and 357 nm. Each chromatographic run was performed in duplicate.

A protein standard mix 15–600 kDa (69385, Sigma Aldrich, Germany) containing bovine thyroglobulin (MW = 670 kDa), γ-globulin from bovine blood (MW = 150 kDa), chicken egg albumin, grade VI (MW = 44.3 kDa), and ribonuclease A (MW = 13.7 kDa) was used for SEC column calibration (elution time *vs.* log MW). Details can be found in Fig. S2 in the ESI.† It should be noted that SEC separates molecules according to their hydrodynamic sizes, thus only approximate molecular masses can be obtained by this calibration method.

2.4. Determination of protein oligomer mass fractions and nitration degrees

A detailed description of the simultaneous determination of protein oligomer mass fractions and nitration degrees using the SEC-HPLC-DAD analysis described above can be found in Liu *et al.*²⁸ Briefly, we report the formation of BSA oligomers as the temporal evolution in the ratios of the respective oligomers (dimers, trimers, and oligomers with $n \geq 4$) to the sum of monomer and all oligomer peak areas at the absorption wavelength of 220 nm. Assuming that the molar extinction coefficients of the individual protein oligomer fractions are multiples of the monomer extinction coefficient, the calculated oligomer ratios correspond to the mass fractions (ω) of the individual oligomers. Nitration degrees (NDs), defined as the concentration of nitrotyrosine (NTyr) divided by the sum of the concentrations of NTyr and Tyr, were obtained for BSA monomers and dimers, using the respective peak areas of the monomer and dimer signals at wavelengths of 280 nm and 357 nm. For calculation of the total protein ND, the sum of the peak areas of all protein signals at wavelengths of 280 nm and 357 nm was used. Note that corresponding to the definition of the ND, the same number of nitrated Tyr



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residues in a BSA monomer and dimer will lead to a factor of 2 difference in the individual NDs, because a BSA dimer contains twice the number of Tyr residues compared to the monomer. Further information on the calculation of NDs can be found in Liu *et al.*²⁸ The values and errors of the calculated NDs and oligomer mass fractions represent the arithmetic mean values and standard deviations of duplicate experiments.²⁹ The commercially available BSA also contains dimers and trimers of the protein as well as pre-nitrated monomers and dimers (~NDs 0.9%). Therefore, the reported values of oligomer mass fractions and NDs were corrected for these background signals.

3. Results and discussion

3.1. Protein oligomerization

Fig. 1 and 2 show the effects of varying NO_2 and O_3 concentrations on protein oligomerization for homogeneous bulk solution and coated-wall flow-tube experiments, respectively. Exposures were carried out at fixed NO_2

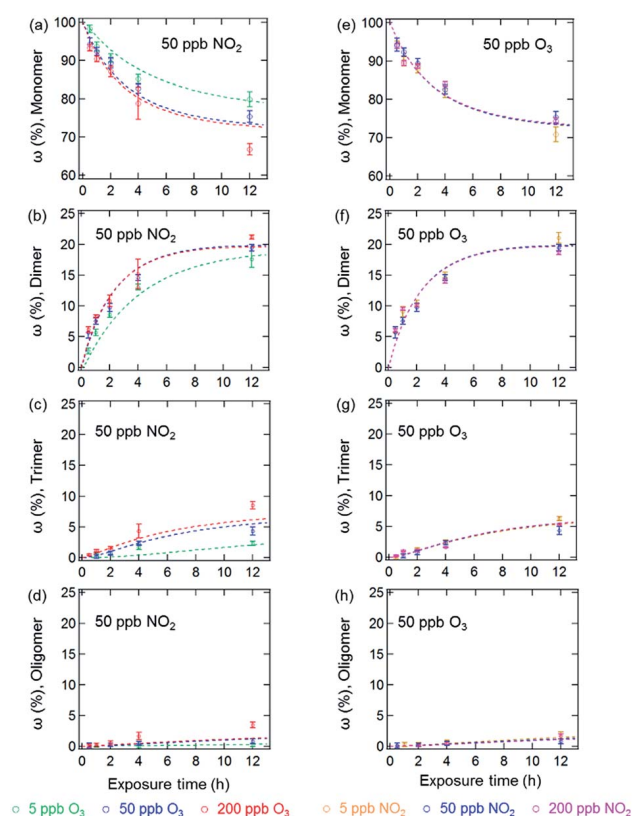


Fig. 1 Temporal evolution of protein oligomer mass fractions (ω (%), monomer, dimer, trimer and oligomer ($n \geq 4$)) in the aqueous phase reaction of BSA with O_3/NO_2 : ((a)–(d)) at a fixed NO_2 concentration of 50 ppb with varied O_3 concentrations; ((e)–(h)) at a fixed O_3 concentration of 50 ppb with varied NO_2 concentrations. The data points and error bars represent the arithmetic mean values and standard deviations of duplicate experiments. The dashed lines are the results of the kinetic model.



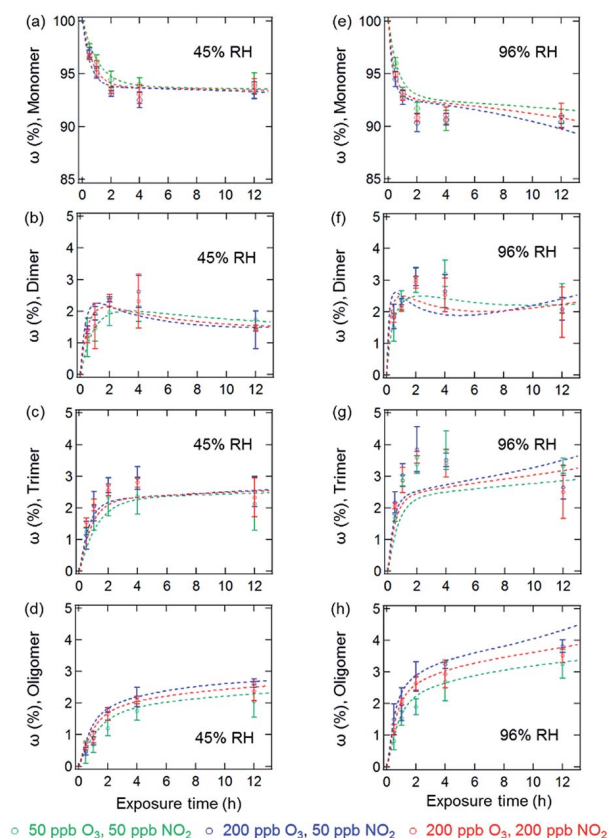


Fig. 2 Temporal evolution of protein oligomer mass fractions (ω (%), monomer, dimer, trimer and oligomer ($n \geq 4$)) upon exposure of BSA films (thickness 34 nm) to various O_3 /NO $_2$ concentrations: ((a)–(d)) at 45% RH; ((e)–(h)) at 96% RH. The data points and error bars represent the arithmetic mean values and standard deviations of duplicate experiments. The dashed lines are the results of the kinetic model.

concentrations with varying O_3 concentrations and *vice versa*. The exposure time was varied from 0.5 up to 12 h. While in bulk solution experiments dimers were generally observed as the major reaction products of BSA with O_3 and NO $_2$, trimers or higher oligomers can be dominant products in the coated-wall flow-tube experiments at longer exposure times, depending on the experimental conditions.

The results of the bulk solution experiments on protein oligomerization are illustrated in Fig. 1. Generally, the mass fractions of dimers, trimers, and higher oligomers increase with increasing reaction times, reaching up to $21 \pm 1\%$ for dimers, $9 \pm 1\%$ for trimers, and $4 \pm 1\%$ for oligomers with $n \geq 4$ after 12 h of exposure. The minimum mass fraction of monomers correspondingly was found to be 66%. While varying the O_3 concentrations (Fig. 1a–d, fixed 50 ppb of NO $_2$) from 5 to 200 ppb significantly affected the temporal evolutions observed for the mass fractions of the different oligomers, changing the NO $_2$ concentration (Fig. 1e–h, fixed 50 ppb of O_3) in the same range did not result in significant changes in oligomer mass fractions. The solubility of O_3 and NO $_2$ in water is



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$\sim 10^{-5}$ mol mL $^{-1}$ (derived from their Henry's law constants: $K_{\text{sol,cc,O}_3} \approx K_{\text{sol,cc,NO}_2} \approx 10^{-2}$ M atm $^{-1}$)³⁰ under our experimental conditions. Increasing O₃ and NO₂ gas concentrations between 5 to 200 ppb should result in concentrations of O₃ and NO₂ between 7×10^{-11} to 3×10^{-9} M in the aqueous phase. Thus, the insignificant change of oligomer mass fractions with varied NO₂ concentration should not be caused by a saturation of dissolved NO₂ in the concentration range investigated (5 to 200 ppb).

Mechanistically, the reactions between O₃/NO₂ and protein Tyr residues involve the formation of ROIs (tyrosyl radicals) resulting from the reaction of Tyr with O₃. These ROIs can then either react with NO₂ to form NTyr residues or cross-link due to intermolecular DTyr formation.^{6,20} Ozonolysis of Tyr may also result in other oxidized products such as 3,4-dihydroxyphenylalanine (DOPA).³¹ However, the reaction mechanism for the formation of these oxidized products is not the focus of this study and we only consider these modified Tyr derivatives in the proposed mechanism (Table S1†) as a portion of the oxidized amino acid residues. Regardless, an inhibition of intermolecular DTyr cross-linking would be expected with increasing NO₂ concentrations. However, no such behavior was observed. Furthermore, similar protein oligomer mass fractions were observed previously in the absence of NO₂ for BSA exposed to O₃ in bulk solution experiments with comparable levels of O₃ (50 and 200 ppb).²⁰ This observation may be due to the high number of accessible Tyr residues on the dissolved BSA molecules in solution, because after 12 h of exposure, 66% of BSA (Fig. 1a) is still present in monomeric form.

The results of the coated-wall flow-tube experiments on protein oligomerization are illustrated in Fig. 2. In these experiments, thin protein films were exposed to O₃/NO₂ mixtures. A film thickness of ~ 34 nm, or roughly five layers of BSA, can be calculated assuming an even distribution of the BSA molecules on the inner surface of the glass tube.²⁰ The experiments were performed at 45% and 96% RH with O₃/NO₂ concentrations of 50/50, 200/50, and 50/200 ppb, respectively. Generally, the reactive sites accessible for oligomerization reactions of the proteins are limited here compared to aqueous solutions, leading to smaller mass fractions of protein oligomers. Furthermore, we observed a 30–40% reduction of the overall oligomer mass fraction (dimer, trimer, and oligomer ≥ 4) compared to similar flow-tube experiments in the absence of NO₂ for comparable RH and O₃ concentrations.²⁰ Apparently, the lower diffusivity of the proteins in this solid (45% RH) or semi-solid (96% RH) state induces a competition between DTyr and NTyr formation, also indicated by the observation of slower reaction rates for oligomerization in the bulk of the thin protein film compared to its surface.^{20,23}

In contrast to the bulk solution experiments, which show a steady increase of the oligomer mass fractions with exposure time, dimer and trimer mass fractions in the flow tube experiments peaked at exposure times of 2–4 h, depending on RH and trace gas concentrations, while only higher oligomers steadily increased over the course of the reaction time (see Fig. 2). This result indicates that as the exposure proceeds, the formation of dimers and trimers becomes slower than their consumption converting them into higher oligomers. The characteristic residence times (lifetimes) of biological particles in the atmosphere can range from hours to weeks, depending on their sizes, aerodynamic, and hygroscopic properties.³² Our observation indicates that the initial exposure (<2–4 h) of proteins to O₃ and NO₂ mainly leads to the formation of protein dimers and



trimers, and as the exposure proceeds, protein oligomers could be the dominant protein species, *e.g.*, on the surface of bioaerosol particles.

3.2. Protein nitration

Fig. 3 and 4 show the effects of varying NO_2 and O_3 concentrations on the nitration of protein monomers and dimers in homogeneous bulk solution and coated-wall flow-tube experiments, respectively. Exposures were carried out at fixed NO_2 concentrations with varying O_3 concentrations and *vice versa*. The exposure time was varied from 0.5 up to 12 h. While in previous studies total protein nitration degrees (NDs) were investigated upon exposure to O_3 and NO_2 ,^{12,33} we explicitly explored and characterized the reaction kinetics of protein nitration, resolving the individual NDs of the protein monomer and its various oligomers over the course of reaction time.

The results of the bulk solution experiments on protein nitration are illustrated in Fig. 3. The maximum ND of protein monomers and dimers were found to be 7% and 5% after 12 h exposure to 200 ppb O_3 and 50 ppb NO_2 , respectively. These results correspond to 1.4 NTyr residues per monomer molecule and 2 NTyr residues per dimer molecule (NTyr/monomer and NTyr/dimer are shown as secondary y-axes in Fig. 3 and 4). We found a positive relationship between the NDs and O_3 concentration (Fig. 3a and b), particularly that the increase of O_3 concentration by one order of magnitude from 5 to 50 ppb resulted in an increase of NDs from $4.2 \pm 0.2\%$ to $6.9 \pm 0.2\%$, and $2.0 \pm 0.3\%$ to $4.5 \pm 0.3\%$ after 12 h exposure for the monomer and dimer, respectively. Also, for protein nitration, no significant difference was found when concentrations of NO_2 were varied from 5 to 200 ppb at a fixed O_3 concentration of 50 ppb, as shown in Fig. 3c and d. These results are in accordance with the observations made by Shiraiwa *et al.*²³ on the

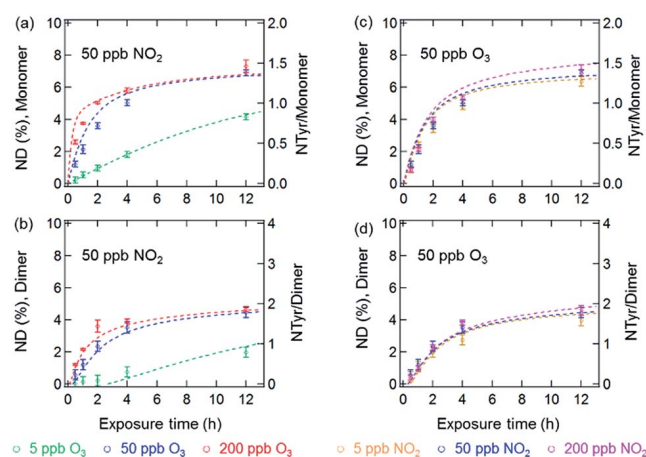


Fig. 3 NDs of protein monomer and dimer (primary y-axis), and NTyr number per monomer and dimer (secondary y-axis) plotted against reaction time in the aqueous phase reaction of BSA with O_3/NO_2 : (a) and (b) at a fixed NO_2 concentration of 50 ppb with varied O_3 concentrations, and (c) and (d) at a fixed O_3 concentration of 50 ppb with varied NO_2 concentrations. The data points and error bars represent the arithmetic mean values and standard deviations of duplicate experiments. The dashed lines are the results of the kinetic model.



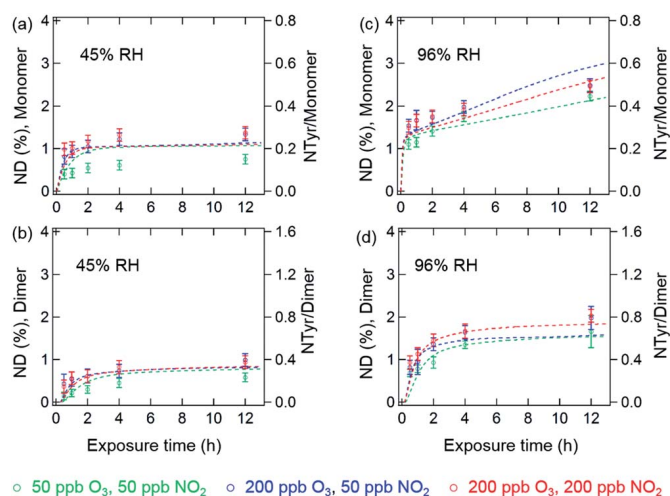


Fig. 4 NDs of protein monomer and dimer (primary y-axis), and NTyr number per monomer and dimer (secondary y-axis) plotted against exposure time upon the exposure of BSA films (thickness 34 nm) to various O_3/NO_2 concentrations: (a) and (b) at 45% RH, and (c) and (d) at 96% RH. The data points and error bars represent the arithmetic mean values and standard deviations of duplicate experiments. The dashed lines are the results of the kinetic model.

study of the reactive uptake of NO_2 by aerosolized proteins. They found that the uptake coefficient of NO_2 (γ_{NO_2}) increased with increasing O_3 concentrations while γ_{NO_2} decreased gradually with increasing NO_2 concentration. Thus, our results confirm that the protein reaction with O_3 and formation of ROI is the rate-limiting step for protein nitration.^{21,23} Shiraiwa *et al.*²³ have excluded NO_3 or N_2O_5 (formed through NO_2 oxidation by O_3) as major contributors to protein nitration. Ghiani *et al.*³⁴ reported that nitration of proteins can also occur by nitrate ions in bulk solutions without UV irradiation under acidic conditions ($\text{pH} < 3$ for BSA). The reaction of NO_2 with water can form nitric acid and thereby nitrate ions might appear in the BSA solution. However, we found that the pH of the BSA solutions stayed relatively constant ($\text{pH} 6.6 \pm 0.2$; pH meter model WTW multi 350i) for a reaction time of 12 h at 200 ppb NO_2 and 50 ppb O_3 , indicating that nitration induced by nitrate ions is likely a minor or negligible pathway in this study. This hypothesis is consistent with the results in Fig. 3c and d that show only a slight increase in ND (monomer, $6.3 \pm 0.3\%$ to $6.9 \pm 0.2\%$, and dimer, $4.0 \pm 0.3\%$ to $4.5 \pm 0.3\%$, for 12 h reaction) for a one order of magnitude increase in the NO_2 concentration from 5 to 50 ppb.

The results of the temporal increase of NDs of monomer and dimer for reactions of the thin protein films with O_3 and NO_2 concentrations of 50 and 200 ppb at 45% and 96% RH are illustrated in Fig. 4. Here, the NDs for monomer and dimer at 45% RH were found to be around 1% and 0.8% for 12 h exposure, corresponding to 0.2 NTyr/monomer molecule and 0.3 NTyr/dimer molecule. Note that the protein coating consisted of ~ 5 layers. Therefore, the results likely indicate that on average one Tyr per BSA monomer molecule located on the surface of the protein film was efficiently nitrated, while the bulk nitration occurred at much slower rates. This highly reactive site could be the Tyr residue at



position 161 (Y161), as suggested by Zhang *et al.*³⁵ The maximum NDs for monomer and dimer reached up to $2.5 \pm 0.1\%$ and $2.0 \pm 0.1\%$ at 96% RH, respectively. However, both of the maximum NDs at 45% and 96% RH were much lower than those obtained for O_3/NO_2 exposure in aqueous solutions (200 ppb O_3 and 50 ppb NO_2). This discrepancy can be explained by a decrease in viscosity and an increase in diffusivity on going from an amorphous solid (45% RH) or semi-solid (96% RH) protein on a coated wall to an aqueous protein solution and was able to be reproduced using a model (see below).^{36,37} Furthermore, the NDs of BSA in this study are comparable to those found by Yang *et al.*³³ using the same protein, whereas the nitration efficiency of BSA in general is found to be much lower than the one observed for the major birch pollen allergen Bet v 1 in similar exposure experiments,¹² indicating a strong influence of molecular structure and potentially the amino acid sequence of the reacting protein. From previous mass spectrometric analysis of the site selectivity of protein nitration by O_3/NO_2 , it is known that only 3 out of 21 Tyr residues in BSA can be detected in the nitrated form, while in Bet v 1, 4 out of 7 Tyr residues can be nitrated.^{12,35} Thus, besides the types of nitrating agents (*e.g.* $ONOO^-$ or O_3/NO_2) and reaction conditions (in aqueous solution or heterogeneous exposure), the nitration efficiency also depends on the fraction of reactive Tyr residues in the investigated protein.

3.3. Kinetics and mechanism of protein nitration and oligomerization by O_3/NO_2

The model results for the reactions of proteins with O_3/NO_2 under the various exposure conditions are shown as the dashed lines in Fig. 1–4. A chemical mechanism involving 19 reactions (see Table S1† for details) was applied in two kinetic models, *i.e.*, a box model for bulk solution experiments and a kinetic multilayer model for aerosol surface and bulk chemistry (KM-SUB)²⁶ for flow tube

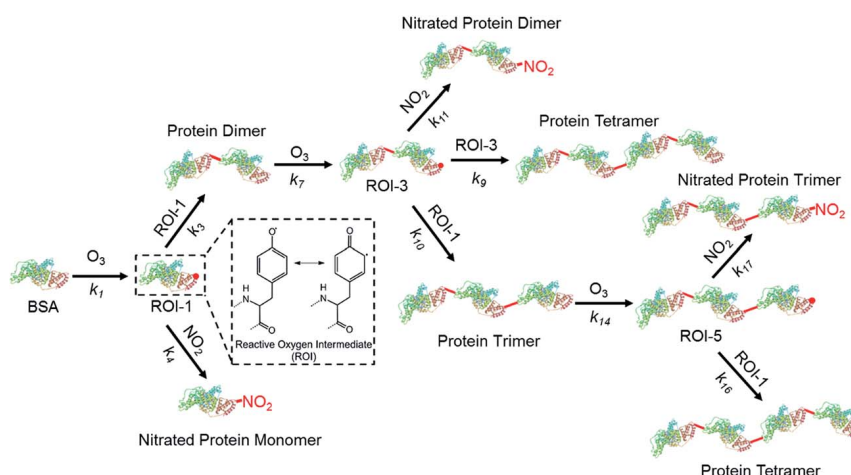


Fig. 5 Schematic overview of the most relevant reactions for protein nitration and oligomerization upon exposure to ozone and nitrogen dioxide. The reactions correspond to protein surface Tyr nitration and oligomerization in the mechanism presented in Table S1.† The molecular structure of the protein (BSA, PDB accession number 3V03) was created using RCSB PDB protein workshop (4.2.0) software.



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experiments to fit the experimental data. The most relevant reactions for this mechanism are illustrated in Fig. 5. The first step in the mechanism is the reaction of a Tyr residue with O_3 forming tyrosyl radicals as long-lived reactive oxygen intermediates (ROIs). In the second step of the process, the ROIs can react with each other to form dimers, or with NO_2 to form nitrated monomers. Note that for simplification, each molecule only contains one reactive tyrosine residue, while nitrated and oxidized monomers, dimers and trimers are unable to take part in further reactions in the kinetic model. A dimer can react further with O_3 to form a dimeric ROI, which may react with NO_2 to form a nitrated protein dimer, with monomeric ROI to form a protein trimer or with another dimeric ROI to form a protein tetramer.

The following assumptions were made to enable modelling of the reaction system for bulk solution and coated-wall flow-tube experiments using the sets of physicochemical parameters shown in Table S1 (ESI†): BSA molecules have reactive amino acid residues on their surface (AA1) and in their bulk (AA2), both of them reactive towards O_3 . While ROIs formed in the protein bulk can only react with NO_2 to form NTyr, they are unable to form intermolecular DTyr due to steric hindrance.³⁸ These assumptions were also applied to dimers and trimers. Besides Tyr, O_3 can also oxidize other amino acid residues, *i.e.*, cysteine (Cys), tryptophan (Trp), methionine (Met) and histidine (His).³¹ Among them, only Cys is able to cross-link proteins directly upon O_3 exposure, yielding intermolecular disulfide bridges, as one free Cys is available in BSA.³⁹ This reversible cross-linking mechanism has been shown to be only a minor contributor to protein oligomerization upon O_3/NO_2 exposure previously.²⁰ The kinetic parameters were obtained using a global optimization method combining a uniformly-sampled Monte Carlo search with a genetic algorithm (MCGA method).^{40,41} The genetic algorithm was terminated when the correlation between experimental data and model output reached an optimum. Concentrations of O_3 and NO_2 in the aqueous phase can be estimated using the published Henry's law constant of $K_{sol,cc,O_3} \approx K_{sol,cc,NO_2} \approx 10^{-2} \text{ M atm}^{-1}$, which were used as fixed values in the model.⁴⁰

The temporal evolution of NDs and oligomer mass fractions in aqueous solution is well reproduced by the model (Fig. 1 and 3). For the heterogeneous reactions studied in the coated-wall flow-tube experiments at 45% RH and 96% RH, some substantial deviations between modelled and measured data can be observed, and the coefficient of determination (R^2 value) is approximately a factor of two lower than for the aqueous data. For example, the oligomer mass fractions ω at 45% RH in Fig. 2a–d are qualitatively captured fairly well, while the model fails to reproduce their evolution at higher RH, especially for the dimer and trimer (Fig. 2f and g). The observed reduction of dimers in flow tube experiments could be reproduced by the model including the reactions on the surface, where the rates are four orders of magnitude higher than that of bulk reactions. Despite simple model assumptions when describing the complex reaction system that was studied, the model reproduces the experimental data reasonably well with an overall R^2 value of 0.88 for Fig. 1–4. Most of the optimized parameters obtained for the flow tube experiments were close to or the same as those for aqueous solutions (for details see Table S1, ESI†). However, note that some of the rate coefficients would be expected to change as the liquid water content and viscosity varies. Water could actively take part in some of the reactions and its presence at different concentrations could lead to changes in experimental conditions, such



as pH, which would influence the rate of the reactions. As the viscosity increases it is also expected that some rate coefficients would decrease as they become diffusion-limited as per the Smoluchowski diffusion equation.^{42,43}

The second-order rate coefficients obtained as model outputs and shown in Table S1† are mostly consistent with previous studies.^{20,23} However, as the complex reaction mixture has been reflected in only 19 chemical reactions, the absolute numbers of the rate coefficients obtained for the individual reactions likely do not reflect reality, because further secondary chemistry of various kinds is not included explicitly. It should also be noted that different types of tyrosine residues have not been explicitly included within the model, although these can nitrate at different rates and have different surface accessibilities.^{35,44} Nevertheless, qualitatively, the model results suggest that protein nitration occurs at faster rates than protein oligomerization. The observed and modelled preference of nitration over oligomerization can be rationalized by comparing the mass fraction of nitrated monomer (calculated by multiplying the mass fraction of monomer with NTyr/monomer) with the total oligomer mass fraction. Nitrated monomers were observed to have two times or higher mass fractions compared to all other oligomer mass fractions combined under all experimental conditions. This result indicates that Tyr nitration outcompetes the dimerization/oligomerization process.^{22,45}

4. Implications and conclusions

In this study we investigated the kinetics and mechanism of the nitration and oligomerization of proteins induced by O₃ and NO₂ under different atmospherically relevant conditions. We showed the concentration and time dependence of the formation of dimers, trimers and higher protein oligomers as well as their individual nitration degrees. The temporal evolution of the concentrations of these multiple reaction products was well reproduced by a kinetic model with 19 chemical reactions. Protein nitration was found to be kinetically favored over protein oligomerization under the experimental conditions studied in this work. On the basis of the observation that nitrated oligomers formed upon exposure to O₃/NO₂, we suggest further investigation on the allergenic and immunogenic effects by nitrated protein oligomers. The nitrated oligomers were also found in the physiologically-relevant peroxyxynitrite-induced protein nitration and oligomerization,⁴⁶ for which the mechanism is similar to the one we reported here.^{25,47}

The implications of protein chemistry with O₃ and NO₂ under atmospherically relevant conditions are illustrated in Fig. 6. The overall nitration and oligomerization rates were both almost one order of magnitude higher in the aqueous phase than for 45% RH, indicating an increased relevance of the processes under cloud-processing conditions. Also, the yields of protein nitration and oligomerization can be strongly influenced by changes in relative humidity leading to changes of phase states. The protein dimers can yield up to 20% (by mass) for 12 h exposure in the liquid phase and the NDs of monomers and dimers can be up to 7% and 5%, respectively. This result indicates that on average, 1.4 Tyr residues in each monomeric protein molecule and 2 Tyr residues in each dimeric molecule are present in their nitrated forms. For proteins in solid or semi-solid phase states, our measurement and model results suggest that higher protein oligomers are likely to be found at lower RH, *e.g.*, on the surface of bioaerosols, whereas the



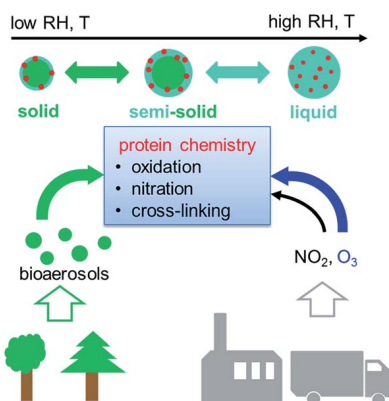


Fig. 6 Atmospheric protein chemistry by ozone (O_3) and nitrogen dioxide (NO_2). Reaction rates are limited by the phase state of proteins and hence the diffusivity of oxidants and protein molecules, which changes with relative humidity (RH) and temperature (T). Air pollutants such as NO_2 and O_3 can enhance allergen release from bioaerosols (e.g., pollen) with O_3 being more important in triggering the nitration, cross-linking and oxidation of allergenic proteins in bioaerosols and other protein-containing particles (e.g., Bet v 1 on urban road dust¹⁵).

NDs of monomers and dimers remain at ~ 1 –2%. Using *ab initio* calculations, Sandhiya *et al.*²² also showed that increased temperature can speed up the formation of tyrosyl radicals by ozonolysis. Thus, it is expected that nitrated and dimeric protein species could be important products of exposure to O_3 and NO_2 under tropical or summer smog conditions. These differences in reaction kinetics should be taken into account in studies on the physiological effects of proteins exposed to anthropogenic air pollutants, as the allergenic proteins in various oligomerization and nitration states might have a different immunogenic potential.

Both increasing levels of O_3 and NO_2 are able to damage pollen grains and facilitate the release of allergens in polluted environments.^{3,8,9} However, our results show that the tyrosine nitration and cross-linking of proteins are sensitive towards an increase in O_3 concentration and rather insensitive towards changes in ambient NO_2 concentrations. Therefore, effective control of the enhanced allergenicity induced by air pollutant-modified aeroallergens should point towards the decrease of ambient ozone concentrations.

Acknowledgements

F. L. and S. L. acknowledge financial support from the China Scholarship Council (CSC). C. J. K. and T. B. acknowledge support by the Max Planck Graduate Center with the Johannes Gutenberg University Mainz (MPGC). C. J. K. acknowledges financial support by the German Research Foundation (DFG; grant no. KA 4008/1-2).

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